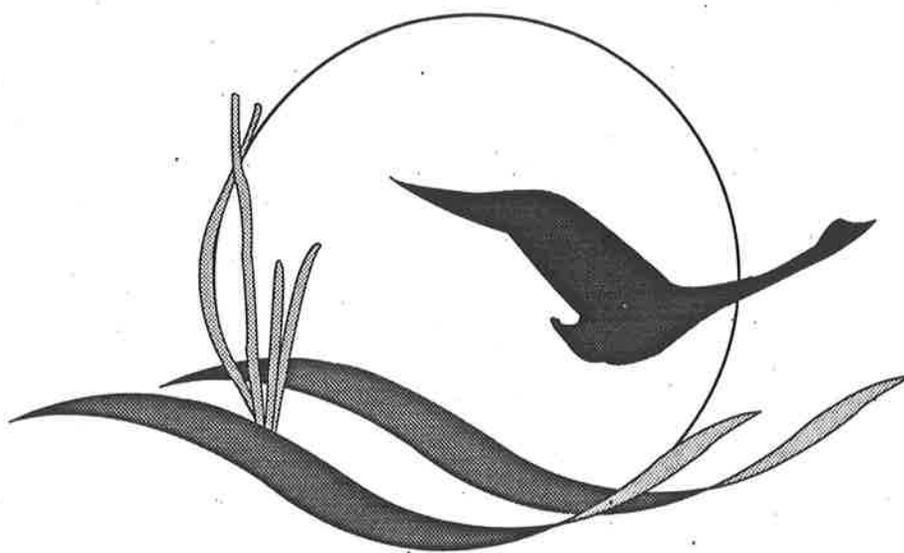


Chesapeake Bay Atmospheric Deposition Study Final Report



Chesapeake Bay Program



Chesapeake Bay Atmospheric Deposition Study Final Report

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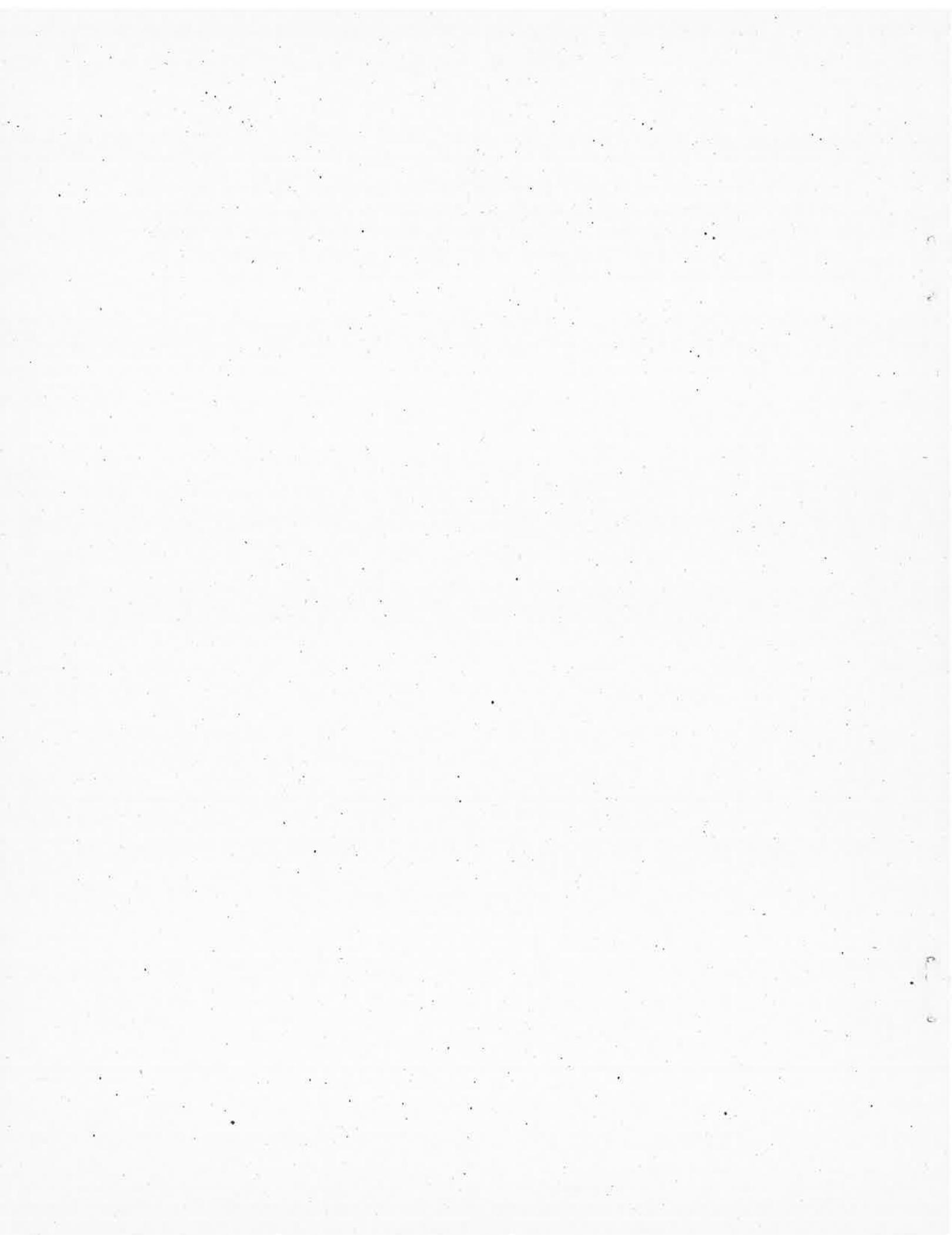


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Abstract

I. Introduction

I.1. Background. The Chesapeake Bay is the largest estuary in North America and one of the most productive aquatic ecosystems on the planet. Its 166,000 square kilometer watershed drains farmlands, forests, marshes, and urban areas in six states, and is home to 14.2 million people. Historically, the Chesapeake Bay has supported a large commercial fisheries. The wetlands surrounding the bay provide winter habitat for many migratory waterfowl. Dramatic decreases in natural habitat and water quality during the past four decades includes the virtual elimination of submerged aquatic vegetation and the substantial increase in seasonal anoxia in the Bay's main channel. This degradation has been attributed to increased erosion (with subsequent increases in turbidity) and enhanced external loadings of the nutrients phosphorus and, especially, nitrogen to the Bay's waters. Overharvesting of filter feeding oysters may have also altered the Bay's nutrient and energy cycles, amplifying eutrophication. In response to seasonal anoxia and the loss of submerged aquatic vegetation, stringent nutrient control strategies were adopted in 1987. To date, more than \$400 million has been spent to reduce nutrient loadings to the Chesapeake Bay. Chemical contaminants enter the Chesapeake Bay from a wide variety of point and diffuse sources. Historically, readily-identifiable industrial and municipal point source discharges in the cities on the major tributaries of the Bay (*e.g.*, Baltimore, MD, Washington, D.C., Richmond and Norfolk, VA) have been the focus of nutrient control strategies. As these sources were controlled, and as our understanding of diffuse sources increased, contaminant inputs from run-off from the watershed became more important. More recent water quality strategies in the Chesapeake Bay region are based on controlling activities within the watershed and tributaries (1991 Amendments to the Chesapeake Bay Agreement).

In 1988, Fisher *et al.* suggested that the atmosphere is a significant source of nitrogen to the Chesapeake Bay ecosystem. Based on nitrogen loads by wet and dry deposition to the watershed, these authors estimated the fraction of deposited nitrogen which was transported through the watershed to the surface waters of the Chesapeake Bay under differing land uses. Their calculations, which suggested that atmospheric deposition accounts for 25% of the external nitrogen entering the Bay, were thoroughly examined at an EPA-sponsored workshop in 1990. Despite considerable uncertainty in the 'watershed transmission' term of their model, the consensus of the workshop was that the atmosphere supplies anywhere between one-quarter to one-third of nitrogen to the Chesapeake Bay (Hinga *et al.*, 1991). Jaworski and Linker (1990), using several approaches to model watershed export of nitrogen, conclude that atmospheric deposition contributes between 35 and 39% of the total nitrogen loading to the Potomac River watershed. Further, Fisher *et al.* (1988) argued that the atmosphere will become a relatively more important source of nitrogen during the next decades as point sources are controlled and emissions of nitrogen oxides into the Chesapeake Bay airshed increase.

Studies conducted in the southern Chesapeake Bay in the early 1980's suggested that the atmosphere is also a significant source of organic contaminants. Velinsky *et al.* (1986)

determined that organic carbon loadings in bulk atmospheric deposition were greater than the estimated tributary loadings of organic carbon to the bay. Anthropogenic hydrocarbons, including polycyclic aromatic hydrocarbons, were detected in bulk deposition samples, with calculated loadings estimated to be comparable to hydrocarbon loadings from the municipal wastewater discharges to the bay (Webber, 1993). While these earlier studies demonstrated the potential importance of the atmosphere in supplying contaminants to the Chesapeake Bay, they were limited by their methodologies (*i.e.*, bulk deposition sampling) and their relatively limited temporal and spatial scope. To further explore this issue, and to quantify the loadings of atmospheric contaminants to the Chesapeake Bay, we operated a network of three stations along the shoreline of the bay during 1990-92. In this chapter, the results of the Chesapeake Bay Atmospheric Deposition Study (CBADS) are presented and discussed relative to the issues raised by the U.S. Environmental Protection Agency's Great Waters Program.

I.2. Objectives. The primary objective of the CBADS network was to provide the best possible estimates of *total, annual* atmospheric loadings of a variety of trace elements and organic contaminants directly to the surface waters of the Chesapeake Bay. Accurate estimates of bay-wide annual loadings required that we characterize the spatial and temporal variability in contaminant inventories in the atmosphere and in depositional fluxes to the bay. Based on previous studies of wet deposition in the region (*e.g.*, Tyler, 1988) and given the resources available for the network, we selected three non-urban shoreline locations for sampling sites. In establishing this initial network, we purposefully avoided the influence of urban areas by locating our sites more than 50 kilometers from metropolitan areas. By minimizing possible urban influences, we insured that the resulting loading estimates are minimum values. Although it is most likely that some fraction of the materials deposited from the atmosphere to the considerable watershed of the Chesapeake Bay are ultimately transported to the surface waters, we did not attempt to characterize loadings to the watershed in this study. Because deposition to the various land surfaces are likely much different than that to the water surface, fluxes measured at our shore-based stations can not be extrapolated with confidence to the watershed. Additionally, the large uncertainty in our understanding of the fate of materials deposited to the land surface (*i.e.*, the fraction transmitted to the receiving water), preclude the simple estimation of the indirect atmospheric loading of contaminants to the Chesapeake Bay.

I.3. Project Description. The sampling strategy employed in CBADS is a compromise between the need to monitor all deposition events and our desire to understand sources of air masses and the processes of deposition. In order to capture all wet deposition during the study period at these relatively remote sites, automated, wet-only precipitation samplers were used to collect temporally integrated samples. Initially, weekly integrated aerosol samples were collected for trace elemental analysis in order to 'smooth' the inherently variable atmospheric signal. However, because these weekly-integrated samples were collected under a wide variety of atmospheric conditions, they were of little use either in estimating air mass sources or in calculating dry aerosol depositional fluxes. Therefore, after the first six months of CBADS, we altered the aerosol sampling schedule to 12 hours every seven days. Similarly, while the integrated precipitation sampling provided the best measure to total wet depositional fluxes, any

information about the mechanism of contaminant incorporation into the precipitation was lost. To investigate the precipitation scavenging in detail, a series of rain storm were intensively sampled during the summer of 1993 (Poster and Baker, 1995; Poster and Baker, 1996 a;1996b).

Sampling was initiated at the mid-bay site in Summer, 1990, with the northern and southern stations beginning operation later in 1990 and in early 1991, respectively (Table II.1, Figure II.1). The three sites were operated continuously until the end of September, 1993. In this chapter, measurements made at the three sites during this period are used to calculate atmospheric depositional fluxes and loading of contaminants to the Chesapeake Bay region, to explore spatial and temporal variability in contaminant inventories and fluxes, and to compare contaminant levels and fluxes in the Chesapeake Bay region to those measured elsewhere.

II. Methods

II.1. Sampling. Sampling and analytical methods employed in the Chesapeake Bay Atmospheric Deposition Study are described in detail in a series of papers published elsewhere (Leister and Baker, 1994; Wu *et al.*, 1994; Scudlark *et al.*, 1994), and will be summarized here. Each of the three sampling sites were chosen to be as close as possible to the Chesapeake Bay in secured, open grassy fields away from local sources. The northern site is located at the University of Maryland's Wye Agricultural Research and Experimental Station on the shore of the Wye River in Queenstown, MD (38°53'N, 76°08'W). The northern site is located in a grassy field approximately 50 m from the road servicing the Wye Station and is a NADAP site. Land use in the area is largely agricultural. The mid-Bay site is located in a 75 x 75 m grassy clearing in a wooded area on the western shore of the Bay between the mouths of the Patuxent and Potomac Rivers (38°13'N, 76°23'W). This site is owned by the State of Maryland and is used for environmental educational activities. A dirt road which is very infrequently used is approximately 30 m from the mid-Bay site. The land use in the area is primarily agricultural and low density housing, and the Patuxent River Naval Air Station is 4.5 km to the north of this site. The southern site is located at Haven Beach in Mathews County, Virginia (37°26'N, 76°15'W). This site, which approximately 100 km due east of Richmond, VA, consists of a wooden platform in a salt marsh directly adjacent to the western shoreline of the bay.

Air samples were collected for elemental (aerosol particles) and organic contaminants (gaseous and aerosol particle-associated) as follows. Initially, 7 day (168 hour) air samples for elemental analysis were collected by drawing air at a flow rate of 10 liters/minute through a dichotomous sampler equipped with a single stage impactor to remove particles $>10\mu\text{m}$. Aerosol particles were collected on 2 μm -pore, 47-mm Gelman "Teflo" filters. A dry gas meter placed after the impactor was used to determine the sample volume and average flow rate for each sample. Beginning in 1992, the sampling protocol for elemental analysis was changed to 12 hours once every seven days. A high-volume air sampler was deployed to collect air samples for gaseous and particle-associated organic contaminants (Leister and Baker, 1994). Air is drawn through a 20.3 cm x 25.4 cm glass fiber filter to collect particles and through a polyurethane foam

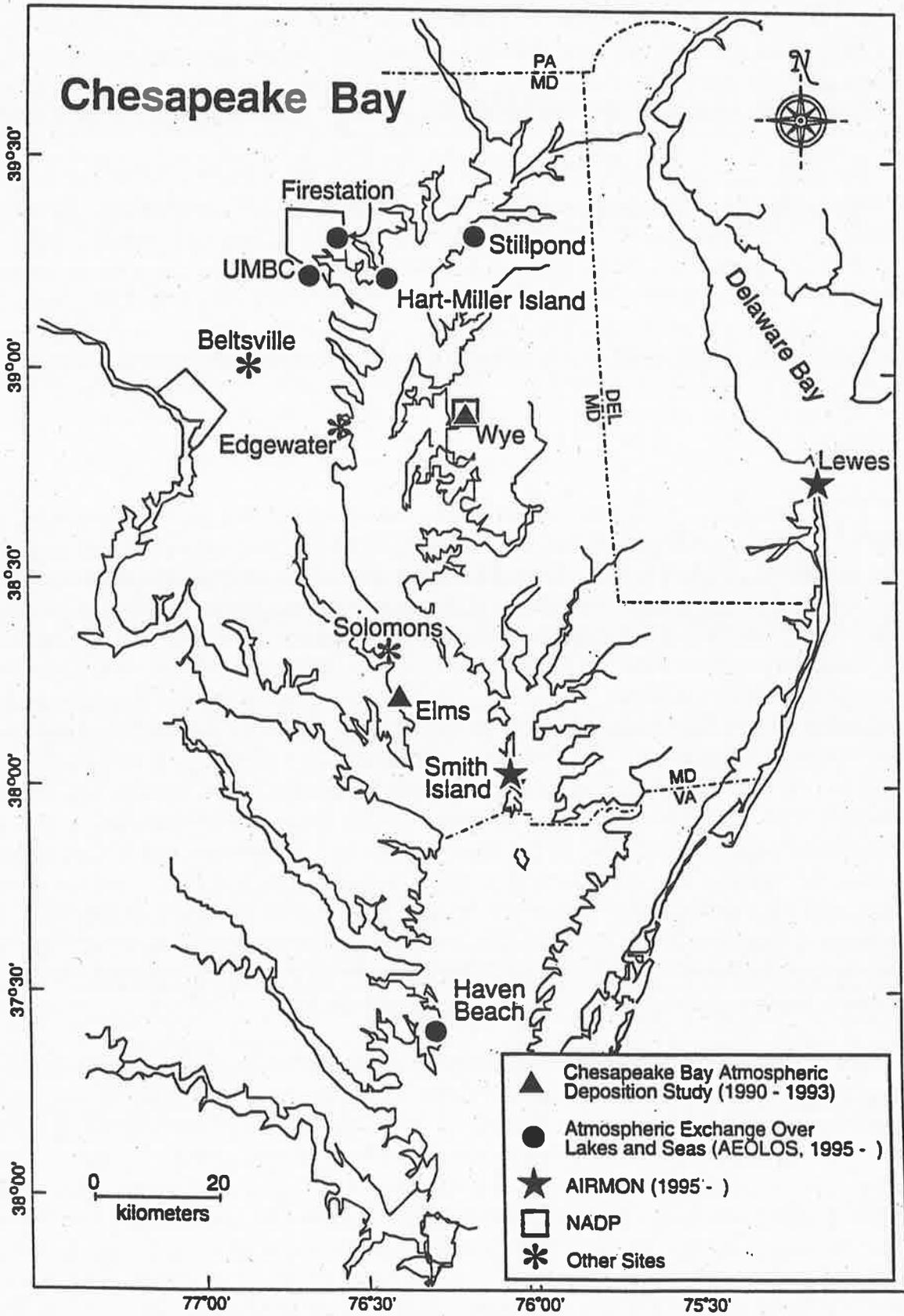


Figure II.1. Chesapeake Bay atmospheric deposition sampling sites.

Table II.1. Sampling schedule, Chesapeake Bay Atmospheric Deposition Study

	Date		Sampling	
	Start	End	Frequency	Interval
Wye				
Trace Elements-Precipitation	5 July 1990	28 September 1993	7 Days	7 Days
Trace Elements-Aerosols	5 June 1990	28 September 1993	6 Days	0.5 ^a Days
Organics-Precipitation	14 July 1992	28 September 1993	14 Days	14 Days
Organics-Air	14 July 1992	28 September 1993	14 Days	0.5-1 Days
Elms				
Trace Elements-Precipitation	5 June 1990	28 September 1993	7 Days	7 Days
Trace Elements-Aerosols	5 June 1990	28 September 1993	7 Days	0.5 ^a Days
Organics-Precipitation	26 June 1990	28 September 1993	14 Days	14 Days
Organics-Air	12 June 1990	28 September 1993	14 Days	0.5-1 Days
Haven Beach				
Trace Elements-Precipitation	4 December 1990	28 September 1993	7 Days	7 Days
Trace Elements-Aerosols	23 November 1990	28 September 1993	7 Days	0.5 ^a Days
Organics-Precipitation	18 December 1990	28 September 1993	14 Days	14 Days
Organics-Air	18 December 1990	28 September 1993	14 Days	0.5-1 Days

^a7 day interval in 1990-91.

plug to collect gaseous organic chemicals. Air sampling times ranged from 12 to 24 hours, depending upon season. Flow rates through the organics air sampler were estimated for each sample from the measured pressure drop across the sampling train. Ambient air samples were collected bi-weekly (*i.e.*, every other Tuesday) to correspond with the precipitation sampling schedule.

Weekly integrated precipitation samples were collected for elemental analysis using an automated wet-only collector which was specially modified for trace element sampling (Aerochem Metrics, Inc., Bushnell, FL). The rigorous trace element "clean" sampling, handling, and analytical procedures described by Tramontano *et al.* (1987) and Scudlark *et al.* (1992) were used. Precipitation collected in an acid-washed, clear high density polyethylene bucket was acidified with ultra-pure HCl to 0.4% v/v, leached for 24 hours, transferred to acid-washed low density polyethylene storage bottles and frozen (Scudlark *et al.*, 1994). Field blanks, consisting of acidified water rinses of buckets deployed during weeks when no precipitation fell were analyzed to assess the magnitude of contamination during sample processing. Organic contaminants in precipitation were collected with an automated sampler consisting of a 1 m² stainless steel funnel, and in-line glass fiber filter and resin cartridge to isolate particulate-associated and dissolved organic contaminants from the precipitation. A peristaltic pump downstream of the sampling train pulled rainwater through the filter and resin cartridge during each storm, minimizing potential analyte losses or repartitioning during sample storage (Leister and Baker, 1994). Precipitation samples for organic contaminant analyses were composited for fourteen days. Sampling and analytical details are described in Leister and Baker (1994). Concentrations of major ions in weekly precipitation samples were determined by separate sampling at each site, either as part of the NAPAP (Wye), the Maryland STAR network (Elms), or as part of this study (Haven). Precipitation intensities and amounts were measured using Belfort gauges or, when not available, estimated from the volume of precipitation collected for chemical analyses (Table II.2).

Sampling at the three Chesapeake Bay Atmospheric Deposition stations began during the second half of 1990 and continued through September 1993. During the Summer, 1993, an intensive sampling program, consisting of event-based precipitation sampling and daily air sampling for 28 days, was conducted in order to study sources of atmospheric contaminants and to better understand scavenging by precipitation. Some of the results of this intensive program are described by Poster and Baker (1995; 1996 a,b).

II.2. Analytical Techniques. Trace elements on aerosol particles were measured using the methods described in Wu *et al.* (1994). Each filter was divided into two equal parts, and one half was analyzed for up to forty elements, including aluminum, arsenic, chromium, iron, manganese, sulfur, selenium, and zinc by instrumental neutron activation analysis. Particles on the remaining half of each filter were digested and analyzed for copper, nickel and lead by inductively coupled plasma-atomic emission spectrometry and for cadmium (and sometimes nickel) by graphite furnace atomic adsorption spectrometry with Zeeman background correction. Precipitation samples were analyzed for elements using the methods of Scudlark *et al.* (1994). Aluminum, cadmium, copper, chromium, iron, manganese, nickel, lead and zinc were quantified

Table II.2. Monthly precipitation amounts (cm/month) at the three Chesapeake Bay Atmospheric Deposition Study sites, 1990-1993.

Month	Wye	Elms	Haven Beach	Average	
1990	June	NA	13.08	NA	13.08
	July	9.48	12.40	NA	10.94
	August	10.90	14.94	NA	12.92
	September	4.11	3.63	NA	3.87
	October	8.59	10.54	NA	9.57
	November	5.66	4.39	0.05	3.37
	December	7.80	9.32	9.51	13.32
1991	January	9.73	13.06	10.94	11.24
	February	4.98	5.49	3.83	4.76
	March	10.59	11.00	8.23	10.11
	April	11.53	8.36	16.92	12.27
	May	3.15	6.43	2.55	4.04
	June	3.40	6.91	23.47	11.26
	July	12.98	15.93	4.72	11.21
	August	9.50	13.11	17.22	13.27
	September	11.59	9.09	6.24	8.97
	October	3.66	7.39	10.08	7.04
	November	1.60	1.47	0.77	1.28
	December	11.93	10.59	7.97	10.16
1992	January	8.2	3.23	3.48	4.96
	February	6.3	5.58	6.36	6.08
	March	12.9	14.94	8.08	11.97
	April	3.3	3.2	6.58	6.53
	May	6.0	10.2	8.52	8.23
	June	4.6	9.525	9.37	7.83
	July	10.1	4.32	16.38	10.26
	August	15.1	24.15	29.87	23.04
	September	11.0	11.15	8.39	10.17
	October	5.4	6.93	10.50	7.61
	November	8.1	3.10	3.43	4.87
	December	8.8	10.92	11.35	10.36
1993	January	5.89	6.93	11.66	8.15
	February	6.27	6.84	4.97	6.02
	March	20.39	10.39	18.88	16.55
	April	7.57	5.17	7.94	6.89
	May	8.28	22.71	19.82	16.94
	June	4.77	4.22	5.44	4.81
	July	5.96	2.80	4.54	4.43
	August	10.25	7.38	6.09	7.90
	September	6.14	6.55	5.03	5.90
Totals	1991	94.6	108.8	112.9	105.5
	1992	99.8	107.3	122.3	109.8
	Overall	316.5	357.3	337.6	337.1

by graphite furnace atomic adsorption with a L'Vov platform to minimize matrix suppression. Citric acid was used as a matrix modifier for aluminum and iron measurements. Arsenic and selenium concentrations in precipitation were measured by hydride generation, cryogenic preconcentration, selective volatilization and detection *via* photoionization (arsenic) or atomic absorption spectrophotometry (selenium), using the methods of Cutter (1986) and Cutter *et al.* (1991). Elemental detection limits and other quality assurance information for these analyses are described in Wu *et al.* (1994) and Scudlark *et al.* (1994).

Organic contaminants in air and precipitation were identified and quantified using high resolution capillary gas chromatography with electron capture (organochlorines) and mass spectrometric (polycyclic aromatic hydrocarbons) detection, as described in detail in Leister and Baker (1994). Seventy-four chromatographic peaks containing one or more polychlorinated biphenyl (PCB) congener were quantified by comparison with a mixed Aroclor standard (Mullin, 1985; Mullin *et al.*, 1984), with operational method detection limits set equal to three times the level measured in the corresponding matrix blank (Leister and Baker, 1994). Fourteen unsubstituted polycyclic aromatic hydrocarbons (PAHs) were identified and quantified by isotope-dilution selected ion mass spectrometry, using perdeuterated PAHs as internal standards. As PAHs were generally present in matrix blanks at undetectable levels, operational detection limits were determined by the instrumental sensitivity. Quality assurance details for the organic contaminant analyses, including potential sampling artifacts and the procedures used to insure sufficiently low matrix blanks, are described in Leister and Baker (1994) and Leister (1994).

III. Results

III.1 Concentrations in Air. The elemental composition of aerosol particles less than 10 μm diameter was dominated by the crustal elements aluminum and iron, and by sulfur (in the form of sulfate). Concentrations of these three elements averaged over the three sampling sites during 1991 and 1992 were 119, 111, and 2123 ng/m^3 , respectively (Table III.1). Trace element concentrations averaged over the same period ranged from 0.16 ng/m^3 for cadmium to 12.6 ng/m^3 for zinc (Figure III.1, Table III.1). The fraction of each element derived from non-crustal (*e.g.*, combustion) sources was estimated based on the average concentration of elements in the Earth's crust (Turekian and Wedepohl, 1961), and assuming all of the measured aluminum associated with aerosol particles is derived from erosion of soils. In the Chesapeake Bay region, non-crustal sources supply greater than 95% of most of the elements measured on aerosol particles. Exceptions include iron, manganese, and chromium (Wu *et al.*, 1994). Arsenic, cadmium, lead, sulfur and selenium are almost exclusively non-crustal, and are likely introduced into the atmosphere as a result of combustion of fossil fuels and incineration of municipal wastes.

Not unexpectedly, concentration of elements on aerosol particles were highly variable throughout the more than three years of continuous sampling. Variations of more than one order of magnitude likely result not only from changes in the total suspended particulate concentrations in the air but also from temporal changes in regional emissions. For example, the concentration

Table III.1. Elemental composition of aerosol particles (ng/m³) collected at the three Chesapeake Bay Atmospheric Deposition Study sites, 1991 and 1992.

	Al	As	Br	Cd	Cr	Cu	Fe	Mn	Ni	Pb	S	Sc	V	Zn
1991														
Wye	109	0.63	2.96	0.13	0.72	1.8	108	2.66	2.86	3.4	2548	1.8	3.5	14.4
Elms	147	0.56	3.31	0.12	0.60	1.5	149	2.74	2.41	2.7	2953	1.8	4.7	12.5
Haven Beach	133	0.47	3.55	0.11	0.46	2.0	120	2.87	2.81	2.3	2140	1.5	4.8	11.5
Bay Average	130	0.55	3.27	0.12	0.59	1.77	126	2.76	2.69	2.8	2547	1.7	4.33	12.8
1992														
Wye	108	0.66	2.65	0.22	0.88	1.73	113	2.90	4.42	5.8	1547	1.8	4.5	13.2
Elms	123	0.57	2.53	0.19	0.78	2.11	107	2.40	3.34	4.9	1779	1.6	5.0	13.1
Haven Beach	75	0.59	2.83	0.18	0.46	1.89	66	1.81	3.25	4.2	1772	1.4	4.75	10.7
Bay Average	102	0.61	2.67	0.20	0.71	1.91	95	2.37	3.67	5.0	1599	1.6	4.75	12.3
Overall														
Wye	108	0.65	2.81	0.18	0.80	1.77	110	2.78	3.64	4.6	2048	1.8	4.0	13.8
Elms	135	0.57	2.92	0.16	0.69	1.81	128	2.57	2.88	3.8	2366	1.7	4.9	12.8
Haven Beach	104	0.53	3.19	0.15	0.46	1.95	93	2.34	3.03	3.25	1956	1.5	4.8	11.1
Bay Average	116	0.58	2.97	0.16	0.65	1.84	110.3	2.56	3.18	3.88	2123	1.67	4.57	12.6

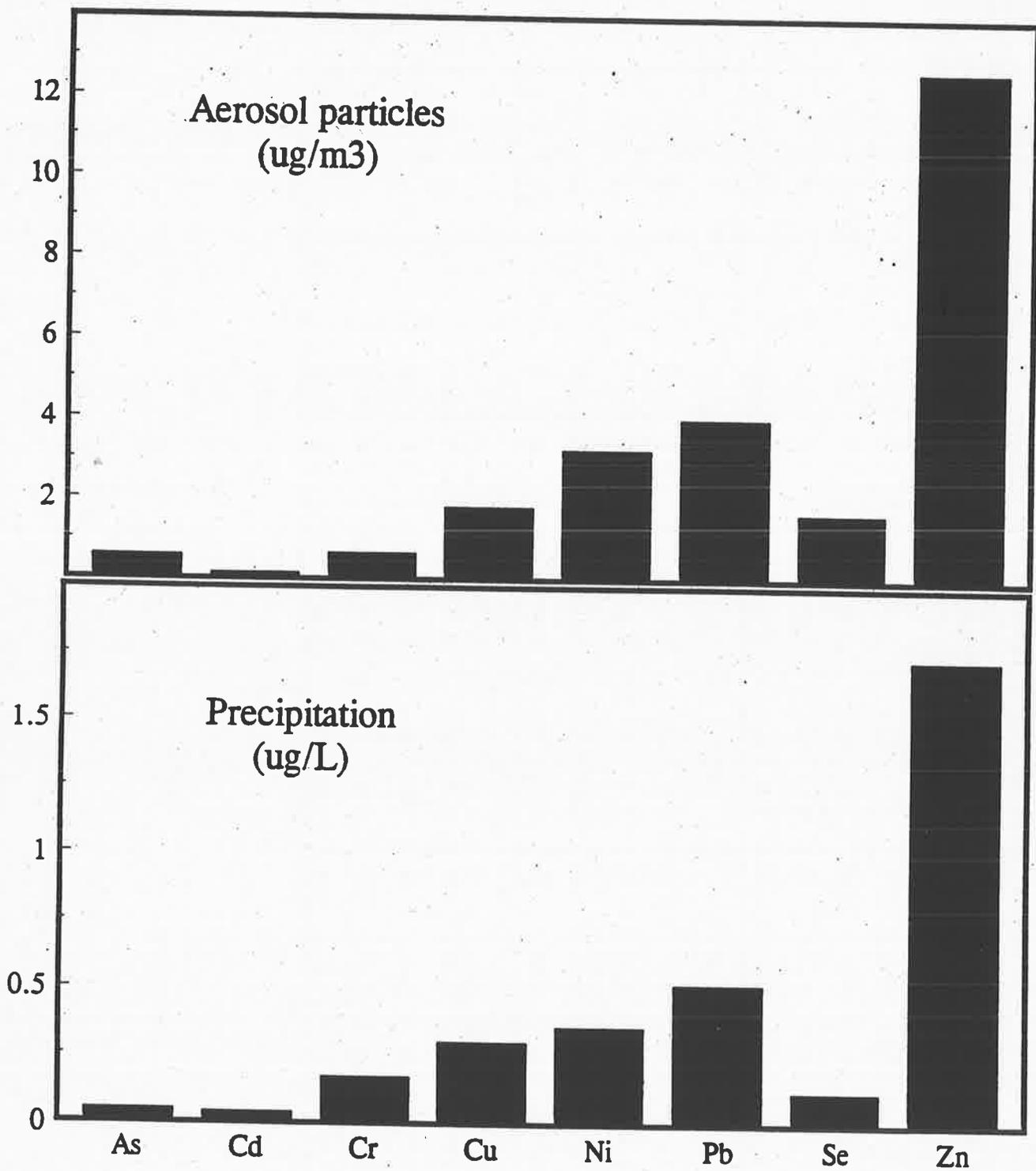


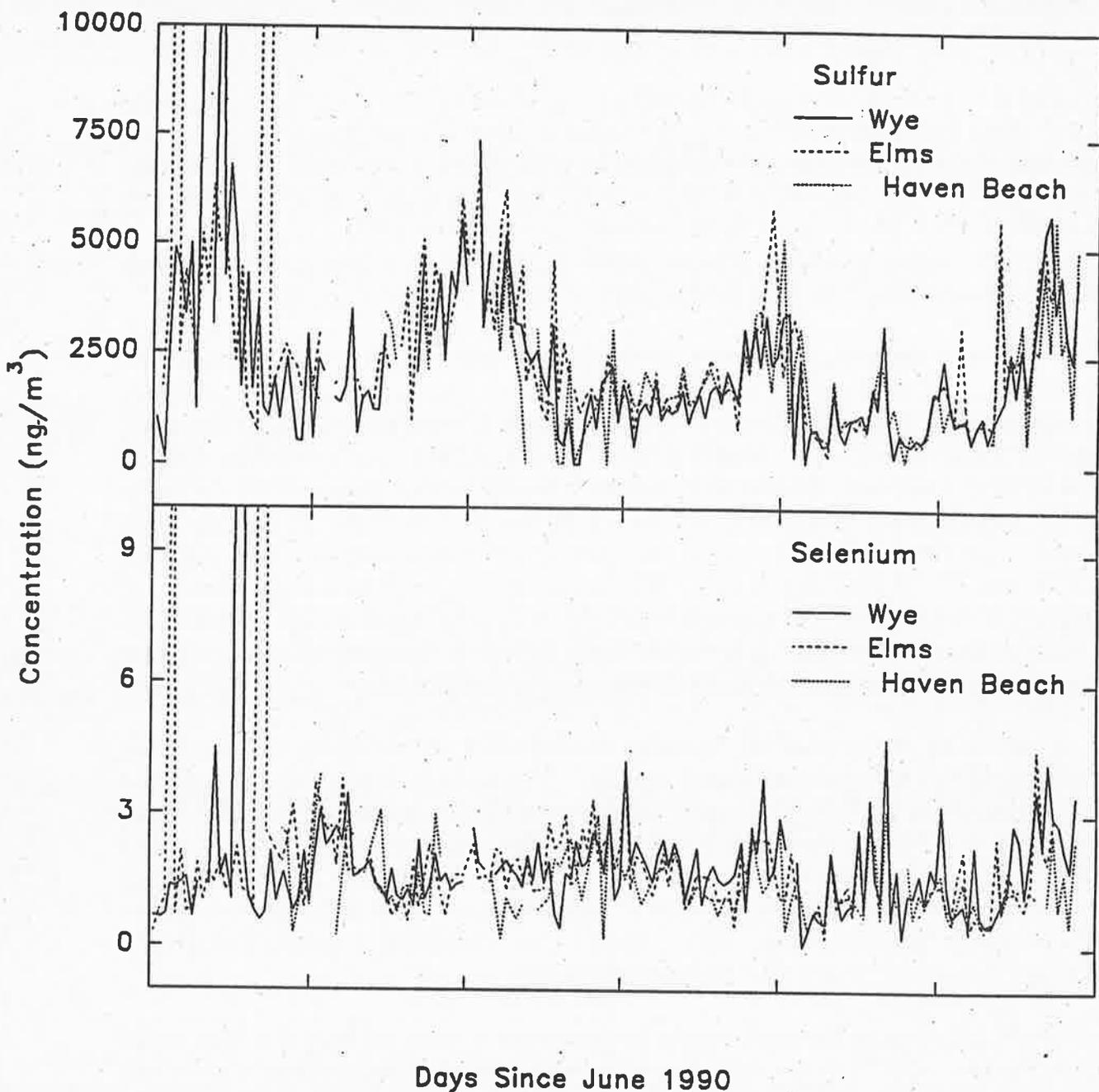
Figure III.1. Average elemental composition of aerosol particles and precipitation, Chesapeake Bay Atmospheric Deposition Study, Wye, 1991.

of selenium, a tracer of coal combustion, averaged 1.7 ng/m^3 during 1991 and 1992, and generally increased during winter months (Figure III.2). Spikes in particulate selenium in the atmosphere suggests that the sampling sites were occasionally impacted by power plants in the region. Concentrations of other combustion-related elements were also slightly higher in the winter, although clear trends are obscured by the inherent short term variability. Concentrations of particulate sulfur are elevated in the summer (Figure III.2), when more photochemical oxidants and sunlight are available to convert gaseous sulfur to sulfate (Gordon *et al.*, 1994).

To compare interannual variations in elemental concentrations, annual averages were calculated for each of the three sites for 1991 and 1992. Several measurements of extraordinarily high concentrations, likely caused by localized contamination, were not included in the annual averages. Concentrations of the crustal elements aluminum, iron, and manganese declined between 10 to 50% between 1991 and 1992, perhaps reflecting less soil erosion during the wetter 1992. The average annual sulfate concentration in 1992 was 60-80% of the 1991 average at the three sites. Conversely, concentrations of atmospheric cadmium, nickel, lead were higher by 60%, 35%, and 70% in 1992 than in 1991. Whether this reflects differences in meteorology or real changes in emissions of these elements to the regional atmosphere is unclear. Even though these data represent a large effort, it is obviously still difficult to argue for systematic trends in the concentrations of trace elements in the Chesapeake Bay atmosphere.

As shown in Figure III.2, the measured concentrations of trace elements are generally within a factor of two among the three sampling sites. On an average annual basis, concentrations slightly decrease from north (Wye) to south (Haven Beach), except sulfur, which is 15% higher at Elms (1991 and 1992) and Haven Beach (1992) than at the northernmost Wye site. The general decreasing trends observed from north to south, in light of increasing sulfate, may indicate higher levels of sulfur-depleted combustion sources (*e.g.*, incinerators, vehicles) in the northern reaches of the Chesapeake Bay. In general, spatial trends in the atmospheric concentrations of trace elements are substantially lower than corresponding temporal trends.

Semivolatile organic chemicals exist in the atmosphere as gases and associated with aerosol particles (Pankow, 1987). In this study, bay-wide annual average concentrations of polycyclic aromatic hydrocarbons ranged from 16 ng/m^3 for dibenz[*a,h*]anthracene to 2590 ng/m^3 for phenanthrene (Tables III.2-4). Atmospheric concentrations were quite variable between the 12-24 hour samples collected semi-weekly, with individual measurements ranging from one tenth to ten times the annual average concentrations. These variations likely results from sampling air masses coming from differing directions, from changes in local and regional emissions, and from differences in atmospheric degradation and deposition rates. For example, increased concentrations of gas-phase PAHs such as pyrene during the summer months may reflect both higher temperatures (*i.e.*, enhanced volatilization) and increased coal and oil combustion to meet electrical demand for air conditioning (Figure III.3). Increases in particulate PAHs, such as benzo[*a*]pyrene, in the atmosphere may result from local burning of yard wastes and of wood for home heating. Some variation in atmospheric levels of organic chemicals may result from the efficient removal of particulate PAHs by precipitation (Poster and Baker, 1996a and b). In



Days Since June 1990

Figure III.2 Concentrations of particulate sulfur and selenium at three Chesapeake Bay sites, 1990-1992.

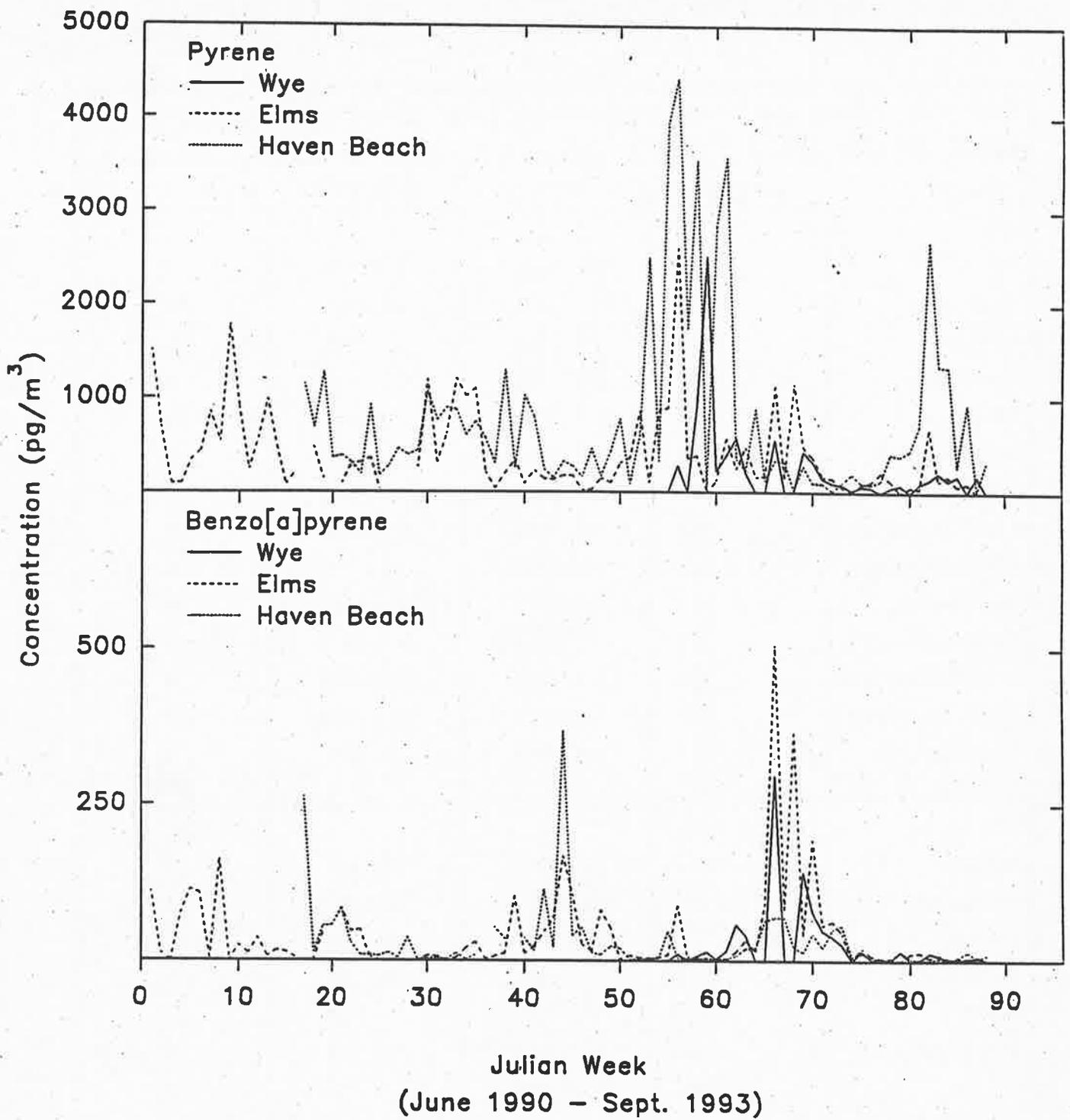


Figure III.3 Pyrene and Benzo[a]pyrene concentrations measured in air at the Wye, Elms, and Haven Beach sites.

Table III.2. Average annual concentrations of total (vapor plus particulate) organic contaminants in air (pg/m³), Wye.

Compound	Mean Conc.	Range	Range % Part.	N
1992				
Fluorene	869.1	194.8 - 2645.4	0.0 - 1.7	26
Phenanthrene	3107.7	967.5 - 5361.9	0.4 - 14.0	26
Anthracene	36.8	9.0 - 272.1	6.6 - 98.6	26
Fluoranthene	646.1	188.3 - 1719.8	1.8 - 74.0	26
Pyrene	706.3	202.6 - 2512.7	1.2 - 71.5	26
Benz[a]anthracene	34.9	3.5 - 177.3	33.5 - 99.3	26
Chrysene	95.6	26.0 - 417.4	33.2 - 96.5	26
Benzo[b]fluoranthene	105.6	9.0 - 489.0	82.1 - 99.1	26
Benzo[k]fluoranthene	76.1	6.8 - 379.3	82.8 - 99	26
Benzo[e]pyrene	77.5	8.1 - 403.0	84.2 - 99.6	26
Benzo[a]pyrene	56.1	3.1 - 297.0	94.2 - 99.8	26
Indeno[123cd]pyrene	72.1	6.4 - 405.6	82.5 - 99.8	26
Dibenz[ah]anthracene	23.0	2.1 - 90.0	57.1 - 99.6	26
Benzo[ghi]perylene	85.1	1.6 - 442.0	89.7 - 100	26
Total PCB	n/a	n/a		
1993				
Fluorene	523.1	0.0 - 3090.1	0.0 - 100	21
Phenanthrene	1153.4	0.0 - 6239.2	0.0 - 100	21
Anthracene	19.8	0.0 - 239.8	0.0 - 100	21
Fluoranthene	152.6	0.0 - 591.4	0.0 - 100	21
Pyrene	114.5	0.0 - 432	0.0 - 100	21
Benz[a]anthracene	19.3	0.0 - 161.7	0.0 - 100	21
Chrysene	62.2	0.0 - 432.6	0.0 - 100	21
Benzo[b]fluoranthene	47.9	0.0 - 332.8	0.0 - 100	21
Benzo[k]fluoranthene	31.1	0.0 - 138.3	0.0 - 100	21
Benzo[e]pyrene	44.7	0.0 - 342.4	0.0 - 100	21
Benzo[a]pyrene	20.3	0.0 - 143.1	0.0 - 100	21
Indeno[123cd]pyrene	39.9	0.0 - 340.3	0.0 - 100	21
Dibenz[ah]anthracene	4.1	0.0 - 34.5	0.0 - 100	21
Benzo[ghi]perylene	41.1	0.0 - 377.3	0.0 - 100	21
Total PCB	n/a			

Table III.2. Average annual concentrations of total (vapor plus particulate) organic contaminants in air (pg/m³), Wye.

Compound	Mean Conc.	Range	Range % Part.	N
Overall 1992-1993				
Fluorene	720.8	0.0 - 3090.1	0.0 - 100	47
Phenanthrene	2270.1	0.0 - 6239.2	0.0 - 100	47
Anthracene	29.5	0.0 - 272.1	0.0 - 100	47
Fluoranthene	434.6	0.0 - 1719.8	0.0 - 100	47
Pyrene	452.7	0.0 - 2512.7	0.0 - 100	47
Benz[a]anthracene	28.2	0.0 - 177.3	0.0 - 100	47
Chrysene	81.3	0.0 - 432.6	0.0 - 100	47
Benzo[b]fluoranthene	80.9	0.0 - 489	0.0 - 100	47
Benzo[k]fluoranthene	56.8	0.0 - 379.3	0.0 - 100	47
Benzo[e]pyrene	63.4	0.0 - 403	0.0 - 100	47
Benzo[a]pyrene	40.8	0.0 - 297	0.0 - 100	47
Indeno[123cd]pyrene	58.3	0.0 - 405.6	0.0 - 100	47
Dibenz[ah]anthracene	14.9	0.0 - 90	0.0 - 100	47
Benzo[ghi]perylene	66.3	0.0 - 442	0.0 - 100	47
Total PCB				

units are in pg/m³

Table III.3. Average annual concentrations of total (vapor plus particulate) organic contaminants in air (pg/m³), Elms.

Compound	Mean Conc	Range	Range % Part	N
1990-1991				
Fluorene	573.2	56 - 2724.5	0.0 - 23	35
Phenanthrene	1779.3	357 - 5692.2	0.3 - 23	35
Anthracene	50.4	0.9 - 181.3	0.0 - 100	35
Fluoranthene	404.7	46.9 - 2274.9	1.0 - 100	35
Pyrene	482	34.4 - 1502.6	0.8 - 100	35
Benz[a]anthracene	40	1.9 - 152.8	0.0 - 100	35
Chrysene	96.9	13.9 - 537.5	0.0 - 100	35
Benzo[b]fluoranthene	100.5	2.4 - 463.8	0.0 - 100	35
Benzo[k]fluoranthene	58.3	3.1 - 214.1	12.2 - 100	35
Benzo[e]pyrene	64.8	3.2 - 325.5	0.0 - 100	35
Benzo[a]pyrene	33.9	3.3 - 162.1	0.0 - 100	35
Indeno[123cd]pyrene	58	6.8 - 304.5	7.3 - 100	35
Dibenz[ah]anthracene	33.9	7.3 - 58.9	6.0 - 100	35
Benzo[ghi]perylene	63.8	5.6 - 242.7	12.4 - 100	35
Total PCB	213.9	17.4 - 507.8	n/a	35
1992				
Fluorene	885.9	0 - 7853	0.0 - 100	25
Phenanthrene	2012.8	34 - 11959	0.3 - 100	25
Anthracene	53	340 - 611.7	1.0 - 100	25
Fluoranthene	309	160 - 1512.3	3.1 - 100	25
Pyrene	468.0	160 - 2599.6	0.7 - 100	25
Benz[a]anthracene	49.0	317.7 - 320	38 - 100	25
Chrysene	162.0	440 - 941.0	2.1 - 100	25
Benzo[b]fluoranthene	178.2	0 - 920.5	2.4 - 100	25
Benzo[k]fluoranthene	130.7	590 - 674.7	1.8 - 100	25
Benzo[e]pyrene	121.6	135 - 770.3	69.5 - 100	25
Benzo[a]pyrene	72.7	85 - 507.1	9.1 - 100	25
Indeno[123cd]pyrene	132.6	79 - 911	63.9 - 100	25
Dibenz[ah]anthracene	19.0	9 - 118	0.0 - 100	25
Benzo[ghi]perylene	142.4	152 - 1047.6	84.7 - 100	25
Total PCB	n/a		n/a	

Table III.3. Average annual concentrations of total (vapor plus particulate) organic contaminants in air (pg/m³), Elms.

Compound	Mean Conc.	Range	Range % Parr	N
1993				
Fluorene	675.3	0.0 - 3087.4	0.0 - 4.3	19
Phenanthrene	1702.5	0.0 - 5505.4	0.0 - 33.7	19
Anthracene	33.6	0.0 - 296.8	0.0 - 3.9	19
Fluoranthene	225.6	0.0 - 994.2	0.0 - 100	19
Pyrene	167.9	0.0 - 679.0	0.0 - 100	19
Benz[a]anthracene	17.4	0.0 - 53.6	0.0 - 100	19
Chrysene	61.3	0.0 - 324.8	0.0 - 100	19
Benzo[b]fluoranthene	58.2	0.0 - 378.4	0.0 - 98.7	19
Benzo[k]fluoranthene	48.3	0.0 - 311.0	0.0 - 100	19
Benzo[e]pyrene	51.4	0.0 - 297.2	0.0 - 100	19
Benzo[a]pyrene	27.2	0.0 - 194.6	0.0 - 100	19
Indeno[123cd]pyrene	56.2	0.3 - 133.6	95.6 - 100	19
Dibenz[ah]anthracene	5.2	0.0 - 45.5	0.0 - 100	19
Benzo[ghi]perylene	52.2	1.3 - 347.2	0.0 - 100	19
Total PCB	n/a	n/a	n/a	
Overall 1990-1993				
Fluorene	696.1	0.0 - 7853	0.0 - 100	79
Phenanthrene	1834.8	0.0 - 11959	0.0 - 100	79
Anthracene	47.2	0.0 - 611.7	0.0 - 100	79
Fluoranthene	332.1	0.0 - 2274.9	0.0 - 100	79
Pyrene	403.2	0.0 - 2599.6	0.0 - 100	79
Benz[a]anthracene	37.5	0.0 - 320	0.0 - 100	79
Chrysene	109	0.0 - 941	0.0 - 100	79
Benzo[b]fluoranthene	115	0.0 - 920.5	0.0 - 100	79
Benzo[k]fluoranthene	78.8	0.0 - 674.7	0.0 - 100	79
Benzo[e]pyrene	79.6	0.0 - 770.3	0.0 - 100	79
Benzo[a]pyrene	44.6	0.0 - 507.1	0.0 - 100	79
Indeno[123cd]pyrene	81.1	0.0 - 911	7.3 - 100	79
Dibenz[ah]anthracene	22.4	0.3 - 118	0.0 - 100	79
Benzo[ghi]perylene	85.9	0.0 - 1047.6	0.0 - 100	79
Total PCB				

Table III.4. Average annual concentrations of total (vapor plus particulate) organic contaminants (pg/m³), Haven Beach

Compound	Mean Conc.	Range	Range % Part.	N
1991				
Fluorene	1653.8	298.7 - 3671.8	0.2 - 1.4	27
Phenanthrene	3506.7	1218.4 - 8285.6	0.4 - 10.0	27
Anthracene	48.9	1.1 - 140.0	1.4 - 100.0	25
Fluoranthene	558.0	202.3 - 1591.5	2.7 - 67.3	27
Pyrene	622.1	177.4 - 1302.4	2.2 - 74.1	27
Benz[a]anthracene	31.7	3.2 - 213.6	14.9 - 100	24
Chrysene	122.2	13.7 - 548.5	9.7 - 97.1	27
Benzo[b]fluoranthene	83.9	1.9 - 601.8	56.8 - 100	26
Benzo[k]fluoranthene	67.0	1.5 - 416.1	56.8 - 100	24
Benzo[e]pyrene	53.6	2.7 - 375.3	29.1 - 100	27
Benzo[a]pyrene	36.9	1.7 - 262.4	66.8 - 100	25
Indeno[123cd]pyrene	92.1	7.1 - 443.3	27.8 - 100	22
Dibenz[ah]anthracene	19.4	8.8 - 51.0	38.5 - 100	9
Benzo[ghi]perylene	71.9	3.7 - 369.1	50.7 - 100	23
Total PCB	298.3	26 - 730	n/a	27
1992				
Fluorene	984.7	203.0 - 4722	0.0 - 100	26
Phenanthrene	2497.7	19 - 16509	0.0 - 100	26
Anthracene	14.7	1.6 - 175	1.1 - 100	26
Fluoranthene	567.7	11 - 2228	0.2 - 100	26
Pyrene	1095.6	11.2 - 4427	0.3 - 100	26
Benz[a]anthracene	24.6	0.6 - 206.2	9.7 - 100	26
Chrysene	112.9	13.7 - 850.3	8.5 - 100	26
Benzo[b]fluoranthene	67.1	5.9 - 407.7	26.1 - 100	26
Benzo[k]fluoranthene	53.1	4.9 - 375.1	24.4 - 100	26
Benzo[e]pyrene	59.5	6.4 - 356.6	15.0 - 100	26
Benzo[a]pyrene	34.8	2.1 - 368	10.8 - 100	26
Indeno[123cd]pyrene	62.5	3.6 - 583	14.5 - 100	26
Dibenz[ah]anthracene	12.9	1.0 - 85.2	15.1 - 100	26
Benzo[ghi]perylene	67.8	4.5 - 502.8	7.4 - 100	26
Total PCB				

Table III.4. Average annual concentrations of total (vapor plus particulate) organic contaminants (pg/m³), Haven Beach

Compound	Mean Conc.	Range	Range % Part.	N
1993				
Fluorene	562.8	147.5 - 1372.5	0.0 - 2.8	21
Phenanthrene	1684.7	36.1 - 3882.0	0.0 - 100	21
Anthracene	36.6	0.0 - 304.6	0.0 - 100	21
Fluoranthene	258.1	14.0 - 1183.4	0.0 - 100	21
Pyrene	498.6	16.8 - 2681.1	0.0 - 100	21
Benz[a]anthracene	10.4	0.0 - 49.0	0.0 - 100	21
Chrysene	55.8	9.7 - 109.8	16.9 - 100	21
Benzo[b]fluoranthene	39.9	2.1 - 129.8	83.2 - 100	21
Benzo[k]fluoranthene	33.8	1.3 - 129.8	0.0 - 100	21
Benzo[e]pyrene	34.0	4.9 - 102.2	84.5 - 100	21
Benzo[a]pyrene	14.1	0.8 - 56.7	77.0 - 100	21
Indeno[123cd]pyrene	29.5	0.7 - 116.4	100 - 100	21
Dibenz[ah]anthracene	3.3	0.0 - 12.9	0.0 - 100	21
Benzo[ghi]perylene	28.8	90.8 - 77	77.0 - 100	21
Total PCB	n/a	n/a	n/a	
Overall Average				
Fluorene	1106.4	147.5 - 4722	0.0 - 100	74
Phenanthrene	2642.9	19 - 16509	0.0 - 100	74
Anthracene	33.1	0.0 - 304.6	0.0 - 100	72
Fluoranthene	479.7	11.0 - 2228	0.0 - 100	74
Pyrene	760.6	11.2 - 4427	0.0 - 100	74
Benz[a]anthracene	23.1	0.0 - 213.6	0.0 - 100	71
Chrysene	100.7	9.7 - 850.3	8.0 - 100	74
Benzo[b]fluoranthene	65.8	1.9 - 601.8	261 - 100	73
Benzo[k]fluoranthene	52.9	1.3 - 416.1	0.0 - 100	71
Benzo[e]pyrene	50.4	2.7 - 375.3	15 - 100	74
Benzo[a]pyrene	29.9	0.8 - 368	10.8 - 100	72
Indeno[123cd]pyrene	64.3	0.7 - 583	14.5 - 100	69
Dibenz[ah]anthracene	12.7	0.0 - 85.2	0.0 - 100	56
Benzo[ghi]perylene	58.7	1.8 - 502.8	7.4 - 100	70

general, the magnitude of fluctuations in atmospheric levels of organic chemicals is larger than the corresponding variations in trace elements and sulfur discussed above, suggesting the importance of local combustion sources.

Annual average concentrations of the more volatile PAHs fluorene, phenanthrene, anthracene, and fluoranthene changed less than 20% between 1991 and 1992 (Tables III.2-4). Conversely, atmospheric concentrations of particulate-dominated PAHs increased two to three fold at the mid-bay site between 1991 and 1992. No comparable increase was observed at the southern site, suggesting perhaps that the mid-bay site was impacted by increased local sources of combustion during 1992. Annual average concentrations of PAHs and t-PCB were approximately 20% higher in the southern Chesapeake Bay relative to the mid Bay in 1991. Conversely, we observed substantially higher concentrations of PAHs at the mid-bay site in 1992. Annual average particulate PAH concentrations in 1992 were approximately 200-250 percent and 20-70 percent higher at the mid-bay site than at the southern and northern sites, respectively. PAHs which occur predominately in the gas phase were less spatially variable than those associated with particles. Unfortunately, few air samples for organic contaminants were collected at the northern site in 1991. Nonetheless, higher concentrations of PAHs at the mid-bay site relative to the southern site are consistent with the trends observed for the atmospheric levels of trace elements.

III.2. Concentrations in Precipitation. The volume-weighted mean (VWM) concentrations of trace elements in precipitation collected at the three sites is summarized in Table III.5. Overall annual VWM concentrations range from 0.05 $\mu\text{g/L}$ for arsenic at Wye to 16.8 $\mu\text{g/L}$ for aluminum at Haven Beach. The relative proportion of trace elements in precipitation is nearly identical to that in Chesapeake Bay aerosol particles (Figure III.1), confirming that aerosol scavenging is the source of trace metals to wet deposition. Trace metal wet depositional fluxes are highly variable, changing more than ten-fold between consecutive months with little apparent seasonal dependence (Figure III.4). This variability, which was similar at each of the three sites, results not only from fluctuations in the atmospheric inventories of trace metals, but also from changes in the amount of precipitation. On an annual basis, the volume-weighted mean concentrations of most trace metals did not systematically vary between the Summer 1990 and Fall 1993 (Table III.5), again suggesting that these rural sites were most strongly influence by the same regional background signal throughout the study.

While individual precipitation events result in spikes in trace metal deposition at one site that are not observed at the other two locations (Figure III.4), these isolated events average out over the year. Annual volume-weighted mean concentrations of trace metals in precipitation are generally within a factor of two among the three sites, with no clear systematic spatial trend observed for all metals. Concentrations of soil component metals aluminum and iron in precipitation are only one half as high at Elms as at the northern and southern sites, perhaps reflecting less localized inputs from agricultural activities at this sites. The overall volume-weighted mean concentrations of copper, nickel, lead and zinc are 50-100% higher at the southern Haven Beach site relative to the other sites. This enrichment was not observed in aerosol particles. It should be noted that the Haven Beach trace metal measurements contained several

Table III.5. Volume-weighted mean concentrations of metals in precipitation, Chesapeake Bay Atmospheric Deposition Study.

Compound	Jun 1990- Dec 1991	1991 VWM	1992 VWM	Jan 1993- Sept 1993	Overall VWM
Wye					
Aluminum	15.8	17.5	12	12.7	13.9
Arsenic	0.05	0.06	0.05	0.06	0.05
Cadmium	0.04	0.05	0.04	0.04	0.04
Chromium	0.14	0.11	0.25	0.13	0.17
Copper	0.28	0.23	0.30	0.35	0.30
Iron	13.2	13.4	11.9	12.5	12.6
Manganese	1.42	1.5	1.34	1.64	1.45
Nickel	0.29	0.29	0.42	0.40	0.36
Lead	0.64	0.69	0.43	0.43	0.52
Selenium	0.16	0.14	0.10	0.06	0.12
Zinc	1.58	1.47	1.89	1.8	1.73
Elms					
Aluminum	7.41	7.73	6.01	NA	6.86
Arsenic	0.05	0.05	0.05	NA	0.05
Cadmium	0.03	0.05	0.04	NA	0.04
Chromium	0.05	0.07	0.01	NA	0.04
Copper	0.23	0.2	0.49	NA	0.34
Iron	5.76	5.97	4.82	NA	5.39
Manganese	0.82	0.82	1.01	NA	0.89
Nickel	0.20	0.21	0.15	NA	0.18
Lead	0.54	0.57	0.27	NA	0.43
Selenium	0.12	0.11	0.07	NA	0.09
Zinc	1.17	1.26	1.79	NA	1.41
Haven Beach					
Aluminum	NA	21.3	12.06	18.1	16.8
Arsenic	NA	0.07	0.07	NA	0.07
Cadmium	NA	0.18	0.14	0.06	0.13
Chromium	NA	ND	ND	ND	ND
Copper	NA	0.82	0.51	0.37	0.57
Iron	NA	13.9	22.8	12.1	16.0
Manganese	NA	1.25	1.2	1.11	0.51
Nickel	NA	0.21	2.63	0.34	1.16
Lead	NA	0.69	0.4	0.49	1.03
Selenium	NA	0.13	0.17	NA	0.15
Zinc	NA	6.29	5.84	4.74	5.48

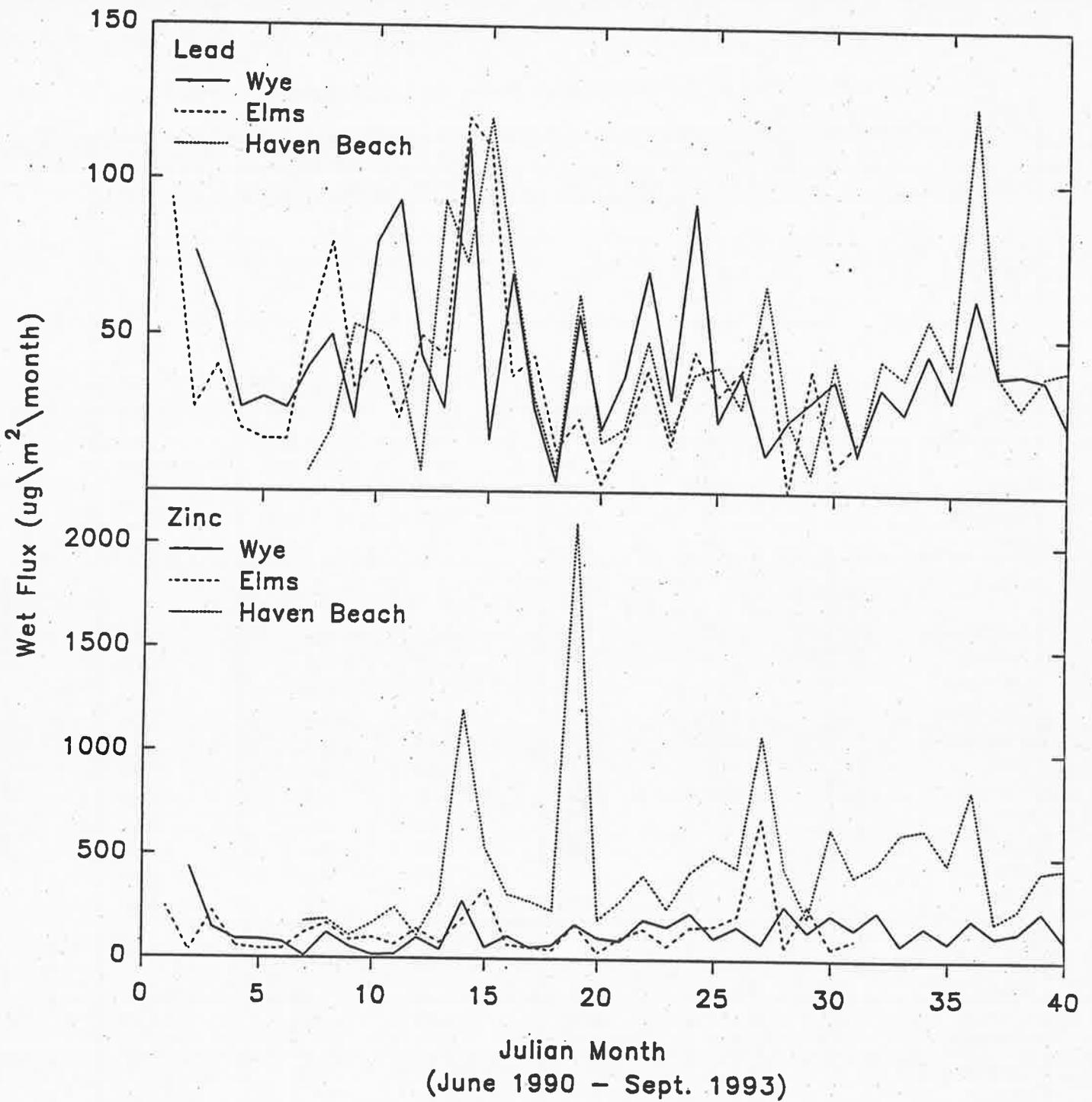


Figure III.4 Lead and zinc wet depositional fluxes at the Wye, Elms, and Haven Beach sites.

extremely high values which were included in these calculation. Whether these values reflect a local source of trace metals to the precipitation or an analytical artifact is unclear.

Overall volume-weighted mean concentrations of polycyclic aromatic hydrocarbons in precipitation ranged from 0.3 ng/L for anthracene and dibenz[*a,h*]anthracene at the Elms site to 9 ng/L for pyrene at the Wye site (Table III.6). Volume-weighted mean concentrations for total PCBs were 1.1 and 0.9 ng/L at the Wye and Elms sites, and for a shorter period of time, 0.35 ng/L at the Haven Beach site. As was observed for trace metals, wet depositional fluxes of organic contaminants varied considerable with time, dominated by episodic spikes at each location (Figure III.5). Extremely high concentrations of some analytes, including pyrene, that were measured in both air and precipitation at the Elms site in Summer, 1990 may have resulted from local vegetation burning. Unlike PAH and PCB levels in the atmosphere, concentrations in precipitation do not systematically vary with season. Based on the overall volume-weighted mean concentrations, levels of PAHs and PCBs are higher in precipitation collected at the mid-bay Elms site than at the northern or southern locations. While this trend is influenced by the few elevated concentrations measured at Elms during Summer 1990 when the other two sites were not operating, the same spatial variation is seen when comparing concentrations from the same period at the three sites (Table III.6). The enrichment at Elms is especially pronounced for higher molecular weight PAHs, implicating a local combustion source (*i.e.*, wood burning for residential heating).

Concentrations of organic contaminants in precipitation measured in this study are consistently lower than those observed in the Great Lakes region (Figure III.6). Polycyclic aromatic hydrocarbon concentrations are two to three times higher at the three Integrated Atmospheric Deposition Network (IADN) sites located at rural, shoreline sites on Lakes Ontario, Michigan, and Superior relative to the Chesapeake Bay data (Gatz *et al.*, 1994). In contrast, levels of the same PAHs in the atmosphere over the Great Lakes is equal to or perhaps less than those measured over the Chesapeake Bay. Whether the apparent enrichment in PAHs in Great Lakes precipitation relative to that in the Chesapeake Bay region shown in Figure III.6 is due to more efficient scavenging in the colder, relatively drier region, or simply reflects methodological differences between the two networks is unclear.

III.3 Wet and Dry Aerosol Depositional Fluxes. Annual wet depositional fluxes ($\mu\text{g}/\text{m}^2\text{-year}$) of each chemical were calculated at each sampling site as follows. Monthly volume weighted mean concentrations were calculated by dividing the total mass of the analyte measured in rainfall by the amount of precipitation collected for analysis each month. This monthly volume weighted mean concentration was then multiplied by the total amount of precipitation which fell at the site during the month, as determined by a Belfort gauge (northern and mid-bay sites) or by the total sample volume (southern site). In cases where the precipitation sampling period spanned the end of a month, the precipitation gauge record was used to partition the wet flux between two months. This method of calculating monthly fluxes accounts for missing samples (*e.g.* due to sampler malfunction or loss of sample during analysis) by assuming that the missing samples contained concentrations of analytes equal to the monthly volume weighted mean concentration.

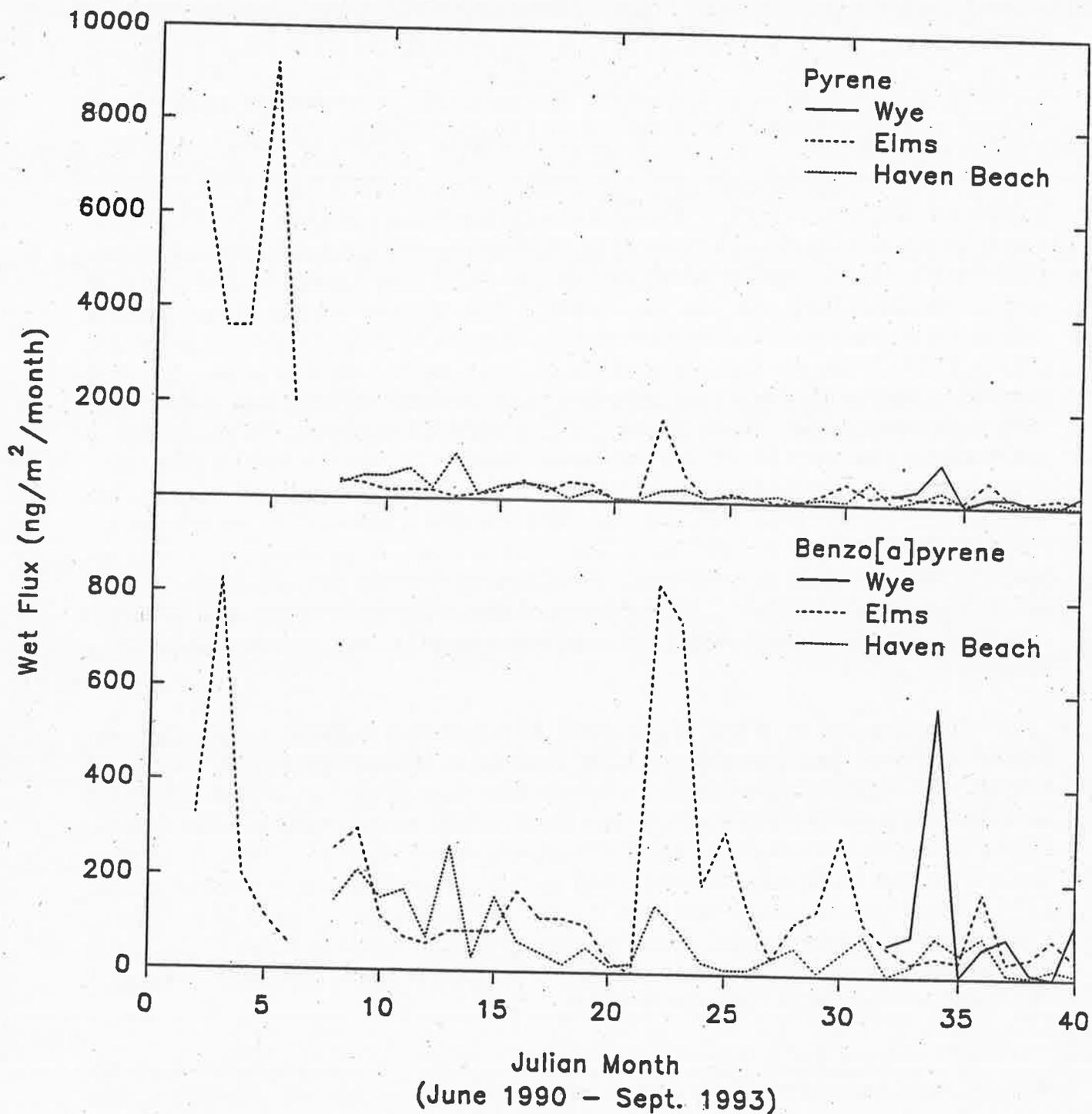


Figure III.5 Pyrene and Benzo[a]pyrene wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites.

Annual wet depositional fluxes ($\mu\text{g}/\text{m}^2\text{-year}$) were calculated by summing the twelve monthly fluxes.

Dry aerosol depositional fluxes of particulate-borne contaminants were calculated as the product of the atmospheric concentration of each chemical and a depositional velocity (V_d) appropriate to water surfaces (Wu *et al.*, 1992; Wu *et al.*, 1994). The magnitude of the depositional flux depends directly on the value of V_d , which is highly uncertain. V_d depends upon the size distribution and composition of the aerosol particles, upon atmospheric stability and turbulence, and upon characteristics of the receiving surfaces. Aerosol deposition rates over water are likely lower than those over land because of less turbulence over the smooth water surface (Williams, 1982). Deposition to water may be enhanced during unstable conditions (*e.g.*, when the air is colder than the water), as mixing and vertical transport through the air-water boundary layer is rapid. Williams (1982) suggests that the formation of whitecaps during high wind stress greatly increases aerosol deposition to the water surface. Contaminant deposition velocities also depend upon the aerosol size distribution, as very large (*i.e.*, $> 1 \mu\text{m}$) and very small (*i.e.*, $< 0.1 \mu\text{m}$) particles are rapidly deposited by gravitational settling and diffusional impaction, respectively, while particles in the 'accumulation mode' (*i.e.*, between 0.1 and $1 \mu\text{m}$) are poorly removed. Anthropogenically-mobilized elements and organic contaminants typically associate with aerosol particles with mass median diameters ranging from 0.2 to $> 0.5 \mu\text{m}$, whereas crustal components (*e.g.*, wind-eroded soils) reside on particles larger than $1 \mu\text{m}$.

In this study, a deposition velocity of 0.26 cm/sec was used to estimate the dry aerosol depositional flux of the non-crustal components of aerosol trace elements. This value, taken from the study in the North Sea by Dedeurwaerder *et al.* (1983), is on the lower end of elemental V_d 's reported in the literature, likely resulting in conservative estimates of the true dry aerosol depositional fluxes. The crustal components of the aerosol were modeled using a depositional velocity of 1.4 cm/sec. Depositional fluxes of those elements that have both crustal and non-crustal components (*i.e.*, chromium, iron, manganese, vanadium) were estimated as linear combinations of the two fluxes, using the aluminum-normalized crustal abundances of Turekian and Wedepohl (1961). Actual deposition velocities likely change dramatically over space and time, especially driven by changing meteorological conditions and atmospheric stability. In this study, we did not have the necessary meteorological data (nor the fundamental understanding) to calculate site and time-specific V_d 's. Because our primary goal is to estimate annual loadings to the Chesapeake Bay, we chose as a first approximation to use the single crustal and non-crustal deposition velocities described above and annual average aerosol concentrations in order to estimate annual dry aerosol fluxes at the three sites.

Wet, dry aerosol, and total depositional fluxes of trace elements at each of the three sites are summarized in Table III.7. Total fluxes in 1992 ranged from 0.07 $\text{mg}/\text{m}^2\text{-year}$ for cadmium at Wye to 121 $\text{mg}/\text{m}^2\text{-year}$ for aluminum at Elms. Not surprisingly, dry aerosol deposition comprises the majority of the total flux for the soil elements aluminum and iron which occur on coarse particles. Wet deposition contributes between one-third and one-half of the total depositional flux of the remaining trace elements. Naturally, spatial trends in total deposition

Table III.7. Trace metal depositional fluxes, Chesapeake Bay Atmospheric Deposition Study.

Wye	Wet Flux (mg/m ² -year)				Dry Aerosol Flux (mg/m ² -year)				Total Annual Flux (mg/m ² -year)	
	Jun 1990- Dec 1991	1992	Jan. 1993- Sept 1993	Jun 1990- Sept 1993	Jun 1990- Dec 1991	1992	Jun 1990- Dec 1992	Jun 1990- Dec 1991	1992	Jun 1990- Dec 1992
Aluminum	13.0	12.0	13.1	12.7	97.0	101	98.6	110	113	111.7
Arsenic	0.043	0.052	0.062	0.050	0.102	0.104	0.103	0.145	0.156	0.165
Cadmium	0.037	0.037	0.040	0.038	0.022	0.034	0.027	0.059	0.071	0.067
Chromium	0.103	0.247	0.132	0.154	0.190	0.211	0.199	0.293	0.458	0.331
Copper	0.201	0.296	0.357	0.266	0.355	0.299	0.331	0.556	0.595	0.688
Iron	10.4	11.9	12.9	11.4	53.5	55.4	54.3	63.9	67.3	67.2
Manganese	1.14	1.34	1.69	1.33	1.15	1.19	1.16	2.29	2.53	2.85
Nickel	0.231	0.416	0.415	0.330	0.508	0.729	0.599	0.739	1.15	1.01
Lead	0.521	0.425	0.445	0.474	0.617	0.912	0.739	1.14	1.34	1.18
Selenium	0.121	0.097	0.066	0.101	0.247	0.278	0.259	0.368	0.375	0.325
Zinc	1.21	1.88	1.82	1.56	2.15	2.09	2.12	3.36	3.97	3.94

Table III.7 (continued). Trace metal depositional fluxes, Chesapeake Bay Atmospheric Deposition Study.

	Wet Flux (mg/m ² -year)			Dry Aerosol Flux (mg/m ² -year)			Total Annual Flux (mg/m ² -year)		
	Jun 1990- Dec 1991	1992	Jun 1990- Dec 1992	Jun 1990- Dec 1991	1992	Jun 1990- Dec 1992	Jun 1990- Dec 1991	1992	Jun 1990- Dec 1992
Elms									
Aluminum	7.05	6.86	6.96	131	114	124	138	121	131
Arsenic	0.047	0.054	0.050	0.098	0.090	0.094	0.145	0.144	0.144
Cadmium	0.037	0.046	0.041	0.022	0.029	0.024	0.059	0.075	0.065
Chromium	0.056	0.011	0.035	0.199	0.205	0.201	0.255	0.216	0.236
Copper	0.209	0.559	0.376	0.335	0.363	0.346	0.544	0.922	0.722
Iron	5.47	5.51	5.49	71.8	59.7	66.8	77.3	65.2	67.3
Manganese	0.582	1.15	0.851	1.37	1.21	1.31	1.95	2.36	2.16
Nickel	0.227	0.173	0.201	0.462	0.569	0.506	0.689	0.742	0.707
Lead	0.515	0.306	0.416	0.576	0.774	0.658	1.09	1.08	1.07
Selenium	0.108	0.079	94.2	0.266	0.251	0.260	0.374	0.33	94.5
Zinc	1.13	2.04	1.56	2.04	2.10	2.06	3.17	4.14	3.62

Table III.7 (continued). Atmospheric depositional fluxes of elements, Chesapeake Bay Atmospheric Deposition Study

	Wet Flux (mg/m ² -year)				Dry Aerosol Flux (mg/m ² -year)			Total Annual Flux (mg/m ² -year)	
	1991	1992	Jan 93- Sept 1993	1991- 1993	1991	1992	1991-1992	1991	1992
Haven Beach									
Aluminum	19.1	10.9	22.0	18.9	110	69.8	98.6	129	80.7
Arsenic	0.06	0.06	NA	0.04	0.083	0.092	0.10	0.14	0.15
Cadmium	0.16	0.13	0.069	0.13	0.022	0.028	0.027	0.18	0.16
Chromium	ND	ND	NA	NA	0.18	0.12	0.20	0.18*	0.12*
Copper	0.74	0.46	0.45	0.60	0.37	0.32	0.33	1.11	0.78
Iron	12.4	20.6	14.7	17.4	58.5	36.6	54.3	70.9	57.2
Manganese	1.12	1.08	1.35	1.29	1.22	0.79	1.16	2.34	1.87
Nickel	0.19	2.38	0.41	1.08	0.48	0.53	0.60	0.67	2.91
Lead	0.62	0.36	0.60	0.57	0.48	0.65	0.74	1.1	1.01
Selenium	0.12	0.16	NA	0.10	0.25	0.22	0.26	0.37	0.38
Zinc	5.63	5.28	5.78	6.07	1.77	1.69	2.12	7.40	6.97

*Dry deposition only

result from variation in precipitation chemistry and amount, and the trace element inventories associated with aerosol particles (given the considerable uncertainty in the dry aerosol deposition calculation, the same deposition velocity was used at each site). Although a distinct north to south trend in precipitation amount occurred in 1992 (100, 107, and 122 cm, respectively), total annual depositional fluxes were remarkably similar among the three stations. As shown in Table III.7, total depositional fluxes were also very similar between years, again indicating that the relatively rare spikes in concentration are dampened against the chronic regional background-signal.

Depositional fluxes of hydrophobic organic chemicals are shown in Table III.8. It is important to remember that these flux estimates do not include passive exchange of gaseous organic contaminants across the air-water interface. Other recent studies have shown that this is the *dominant* atmospheric deposition process for volatile organic contaminants, including PCBs (Baker and Eisenreich, 1990; Achman *et al.*, 1993) and low molecular weight PAHs (Nelson *et al.*, 1995). In those studies, the net direction of exchange is often from the water to the air and diffusive gas exchange is large enough to offset wet deposition and dry aerosol deposition. Therefore, the depositional fluxes presented here (Table III.8) should be considered gross fluxes to the water surface, rather than net exchanges. Total fluxes for 1992 range from 0.8 $\mu\text{g}/\text{m}^2\text{-year}$ for anthracene at Elms to 10.8 $\mu\text{g}/\text{m}^2\text{-year}$ for benzo[b]fluoranthene at Elms (the precipitation sampling record at Wye is incomplete for 1992). Both wet deposition and dry aerosol deposition contribute to total PAH deposition, with dry aerosol deposition becoming relatively more important for the higher molecular weight, less volatile compounds (Table III.8). Because aerosol particle-associated PCBs were present below our analytical detection limits, estimates of PCB dry deposition were made using an aerosol PCB level calculated from the measured gaseous PCB concentration and the Junge-Pankow sorption model (Pankow, 1987; Leister and Baker, 1994). The total depositional flux of total PCBs is about 3.5 $\mu\text{g}/\text{m}^2\text{-year}$, with approximately equal contribution from wet and dry aerosol deposition. Total depositional fluxes of PAHs decrease with time during this study, with the lowest fluxes measured during the first nine months of 1993 (Table III.8). While some of this decrease is attributed to beginning with anomalously high measurement in Summer 1990, we continued to see decreases in both wet and dry aerosol fluxes between 1992 and 1993. Overall, total annual fluxes are generally within 50% among the sites, which given the uncertainty in the dry aerosol deposition estimates, indicates little spatial variability when integrating over annual cycles. One should remember, however, that this study did not specifically address the possible influences of urban areas on atmospheric deposition, and gradient near the cities of Baltimore, Washington, and Norfolk may be important.

To place the atmospheric deposition fluxes measured in this study in perspective, they are compared to similar estimates made for the Great Lakes region (Figure III.7; Eisenreich and Strachan, 1992). Annual Great Lakes fluxes were estimated for each lake based upon several sources of data, including the initial results of the IADN program, and were averaged in Figure III.7. Wet depositional fluxes of trace elements are up to three times higher to the Great Lakes than to the Chesapeake Bay, despite significantly less rainfall (80 *versus* 110 cm/year). In contrast, wet depositional fluxes of total PCBs and PAHs are remarkably similar between the two regions, as higher measured concentrations in the Great Lakes (Figure III.6) are offset by lower

Table III.8. Wet, dry aerosol, and total depositional fluxes of organic contaminants, Chesapeake Bay Atmospheric Deposition Study, 1990-1993.

Wye	Wet Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Dry Aerosol Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Total Flux ($\mu\text{g}/\text{m}^2\text{-year}$)			
	Jun 1990- Dec 1991	Jan 1993- Sept 1993	Jun 1990- Sept 1993	1992	Jun 1990- Dec 1991	1992*	Jan 1993- Sept 1993*	Jun 1990- Sept 1993	Jun 1990- Dec 1991	1992	Jan 1993- Sept 1993	Jun 1990- Sept 1993
Fluorene	NA	0.8		NA	NA	0.3	0.4	0.4	NA	0.3	1.2	1.2
Phenanthrene	NA	3.8		NA	NA	2.2	4.6	3.6	NA	2.2	8.4	7.4
Anthracene	NA	0.5		NA	NA	0.2	0.6	0.4	NA	0.2	1.1	0.9
Fluoranthene	NA	3.3		NA	NA	2.0	3.9	3.2	NA	2.0	7.3	6.5
Pyrene	NA	2.2		NA	NA	2.0	3.4	2.8	NA	2.0	5.6	5.1
Benz[a]anthracene	NA	0.6		NA	NA	0.7	1.4	1.1	NA	0.7	1.09	1.8
Chrysene	NA	1.5		NA	NA	1.9	3.8	3.0	NA	1.9	5.24	4.5
Benzofluoranthene	NA	1.7		NA	NA	2.0	2.9	2.6	NA	2.0	4.63	4.3
Benzokjfluoranthene	NA	0.9		NA	NA	1.5	2.5	2.1	NA	1.5	3.34	2.9
Benzofluoranthene	NA	2.3		NA	NA	1.8	2.7	2.3	NA	1.8	4.97	4.5
Benzofluoranthene	NA	1.0		NA	NA	1.1	1.3	1.2	NA	1.1	2.29	2.2
Indeno[1,2,3-cd]perylene	NA	1.2		NA	NA	1.1	2.1	1.7	NA	1.1	3.32	2.9
Dibenz[a,h]anthracene	NA	0.4		NA	NA	0.4	0.6	0.5	NA	0.4	1.03	0.9
Benzofluoranthene	NA	1.0		NA	NA	1.2	2.4	1.9	NA	1.2	3.33	2.9
Total PCBs	NA	0.2		NA	NA	NA	NA	NA	NA	NA	0.24	0.2

*Calculated from geometric mean.

Table III.8 (continued). Wet, dry aerosol, and total depositional fluxes of organic contaminants, Chesapeake Bay Atmospheric Deposition Study, 1990-1993.

	Wet Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Dry Aerosol Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Total Flux ($\mu\text{g}/\text{m}^2\text{-year}$)			
	Jun 1990- Dec 1991	Jan 1993- Sept 1993	Jun 1990- Sept 1993	1992	Jun 1990- Dec 1991	1992*	Jan 1993- Sept 1993*	Jun 1990- Sept 1993	Jun 1990- Dec 1991	1992	Jan 1993- Sept 1993	Jun 1990- Sept 1993
Elms												
Fluorene	2.6	0.9	1.6	0.6	1.6	0.8	0.6	4.2	1.4	1.5	2.7	
Phenanthrene	8.9	4.1	5.7	2.4	12	7.0	6.2	21	9.4	10.2	14.8	
Anthracene	1.3	0.3	0.8	0.4	0.9	0.5	0.5	2.2	0.8	0.8	1.4	
Fluoranthene	15	3.2	8.3	2.6	16	6.0	5.0	31	8.6	8.2	18.5	
Pyrene	20	2.3	10.2	2.3	15	5.8	4.4	35	8.0	6.8	19.8	
Benz[a]anthracene	1.7	0.6	1.1	0.5	5.6	2.5	1.2	7.3	3.0	1.8	4.6	
Chrysene	5.3	1.7	3.4	1.8	12	7.0	3.0	17	8.8	4.7	11.6	
Benzo[b]fluoranthene	7.6	2.0	4.7	2.6	15	8.2	3.1	23	10.8	5.1	14.7	
Benzo[k]fluoranthene	5	1.1	2.8	1.0	8.2	5.1	2.4	13	6.1	3.5	8.7	
Benzo[e]pyrene	2.3	2.3	2.2	1.9	9	6.6	3.0	11	8.5	5.3	9.0	
Benzo[a]pyrene	2.4	0.7	2.0	2.5	4.3	3.4	1.4	6.7	5.8	2.0	5.3	
Indeno[1,2,3-cd]perylene	2.1	2.4	1.9	1.3	9.6	5.5	2.3	12	6.8	4.6	8.5	
Dibenz[a,h]anthracene	1.1	0.6	0.8	0.5	1.8	0.9	0.4	2.9	1.4	1.0	2.0	
Benzo[ghi]perylene	2.4	2.3	2.0	1.2	1.7	6.4	3.2	2.9	7.6	5.5	5.5	
Total PCBs	1.9	0.4	1.2	0.7	1.4	NA	NA	3.3	NA	NA	NA	

*Calculated from geometric mean.

Table III.8 (continued). Wet, dry aerosol, and total depositional fluxes of organic contaminants, Chesapeake Bay Atmospheric Deposition Study, 1990-1993.

	Wet Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Dry Aerosol Flux ($\mu\text{g}/\text{m}^2\text{-year}$)				Total Flux ($\mu\text{g}/\text{m}^2\text{-year}$)			
	Jun 1990- Dec 1991	Jan 1993- Sept 1993	Jun 1990- Sept 1993	Jun 1990- Dec 1991	1992*	Jan 1993- Sept 1993*	Jun 1990- Sept 1993	Jun 1990- Dec 1991	1992	Jan 1993- Sept 1993	Jun 1990- Sept 1993	
	1992	1992	1992	1992	1992	1992	1992	1992	1992	1992	1992	
Haven Beach												
Fluorene	1.8	0.7	1.1	1.5	0.6	0.7	1.0	3.3	1.3	1.3	2.0	
Phenanthrene	8.0	3.0	4.8	12	5.1	4.0	7.3	20	8.1	7.1	12.1	
Anthracene	0.4	0.2	0.3	0.74	0.2	0.2	0.4	1.2	0.4	0.4	0.7	
Fluoranthene	5.4	3.3	3.7	21	3.5	2.4	9.6	26	6.8	4.5	13.3	
Pyrene	4.3	1.8	2.6	18	4.4	2.4	8.8	22	6.2	3.6	11.4	
Benz[a]anthracene	0.8	0.4	1.0	4.2	1.3	1.1	2.3	5	1.6	3.1	3.3	
Chrysene	2.3	1.1	1.6	11	4.9	3.1	6.6	13	6.0	4.6	8.3	
Benzo[b]fluoranthene	2.0	1.1	1.5	13	4.9	2.5	7.2	15	6.1	3.8	8.7	
Benzo[k]fluoranthene	1.5	0.9	1.0	10	3.5	1.9	5.4	11	4.4	2.5	6.5	
Benzo[e]pyrene	1.5	1.0	1.4	8	4.4	2.4	5.2	9.5	5.4	4.0	6.5	
Benzo[a]pyrene	1.3	0.6	0.8	5.5	1.7	0.9	2.9	6.8	2.3	1.3	3.7	
Indeno[1,2,3-cd]perylene	2.2	1.0	1.5	14	3.2	1.7	6.7	16	4.2	2.8	8.2	
Dibenz[a,h]anthracene	0.4	0.4	0.4	3	0.5	0.4	1.4	3.4	1.0	0.7	1.8	
Benzo[ghi]perylene	2.2	0.6	1.2	10	3.9	2.2	5.6	12	4.5	2.9	6.8	
Total PCBs	1.6	0.8	1.1	2.1	NA	NA	NA	3.7	NA	NA	NA	

*Calculated from geometric mean.

Table III.6. Volume-weighted mean concentrations of organic contaminants in precipitation, Chesapeake Bay Atmospheric Deposition Study.

Compound	Jun 1990- Dec 1991	1991 VWM	1992 VWM	Jan 1993- Sept 1993	Overall VWM
Elms					
Fluorene	1.8	1.4	0.58	0.88	1.2
Phenanthrene	6.7	5.3	2.25	4.18	4.7
Anthracene	1.2	0.7	0.34	0.28	0.7
Fluoranthene	9	4.9	2.46	3.27	5.7
Pyrene	16.8	2.6	2.11	2.39	9.0
Benz[a]anthracene	1.3	0.9	0.48	0.59	0.9
Chrysene	3.4	2.7	1.71	1.76	2.5
Benzo[b]fluoranthene	5.7	3.1	2.40	2.05	3.9
Benzo[k]fluoranthene	3.1	1.5	0.89	1.17	2.0
Benzo[e]pyrene	2.4	2.1	1.74	2.32	2.2
Benzo[a]pyrene	1.9	1.2	2.32	0.69	1.8
Indeno[1,2,3-cd]perylene	2.1	2.4	1.21	2.42	1.9
Dibenz[a,h]anthracene	1.1	0.8	0.43	0.61	0.8
Benzo[g,h,i]perylene	2.3	2.4	1.13	2.36	1.9
Total PCBs	1.6	0.2	0.66	0.42	1.1
Haven Beach					
Fluorene	NA	1.78	0.57	0.53	1.0
Phenanthrene	NA	7.91	2.44	2.38	4.3
Anthracene	NA	0.42	0.16	0.22	0.3
Fluoranthene	NA	4.67	2.69	2.01	3.2
Pyrene	NA	3.57	1.49	1.28	2.1
Benz[a]anthracene	NA	0.68	0.29	0.36	0.4
Chrysene	NA	2.07	0.91	1.09	1.4
Benzo[b]fluoranthene	NA	1.77	0.92	1.24	1.3
Benzo[k]fluoranthene	NA	1.28	0.72	0.62	0.9
Benzo[e]pyrene	NA	1.29	0.83	1.63	1.2
Benzo[a]pyrene	NA	1.14	0.45	0.39	0.7
Indeno[1,2,3-cd]perylene	NA	1.9	0.82	1.06	1.2
Dibenz[a,h]anthracene	NA	0.33	0.35	0.34	0.3
Benzo[g,h,i]perylene	NA	1.8	0.48	0.78	1.0
Total PCBs	NA	1.12	0.64	0.87	0.9
Wye					
Fluorene	NA	NA	NA	0.65	0.65
Phenanthrene	NA	NA	NA	3.03	3.03
Anthracene	NA	NA	NA	0.50	0.50
Fluoranthene	NA	NA	NA	2.84	2.84
Pyrene	NA	NA	NA	1.94	1.94
Benz[a]anthracene	NA	NA	NA	0.56	0.56
Chrysene	NA	NA	NA	1.27	1.27
Benzo[b]fluoranthene	NA	NA	NA	1.49	1.49
Benzo[k]fluoranthene	NA	NA	NA	0.76	0.76
Benzo[e]pyrene	NA	NA	NA	1.92	1.92
Benzo[a]pyrene	NA	NA	NA	0.63	0.63
Indeno[1,2,3-cd]perylene	NA	NA	NA	1.03	1.03
Dibenz[a,h]anthracene	NA	NA	NA	0.35	0.35
Benzo[g,h,i]perylene	NA	NA	NA	0.82	0.82
Total PCBs	NA	NA	NA	0.35	0.35

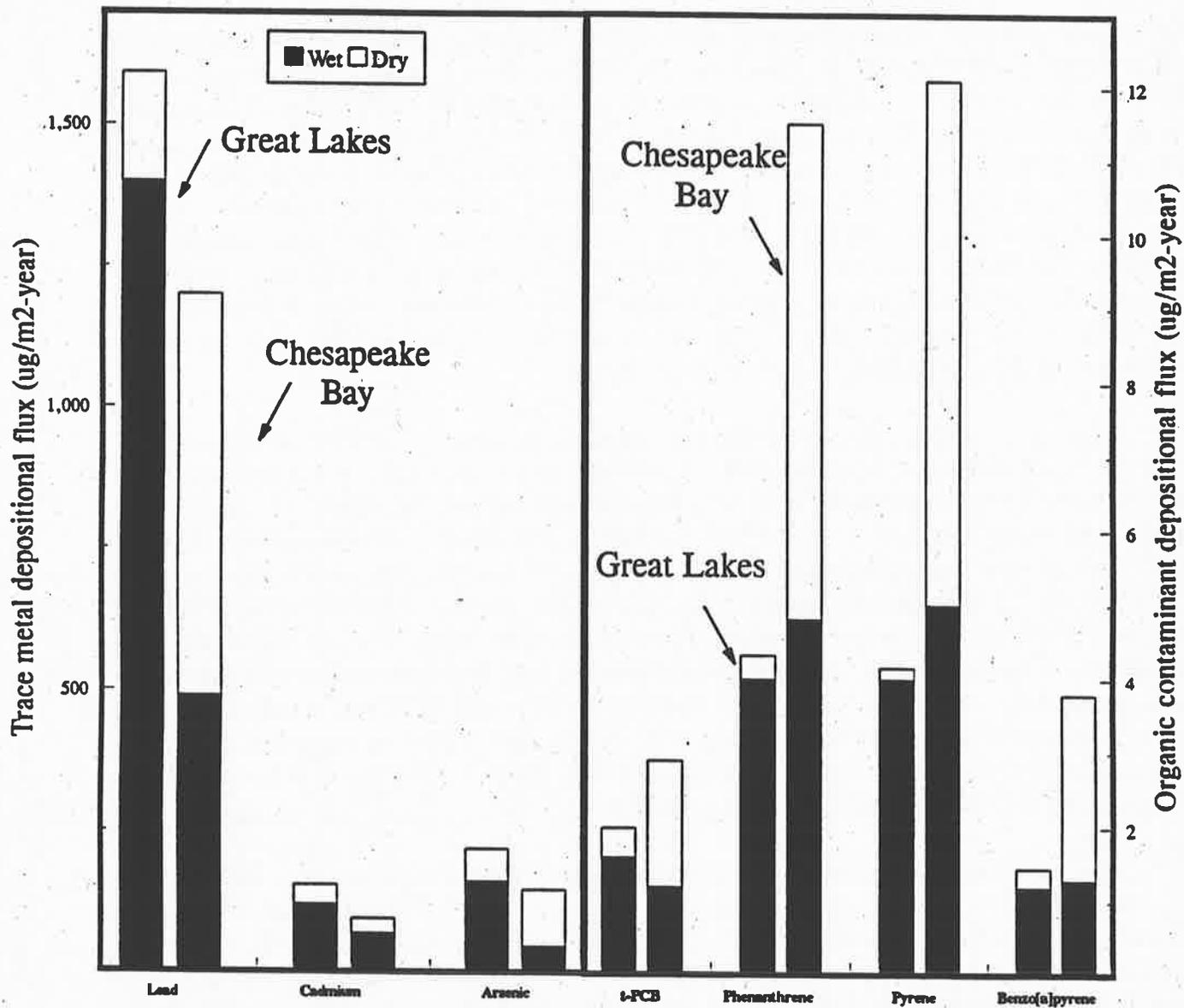


Figure III.7. Comparison of Chesapeake Bay (this study) and Great Lakes (Eisenreich and Strachan, 1992) atmospheric depositional fluxes.

precipitation amounts. Dry aerosol depositional flux estimates are higher in the Chesapeake region, especially for organic contaminants. This is due, in part, to the choice of deposition velocities used for the two estimates (0.2 cm/sec for all species in the Great Lakes, 0.26 cm/sec for elements and 0.49 cm/sec for organics in the Chesapeake Bay). While all of these values are within the generally accepted range for dry aerosol deposition velocities, the impact of this uncertainty can be seen in Figure III.7. In addition, measured aerosol-bound organic concentrations were generally higher than the values used in the Great Lakes dry aerosol deposition calculations (Eisenreich and Strachan, 1992). Despite the differences, estimated atmospheric depositional fluxes are generally within a factor of two between the Chesapeake Bay and Great Lakes regions, which, given the numerous opportunities for error in these measurements and calculations, is quite good agreement.

III.4. Annual Chesapeake Bay-Wide Atmospheric Loadings. To estimate the annual, bay-wide loading of elements and organic contaminants to the entire Chesapeake Bay, the annual site-specific wet and dry aerosol fluxes were averaged and these two average fluxes were multiplied by the surface area of the bay ($1.15 \times 10^{10} \text{ m}^2$). While we could have chosen the partition the area of the bay into three sections represented by each site, simply averaging the three fluxes provides approximately the same estimate, because of the nearly equal spacing of the three sites along the mainstem of the bay and the similar fluxes measured at the sites. Again, because of our inability to estimate dry aerosol loadings on a finer temporal resolution, bay-wide loadings are presented on an annual basis. Fluxes and loadings are presented for 1991 and 1992, and averaged over the entire study period (mid 1990 through mid 1993). While these loading estimates are presented with three significant figures, propagated uncertainties are likely on the order of a factor of two, due primarily to the dry aerosol deposition estimates.

Bay wide atmospheric loadings of aluminum and iron are 1,340,000 and 799,000 kg/year respectively (Table III.9). Loadings of trace elements range from 1110 kg/year for cadmium to 49,400 kg/year for nickel. Loading estimates are generally similar for 1991 and 1992, except for large nickel and zinc due elevated wet deposition measured at the Haven Beach site in 1992. Loadings of PAHs range from 13 kg/year for anthracene to 189 kg/year for fluoroanthene. Total PCB loadings are estimated to be 37 kg/year. Interestingly, both wet deposition and dry aerosol depositional fluxes appear to decrease between 1991 and 1992. Whether this reflects a real interannual variation or simply results from aggregating measurements from different locations is unclear.

To place these loadings in perspective, they are compared to recent estimates of trace metal and organic contaminant loadings delivered to the Chesapeake Bay by the Susquehanna River in Table III.10 (Godfrey *et al.*, 1995; Foster, 1995; Conko, 1995). The Susquehanna River is the largest tributary to the Chesapeake Bay, supplying approximately 60% of the freshwater inflow to the estuary. Annual riverine loadings of dissolved and particulate trace metals and organic contaminants were determined by analyzing flow-weighted samples collected at Conowingo, Maryland between February, 1994 and January, 1995 (Foster, 1995; Conko, 1995). Atmospheric deposition directly to the surface waters of the Chesapeake Bay supplies loads of polycyclic

Table III.9. Annual loadings of trace metals and organic contaminants to the Chesapeake Bay

	Wet Deposition (kg/year)			Dry Deposition (kg/year)			Total Deposition (kg/year)		
	1991	1992	Overall	1991	1992	Overall	1991	1992	Overall
Fluorene	23	7	16	16	7	12	39	14	27
Phenanthrene	85	31	63	127	55	92	211	86	155
Anthracene	7.0	3	6	8.9	3	6	16	7	13
Fluoranthene	61	34	70	176	44	119	237	78	189
Pyrene	43	24	75	147	47	109	190	70	184
Benz[a]anthracene	10	5	9	47	17	34	57	22	44
Chrysene	31	17	29	116	53	85	146	70	114
Benzo[b]fluoranthene	31	21	36	138	58	98	169	79	134
Benzo[k]fluoranthene	18	11	22	99	39	65	117	50	88
Benzo[e]pyrene	21	17	21	90	49	67	111	66	88
Benzo[a]pyrene	18	18	17	52	24	36	70	42	53
Indeno[123cd]pyrene	27	13	20	141	38	78	168	51	98
Dibenz[ah]anthracene	6.7	5	7	27	7	16	34	12	22
Benzo[ghi]perylene	27	10	19	113	44	75	141	54	94
Total PCBs	17	9	13	20	NA	20	37	NA	37
Aluminum (x 10 ³)	160	114	137	1300	1093	1200	1460	1207	1340
Arsenic	580	633	607	1000	1094	1050	1580	1727	1660
Cadmium	920	813	867	130	347	240	1050	1160	1110
Chromium	1100	951	1026	2000	2064	2030	3100	3020	3060
Copper	6100	5049	5575	3500	3745	3620	9600	8790	9200
Iron	120000	145600	132800	750000	582000	666000	870000	727000	799000
Manganese	12700	13700	13200	15000	12200	13600	27700	25900	26800
Nickel	3000	11370	7185	5300	7019	6160	8300	18400	13300
Lead	6700	4180	5440	5200	8960	7080	11900	13100	12500
Selenium	1500	1280	1390	3000	2860	2930	4500	4140	4320
Zinc	18000	35300	26600	23000	22600	22800	41000	57800	49400

Table III.10. Relative importance of sources of trace metals and organic contaminants to the Chesapeake Bay.

	Susquehanna River Load (kg/year) ¹		Atmospheric Deposition
	Dissolved	Particulate	Load (kg/year) ²
Organics			
t-PCBs	97	68	37
Fluorene	37	85	27
Phenanthrene	63	388	155
Fluoranthene	108	1020	189
Pyrene	104	925	184
Benz[a]anthracene	12	364	44
Chrysene	15	316	114
Benzo[a]pyrene	5	436	53
Elements			
Aluminum (x 10 ⁻³)	2,560	64,800	1,340
Arsenic	12,600	ND	1,660
Cadmium	2,130	26,700	1,110
Chromium	4,130	111,000	3,060
Copper	47,800	151,000	9,200
Iron (x 10 ⁻³)	4,100	40,000	799
Manganese (x 10 ⁻³)	3,290	1,530	26.8
Nickel	121,000	65,200	13,300
Lead	6,530	38,600	12,500
Zinc	77,900	360,000	49,400

¹Annual loads entering the Chesapeake Bay via the Susquehanna River, measured at Conowingo, Maryland between February, 1994 and January 1995 by Foster (1995; organics) and Conko (1995, elements).

²Total atmospheric deposition loads directly to the surface of the Chesapeake Bay below the fall lines (1.15×10^{10} m²), as measured by the Chesapeake Bay Atmospheric Deposition Study, 1990-1991.

aromatic hydrocarbons which are comparable to (fluorene) or greater than the loads of *dissolved* PAHs delivered by the Susquehanna River (Table III.10). Dissolved total PCB loads from the river are approximately three times those from the atmosphere. Particulate-bound organic contaminants discharged from the river dominate the external loading of PAHs, with a large contribution from the high sediment burden carried by the river during high flows. Atmospheric depositional fluxes of several elements, including lead, cadmium, and chromium, are within a factor of two of the dissolved load from the Susquehanna River. Again, particulate metal loads from the river dominate over both dissolved riverine loads and atmospheric deposition.

While it is interesting to compare the relative importance of riverine and atmospheric sources of trace elements and organic contaminants to the Chesapeake Bay, one must take care not to overinterpret these results. While the Susquehanna River delivers large quantities of these species, much of this load is removed in the northern extreme of the bay (Helz and Huggett, 1987) and is delivered episodically during high river flows. Whether particulate-bound metals and organic contaminants are available to geochemically repartition within the bay's waters and to be taken up by aquatic organisms is quite unclear. In contrast, atmospheric deposition directly to the water's surface supplies these species directly to water column, without any comparable zone of efficient removal. However, it has recently been suggested that combustion-derived PAHs associated with aerosol particles washed into surface waters also may not be available for repartitioning (McGroddy *et al.*, 1995). Finally, one must remember that the distinction between riverine and atmospheric loadings is not clear. Some fraction of the chemical input from the tributaries results from deposition of atmospheric pollutants to the watershed, with subsequent transmission through the vegetation and soil layers into surface waters.



IV. Results of Intensive, Event-Based Study

Part IV.A. The Influence of Submicron Particles on Hydrophobic Organic Contaminants in Precipitation. 1. Concentrations and Distributions of Polycyclic Aromatic Hydrocarbons and Polychlorinated Biphenyls in Rainwater

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IVA.1 Abstract

Concentrations of non-filter retained, particle-associated organic contaminants in rain are estimated as the difference between measured filtrate concentrations and those supported by gas exchange into raindrops, using field data resulting from the simultaneous collection of air and precipitation. The distributions of organic contaminants in four discrete precipitation events are presented for several polychlorinated biphenyl congeners and for polycyclic aromatic hydrocarbons. These contaminants are up to 80% bound to non-filterable particulate material in rainwater and <9% "truly" dissolved, demonstrating that submicron particles significantly contribute to the overall removal of organic contaminants from the atmosphere by precipitation. In the precipitation events described in this paper, organic carbon-normalized filter-retained particle ($\sim > 0.5 \mu\text{m}$) partition coefficients for PCB congeners and PAHs are larger than those for non-filter-retained particles, suggesting that organic contaminants adsorb differently to large and small (*i.e.*, non-filter retained) particles. Partition coefficients to both large and submicron particles are weakly correlated with hydrophobicity, indicating that the speciation of PCBs and PAHs in rain is not predictable from these compounds' octanol-water partition coefficients. We attribute the poor correlation to slow desorption kinetics and to non-exchangeable fractions of contaminants within atmospheric particles.

IVA.2 Introduction

Hydrophobic organic contaminants (HOCs) exist in the atmosphere as gases and associated with liquid and solid particles (1-4). To predict the atmospheric residence times, deposition rates, and fates of HOCs requires that the mechanisms of HOCs incorporation into atmospheric liquid water and removal from the troposphere be understood. Both gaseous and particle-associated HOCs are deposited to the Earth's surface through dry depositional processes, (*e.g.* gas absorption, gravitational settling, eddy diffusion) or *via* precipitation scavenging, either during droplet formation (in-cloud scavenging) or as the drops fall through the air column (below-cloud scavenging) (5-7). Fogwater can also incorporate gaseous and particle-associated HOCs close to the ground or below clouds prior to rain or drizzle (8-11). Previous studies of the composition of rain- and fogwater have contributed to our understanding of the physical speciation, fate and transport of airborne HOCs (8-16). Murray and Andren (17) argue that the dissolution of gaseous polychlorinated biphenyls (PCBs) to rain is apparently larger than what is predicted by the congeners' air-water partition coefficients due to the presence of non-filter retained ($< 0.7 \mu\text{m}$), contaminant-laden particles in the operationally defined dissolved phase of their precipitation samples. Similarly, Schomburg *et al.*, (9) and others (10, 18-20) conclude from field studies that the enrichment of HOCs in fogwater is due to submicron particulate material in fogwater that is strongly sorptive, surface active, and not retained by standard glass fiber filters. Leuenberger *et al.*, (16) also suggest that the presence of insoluble *n*-alkanes in the filtered fraction ($< 0.2 \mu\text{m}$) of rainwater is likely due to clusters or aggregates in this phase.

The concept that submicron organic particulate matter influences the partitioning behavior of HOCs in rain- and fogwater is similar to that invoked to describe the sorption of HOCs to

particles in freshwater and estuarine environments (21-23). Evidence suggests that HOCs have a high affinity for substrates such as dissolved organic matter (24-28), including humic materials (29,30). These types of substrates may exist in rainwater and influence HOC partitioning, though the relative magnitude of HOC sorption to submicron material in rainwater has not been thoroughly examined. Since observed enhancements of HOCs in rain- and cloudwater (*i.e.*, beyond what is predicted by gas exchange) have been attributed to the presence of submicron particle associated HOCs in the operationally defined filtrate, HOC precipitation scavenging models which do not include submicron particle-associated concentrations may be in error and should be reexamined.

We recently carried out a detailed investigation of HOC speciation in rain in the Chesapeake Bay region. In this paper, we present concentrations (dissolved and particle-associated) of PCB congeners, polycyclic aromatic hydrocarbons (PAHs), organic carbon, and suspended particles in rain, along with HOC atmospheric concentrations (particle-associated and gaseous), all of which were measured during five discrete storm events that occurred adjacent to the Chesapeake Bay in late Summer, 1992. Estimates of submicron particle-associated HOC concentrations in rain are presented as the difference between measured filtrate concentrations and those supported by gaseous absorption. The resulting partitioning of HOCs between particles and rainwater is compared to similar fog studies. An accompanying paper (31) details an HOC scavenging model which includes the scavenging ratios of particle-associated, gaseous, and submicron PAHs.

IVA.3 Methods

Concentrations of 65 PCB congeners and 14 PAHs were measured in atmospheric and precipitation samples collected at a rural site adjacent to the Chesapeake Bay during late Summer 1992 (Table IV.A-1). This site is one of the Chesapeake Bay Atmospheric Deposition network sites where atmospheric (semi-weekly) and precipitation (integrated bi-weekly) samples were collected between July 1990 and September 1993 for HOC analysis (32). Sampling and analytical procedures are presented in detail elsewhere (32, 33) and are summarized here. Particle-associated and gaseous contaminants were isolated from the atmosphere by drawing air through a 20.3 x 25.4 glass fiber filter (Schleicher and Schuell #25) and a polyurethane foam plug (PUF; 8.5 cm x 10 cm) for about 16 hours at a flow rate of 0.8 m³/min. Samples were collected before, during (when logistically possible), or immediately after each rain event. An additional high-volume air sampler was deployed for measuring total suspended particle (TSP) concentrations. Precipitation was collected for HOC analysis in an automated wet-only collector equipped with a 1 m² stainless steel collection funnel that drained precipitation through an *in situ* filtration system consisting of a glass fiber filter (mean pore size 2.9 μ m, 144 mm, Schleicher and Schuell #25) and a column of Amberlite XAD-2 resin (32). To include small particles which adhered to the funnel surface rather than being washed onto the filter, the funnel was scrubbed with wetted glass wool and deionized water after each rain event. This funnel wash was combined with the corresponding precipitation filter sample. An additional automated wet-only precipitation collector (Aerochem Metrics, Bushnell, FL) was deployed to collect precipitation for the analysis of

Table IV.A-1. Storm Events, 1992, Chesapeake Bay, MD

Event	Date	Wind Direction ¹	Air Temp ²	TSP ³	DOC ⁴ mg/L	POC ⁴ mg/L	foc ⁴ %	Rain Vol mm	Duration hrs	Trigger Event ⁵
rain storm #1	31 July	180	27	3.4 mg/L	3.7	0.9	20	13	< 1	pre-frontal trough ⁷
air smp a	28 July	320	27	29 µg/m ³	n/a ⁶	n/a	n/a	n/a	n/a	
rain storm #2	11 August	200	33	n/a	n/a	n/a	n/a	34	no data	pre-frontal trough
air smp b	11 August	340	33	36 µg/m ³	n/a	n/a	n/a	n/a	n/a	
rain storm #3	24 September	010	17	0.9 mg/L	2.1	0.2	9	31	16.8	tropical storm Danielle
air smp c	22 September	360	31	21 µg/m ³	n/a	n/a	n/a	n/a	n/a	
rain storm #4 ⁸	27 September	210	22	2.8 mg/L	1.2	0.2	13	13	17.5	pre-frontal trough
rain storm #5	4 October	045	16	1.4 mg/L	9.7	0.6	6	15	12	coastal low pressure system
air smp d	6 October	040	16	14 µg/m ³	n/a	n/a	n/a	n/a	n/a	

¹ at 18Z

² mean air temperature (°C) during rain storm or during air sample

³ total suspended solids concentrations in rain (mg/L) or air (µg/m³)

⁴ DOC = dissolved organic carbon; POC = particulate organic carbon; foc = fraction of organic particulate carbon in rain

⁵ meteorological disturbance that brought about the precipitation event (ref 36)

⁶ not available

⁷ pre-frontal atmospheric low pressure depression

⁸ 27 September rain sample coupled to 22 September air sample

suspended solids and dissolved and particulate carbon concentrations.

Air and precipitation filter samples were extracted for 24 hours with methanol followed by 24 hours with dichloromethane in Soxhlet flasks. Air vapor samples (PUF) were extracted for 24 hours with petroleum ether and resin (XAD-2) samples were extracted for 24 hours with 1:1 (v/v) acetone:hexane. Laboratory blanks of each matrix were also extracted to account for contamination during processing. Extraction was followed by liquid-liquid partitioning into deionized water to remove the more polar solvent in the matrices that had solvent combinations. The remaining non-polar extracts were then concentrated to < 4 mL *via* rotary-evaporation and were further evaporated under a nitrogen stream. Dichloromethane extracts were exchanged for hexane during the rotary-evaporation process. A 25% subsample of each concentrated extract was taken for PAH analysis, concentrated, and analyzed by capillary gas chromatography-ion selective mass spectrometry. Individual PAHs were identified based on chromatographic retention times, confirmed by the abundance of a secondary mass fragment relative to the molecular ion, and quantified by isotope dilution using five perdeuterated PAHs internal standards (MSD Isotopes, Cambridge, MA; Supelco, Bellefonte, PA). The remaining sample extract was concentrated, fractionated on Florisil columns into two eluents (23), and analyzed by capillary gas chromatography with electron-capture (ECD) detection for PCBs. PCB congeners were identified and quantified by comparison to a mixed Aroclor standard (1232, 1248, 1262) by the procedure of Mullin (34), using two non-industrially synthesized PCB internal standards (35). In cases where two or more PCB congeners were not chromatographically resolved, the total concentration of the mixture is reported.

Recoveries of an analytical surrogate (PCB congener #14, 3,5 dichlorobiphenyl) spiked into each sample prior to extraction were 65-108% (N = 9), exclusive of two outliers at 15 and 40%. Recoveries of d₁₂ chrysene and d₁₂ fluoranthene were 73-108% (N = 13) with one outlier at 162%, and 43-70% (N = 5), respectively. Unusual surrogate recoveries likely result from chromatographic interferences with the specific surrogate compounds in individual samples. Laboratory blanks were generally free of contamination (*i.e.* below instrumental detection limits). In the limited cases where masses of sample analytes were less than three times those in the corresponding matrix blank, the field analytes were not included in the data set. Instrumental detection limits were calculated for each sample by assuming a minimum chromatographic area with a signal to noise ratio equal to three.

Total organic carbon was determined with an OF Corporation (College Station, TX) TO analyzer (Model 700) *via* persulfate wet-oxidation at 100°C and non-dispersive infrared detection. Particulate carbon was measured *via* combustion in pure oxygen under static conditions at 950 °C in a Leeman Laboratory, Inc. (#240-XA) elemental analyzer.

IVA.4 Results and Discussion

Storm descriptions. The duration of each storm sampled during July-October 1992 ranged from less than one hour to 17.5 hours, indicating that the storms were generally

convective, quick moving fronts rather than stationary systems (Table IV.A-1). Atmospheric pressure depressions in front of occluded frontal systems, commonly called troughs, moving toward the Atlantic Ocean triggered rapid precipitation events over the Chesapeake Bay region on 31 July, 11 August, and 27 September as the frontal systems passed over the Appalachian Mountains and the leeward coastal flood plain (36). In contrast, tropical storm Danielle, located off the mouth of the Chesapeake Bay on 24 September 1992, was responsible for the precipitation that fell on this date. An occluded low pressure system located in the Carolinas brought about precipitation to the Chesapeake Bay region on 4 October. Rain intensities ranged from $0.8 \text{ mm}\cdot\text{hr}^{-1}$ on 27 September to as high as $13 \text{ mm}\cdot\text{hr}^{-1}$ on 31 July.

Ancillary data. Suspended solids concentrations in rainwater ranged from 0.89 to $3.4 \text{ mg}\cdot\text{L}^{-1}$ (Table IV.A-1). The carbon content of the particles filtered from the precipitation ranged from 6 to 20% and was highest in the event with the highest suspended solids concentrations (31 July). Dissolved carbon ranges from 1.2 to 9.7 mg/L . Murray and Andren (17) report DOC ranging from 0.5 to $4.9 \text{ mg}\cdot\text{L}^{-1}$ in 10 storms in Madison, Wisconsin. Glotfelty *et al.* (19) report total organic carbon in fogwater ranging from 38 to $55 \text{ mg}\cdot\text{L}^{-1}$. High concentrations of organic carbon in rain and fog are thought to cause an enhancement in the equilibrium gas scavenging of HOCs due to the strong partitioning of contaminants to organic particles and DOC (10).

Atmospheric concentrations of HOCs. Average gaseous concentrations of PCB congeners measured during this study are shown in Figure IV.A-1. Air samples were taken as close as possible to or during each rain event. PCB congeners 19, 40, 119 and 193 (IUPAC numbers) were not detected above the analytical blank in any air sample. Total PCB concentrations (sum of 61 congeners) range from 304 to $542 \text{ pg}\cdot\text{m}^{-3}$ and are similar in magnitude to the mean total PCB concentration measured at the same site during 1990-1991 ($210 \text{ pg}\cdot\text{m}^{-3}$; 32). Particle-associated PCBs were not present in air samples above our operational detection limits ($\Sigma\text{PCB} \sim 2.1 \text{ pg}\cdot\text{m}^{-3}$) but are estimated to be $<3\%$ of the total atmospheric burden in the Chesapeake Bay region based on congener-specific vapor pressures, ambient temperatures, and the ambient TSP levels in the region (3,32).

Phenanthrene, pyrene, and fluorene, respectively, are the dominant PAH species in the gas phase and higher molecular weight compounds, such as indeno[123-*cd*]pyrene and benzo[*ghi*]perylene are predominantly particle-associated (Figure IV.A-1). The highest particle-associated PAH concentrations were observed on October 6th, while the highest gas phase PAH concentrations were observed two weeks earlier. Variations in organic contaminant gas-particle distributions result from changes in temperature and particle composition over relatively short temporal scales and are not unexpected. Similar variations in PAH speciation in the atmosphere were observed during 1990-1991 at the same site (32).

Concentrations of HOCs in rain events. In this study, total PCB concentrations (dissolved + particulate) ranged from 0.85 to $2.2 \text{ ng}\cdot\text{L}^{-1}$ and are similar to the volume weighted mean concentration of ΣPCBs measured during 1990-1991 at the same site ($1.6 \text{ ng}\cdot\text{L}^{-1}$, 31; Figure IV.A-2). Murray and Andren (17) report a total PCB volume weighted mean concentration

measured in 10 storms in Madison, Wisconsin of $3.5 \text{ ng}\cdot\text{L}^{-1}$, with the particulate-bound PCBs contributing up to 50% of the total. The particle-associated (*e.g.* filter-retained) PCBs in this study contribute less to the total, ranging from 7% to as high as 50% on one occasion. Total (dissolved + particulate concentrations of the individual PAHs) PAH concentrations in precipitation range from 0.5 to $6 \text{ ng}\cdot\text{L}^{-1}$ (Figure IV.A-3). PAH concentrations in the first storm event are nearly twice as high as those in the other four storms. Large temporal variability in HOC concentrations in precipitation were previously observed at this site, with relative standard deviations observed in semi-weekly integrated PAH concentrations measured at the same site from 1990 - 1991 greater than 100% (32). PAHs are predominantly associated with filter-retained particles in precipitation with the exception of fluorene and phenanthrene (Figure IV.A-3). Variable patterns of HOC speciation in rainfall imply that the mechanisms of wet deposition of HOCs (*e.g.*, wet scavenging) are highly variable. Differences in contaminant speciation in rain may result from differences in scavenging mechanisms among storms.

Interestingly, particle-associated PAH concentrations on a weight basis of particles ($\mu\text{g}/\text{g}$) measured in rain collected on 24 September are considerably higher than those observed in the atmosphere prior to the storm, perhaps due to dissolution of non-PAH enriched particles in the rain drops (*e.g.* soluble sulfate aerosol particles providing a higher TSP concentration in dry air relative to rain; Figure IV.A-4). The enhancement of PAHs on rain particles may also be due to selective removal of particles which are enriched in PAHs. The removal of particles from the atmosphere by precipitation has been shown to be a function of particle size (38,39). Slinn *et al.* (38) argue that particles much less than $0.1 \mu\text{m}$ are efficiently scavenged by raindrops. Accordingly, several workers suggest that PAHs are generally associated with atmospheric particles less than $0.1 \mu\text{m}$ (40-43). Hence, it is not unlikely that enhanced particle-associated PAH concentrations in rain (on a weight basis) relative to air are the result of selective scavenging of submicron particles which are enriched with PAHs. PAH concentrations observed in the next rain event are considerably lower than those observed in the earlier event (Figure IV.A-4), suggesting that the particles were effectively removed from the atmosphere *via* rain scavenging during the first storm.

Contaminant distributions in rain. Within raindrops, HOCs are either dissolved or associated with filter- and non-filter retained particles. Dissolved HOC concentrations in rain supported by equilibrium partitioning with the surrounding gas phase can be estimated from measured contaminant concentrations in the gas phase and temperature-corrected air-water partition coefficients (Henry's Law Constants, H ; 12), assuming the droplet reaches equilibrium with the surrounding air. Henry's Law describes the equilibrium distribution of a compound between the gas and liquid phases for dilute solutions as:

$$[(R \times T) / H] = C_d / C_g \quad [1]$$

where R is the ideal gas constant ($8.21 \times 10^{-5} \text{ m}^3\cdot\text{atm}\cdot\text{mol}^{-1}\cdot\text{K}^{-1}$), T is temperature (K), H is the compound specific Henry's Law constant (HLC, $\text{m}^3\cdot\text{atm}\cdot\text{mol}^{-1}$), C_d is the dissolved HOC concentration ($\text{ng}\cdot\text{m}^{-3}$) and C_g is the gaseous HOC concentration ($\text{ng}\cdot\text{m}^{-3}$). In order to accurately

predict gas/dissolved HOC distributions, it is necessary to know Henry's Law constants (H) at the ambient temperature. Generally, H values reported in the literature result from laboratory measurements at 25°C or are calculated from measurements taken at different temperatures. Tateya *et al.* (44) have demonstrated that the dependence of H on temperature for PCB congeners is given by $\ln H = 18.58 - 7859 \cdot T^{-1}$. Similar relationships have been reported for other HOC compounds (4,45) including mirex (46), hexachlorocyclohexane (47), and methoxylated phenols (48). For PAHs, a similar dependence of H on temperature as that for PCB congeners is generally assumed (37). In this study, H values for PCBs at ambient air temperatures were estimated at each air temperature measured during atmospheric sampling from the values at 25°C using the slope of Tateya *et al.* (44). Henry's Law constants for PAHs calculated at 25°C using published solubility and vapor pressure data (49-51) were corrected to ambient sample air temperatures using the slope of Tateya *et al.* (44) and compound specific intercepts (37).

In this study, aqueous PCB and PAH concentrations in rain, predicted from ambient gaseous concentrations and Henry's Law, are less than measured filtrate concentrations by an average factor of about 100, indicating that the rain is apparently supersaturated with respect to the ambient air concentrations (Figure IV.A-5). This apparent enrichment of HOCs in rainwater is similar to that reported for contaminants in rain and fogwater (8,10,17). Higher than predicted dissolved PCB levels in precipitation were observed by Murray and Andren (17). Capel *et al.* (10) report that less than 1% of PCBs, *n*-alkanes, PAHs, phenol, and many chlorinated dioxins and furan congeners are truly dissolved in the aqueous phase of fogwater. Pesticides in coastal fogwater collected near Monterey, California are also enriched relative to the gas phase (8).

Calculating dissolved HOC levels supported by gas exchange assumes that the raindrop reaches equilibrium with the surrounding air during its fall, and that our ground-level air measurements adequately characterize the gaseous phase. Scott (7) argued that gas exchange across the droplet surface is rapid and the time required to reach equilibrium is < 1 second. There is some uncertainty in this estimate, largely resulting from the assumption that the droplet is well mixed and from errors in the air-water mass transfer coefficient. Using macroscopic-scale transfer coefficients, we estimate times to equilibrium of less than ten minutes, which is likely an upper bound to Scott's estimate. Schwartz and Frieberg (52) and Warneck (53) also conclude based on theoretical calculations that times to equilibrium are short. It is also conceivable that enrichment results from the raindrops passing through an air layer with higher gaseous HOC concentrations than those which we measured at ground level. A 10-100 fold increase in HOC concentrations at height relative to ground level would be required to quantitatively explain our observed enrichments. Knap and Binkley (54) measured chlorinated aromatic compounds at differing altitudes over the Western Atlantic, and observed no clear trends in concentration with altitude. Increases in gaseous HOC levels with altitude are unlikely for two reasons. First, the sources of HOCs are at ground level. Second, decreasing temperatures with altitude result in gaseous HOCs adsorbing onto aerosol particles. Finally, enrichment of HOCs in precipitation may result from the coincident scavenging of surface-active organic matter which creates organic surface films on raindrops. However, Gill *et al.* (55), using data from Likens *et al.* (56), demonstrated that there is insufficient surface-active organic material in raindrops to form a

cohesive film on drops less than 4 mm. They further conclude that for cloud and fog drops up to 1 μm , only about a 1% monolayer surface coverage will occur. Similarly, Ligocki *et al.* (12) argue that surface films do not control the rates of HOC gaseous sorption to rain drops. Hence, the observed enrichment of contaminants in the filtrate is probably not due to non-equilibrium, nor to vertical gradients in gaseous HOCs, nor to the presence of organic surface films on liquid drops, but rather results from the scavenging of HOC-enriched submicron particles.

This conclusion is consistent with that reported by Duinker and Bouchertall (15), who suggest that precipitation scavenging of particle-associated PCBs is the dominant mechanism for PCB removal from the atmosphere *via* rainfall despite a low ambient aerosol fraction in the atmosphere. Further, they conclude from Doskey and Andren's data (57) that submicron particles may be the main absorptive phase for particle-associated PCBs in the atmosphere. Similarly, Gonzales and Murr (58) determined based on electron micrographs of individual raindrops that the average particle diameter scavenged by a raindrop is $\sim 0.2 \mu\text{m}$, suggesting that submicron scavenging of HOCs contributes substantially to the overall wet removal of HOCs from the atmosphere rather than the presence of surface-active organic films on raindrops.

Contaminant partitioning to filter- and non-filter retained particles in rain. The speciation of HOCs in rainwater may be modeled by assuming that the dissolved and particulate phases achieve a sorptive equilibrium which can be predicted from the HOC's octanol-water partition coefficient (K_{ow}) and the carbon content of the particles (21-30,59,60). To evaluate whether this approach explains the observed PCB and PAH distributions in our event precipitation samples, we calculated partition coefficients to filter-retained and non-filter-retained particles as follows:

$$K_{oc,nf} = [C_{nf} * 10^6 / C_{d,p} * [\text{DOC}]] \quad [2]$$

where $[C_{nf}]$ is the estimated concentration ($\text{ng}\cdot\text{L}^{-1}$) of HOC associated with non-filter retained particles, $[C_{d,p}]$ is the predicted equilibrium dissolved HOC concentration in rain ($\text{ng}\cdot\text{L}^{-1}$) and $[\text{DOC}]$ is the concentration of dissolved organic carbon in rain ($\text{mg}\cdot\text{L}^{-1}$; 9), and:

$$K_{oc,f} = [C_{p,f} * 10^6 / C_d * \text{TSP}_r] / [\text{foc}] \quad [3]$$

where $C_{p,f}$ is the measured filter retained particle-associated HOC concentration in rain ($\text{ng}\cdot\text{L}^{-1}$), $C_{d,p}$ is the predicted equilibrium dissolved HOC concentration in rain ($\text{ng}\cdot\text{L}^{-1}$), TSP is the suspended solids concentration in rain ($\text{mg}\cdot\text{L}^{-1}$), and foc is the fraction of particulate organic carbon. Measured dissolved organic carbon concentrations were used as a surrogate measure of the quantity of non-filter-retained particles in the rainwater, as has been previously done for surface waters (23).

Partition coefficients for PCBs and PAHs in rainwater are generally much higher than those measured in surface waters (61), suggesting that atmospheric particles are more sorptive than solids in surface waters. Partition coefficients of PCB congeners to filter-retained particles

are approximately ten times larger than those to particles which pass through the glass fiber filter (Figure IV.A-6). In contrast, PAH partition coefficients to filter-retained particles, while very large, are approximately ten times lower than the partition coefficients to the non-filter-retained particles. This difference between the partitioning behavior of PCBs and PAHs may reflect their different atmospheric behaviors. PCBs are released as gases into the atmosphere primarily via volatilization from the Earth's surface, and subsequently adsorb onto aerosol particles, most likely through a reversible surface adsorption process. In contrast, PAHs emitted during combustion processes may be incorporated into the matrix of primary aerosols. A fraction of PAHs associated with particulate matter in the atmosphere, therefore, is likely bound within the particle matrix and not exchangeable with the surrounding gas phase (62,63). The extremely high PAH partition coefficients measured in rainwater are consistent with the idea that PAHs are tightly bound by atmospheric particles, and do not achieve sorptive equilibrium with the surrounding liquid within a raindrop.

If HOCs reached sorptive equilibrium within rainwater, their partition coefficients to both large (*i.e.* filter-retained) and small (*i.e.* non-filter-retained) particles should depend directly upon HOC hydrophobicity (as parameterized by their octanol-water partition coefficients). As shown in Figure IV.A-6, estimated organic carbon-normalized partition coefficients of PCBs and PAHs in rainwater are poorly correlated with K_{ow} . It is interesting to note that in previous studies Baker *et al.* (23) and others (21,22) have argued that the lack of dependence of K_d on K_{ow} observed in surface waters resulted from incomplete separation of particles and dissolved HOCs. However, in this study we have eliminated that artifact by calculating dissolved HOC concentrations from the surrounding gas phase concentrations, yet still observe partitioning behavior which is independent of HOC hydrophobicity. Based on this analysis, we conclude that these compounds are enriched in rainwater due to scavenging of contaminant-loaded submicron particles rather than the coincident scavenging of surface active organic matter which subsequently binds dissolved contaminants. In addition, we conclude that classic sorption models derived from studies of HOC associations with sediments and soils (*i.e.*, 59,60) cannot be used to predict HOC speciation in precipitation.

Murray and Andren (17) report similar distributions of PCB congeners in precipitation events, although we observed a larger fraction of submicron particle-associated PCBs in our study. Several of Murray and Andren's precipitation samples were collected during the winter at 0°C when large (> 0.5 μm), filter-retained particles are likely present in rural Wisconsin air due to home heating and wood burning stoves. In order to compare our results to those of Murray and Andren's (17), we calculated observed partition coefficients as the ratio of HOC concentrations on the filterable particles to those in the filtrate:

$$K_d = C_{p,f} * 10^6 / (C_{d,r} * TSP_r) \quad [4]$$

where $C_{p,f}$ is the measured filter retained particle-associated contaminant concentration in rain ($ng \cdot L^{-1}$), TSP_r is the suspended solids concentration ($mg \cdot L^{-1}$), and $C_{d,r}$ is the observed field measured "dissolved" (*e.g.* dissolved + non-filter retained particle-associated contaminants)

contaminant concentration ($\text{ng}\cdot\text{L}^{-1}$). Our calculated distribution coefficients for PCB congeners in precipitation are very similar to those found by Murray and Andren (17) in ten precipitation events in Madison, Wisconsin. In both cases, weak relationships between K_d and K_{ow} are reported. The similarity between these two studies in both the magnitude of K_d and in the lack of dependence on K_{ow} suggests that both the contaminant speciation in precipitation and the type of atmospheric particulate matter in rainfall in both studies was similar.

To demonstrate the quantitative importance of small particles in the overall removal of HOCs from the atmosphere, the distribution of several HOCs in rain on 11 August 1992 are presented in Table IV.A-2. The gaseous HOC concentrations were measured during the rain event, thereby giving our best estimate of atmospheric HOC speciation. Most of the compounds, all with varying physicochemical properties, are $\leq 9\%$ truly dissolved and the remaining fractions are associated with particles. The least volatile PAH (benzo[ghi]perylene) is not predicted to occur in the dissolved phase, as this compound is generally not detected in the ambient gas phase. Rather, the compound is mainly particle-associated due to its relatively low saturation vapor pressure and solubility. Interestingly, particle removal processes contribute a significant portion of PCBs to the total concentration in rain, even though the particle concentrations of these compounds in air are small relative to the total air burden ($< 30\%$).

IVA.5 Summary

Non-filter-retained and filter-retained particulate matter in rainwater play an important role in the overall removal of contaminants from the atmosphere. Estimates of non-filter-retained, particle-associated contaminants in rain from Henry's Law suggest that this fraction comprises up to 80% of the total HOC burden in precipitation, depending on the compound, and that less than 9% of the total HOC concentration in precipitation is truly dissolved. Hence, small particles can significantly contribute to the overall removal of contaminants from the atmosphere *via* rain.

Using dissolved HOC concentrations predicted by gas exchange, we calculated filter retained and non-filter retained particle partition coefficients, utilizing dissolved organic carbon as a surrogate for the non-filter retained suspended solids concentrations. Organic carbon-normalized filter retained particle partition coefficients are larger than those for non-filter retained particles for PCB congeners, but the opposite was observed for PAHs, suggesting that the compounds do not sorb to filter retained and non-filter retained particles in the same way and that particles retained by a glass fiber filter are chemically different than those submicron particles that are able to pass through the filter.

Organic carbon-normalized partition coefficients of PCBs and PAHs in rainwater are poorly correlated with K_{ow} , suggesting that these compounds do not necessarily partition to surface active organic matter within raindrops. Rather, HOCs are mainly removed by the scavenging of contaminant-enriched particles of various sizes. In addition, the lack of dependence of the observed PAH and PCB partition coefficient suggest that HOC partitioning in rainwater does not reach sorptive equilibrium and that extrapolation of sorption models based on HOC binding with

Table IV.A-2. Distribution of Organic Contaminants in Rain, Chesapeake Bay, MD, 11 August 1992

Compound	Total Rain		percent of total in rain		Total Air		Particle-Associated
	Concentration pg/L	Truly Dissolved	Non-filter Retained Particulate	Filter-retained Particulate	Concentration pg/m ³	Gaseous	
16/32 ¹	40	1	71	28	14	99	1
32/28	22	0.7	27	72	13	98	2
33/21/53	70	0.6	95	4	5	97	3
41/84/71	16	0.6	58	41	7	96	4
66/95	36	0.3	45	54	9	92	8
70/76	29	0.5	77	22	13	92	8
118	17	0.6	73	27	5	77	23
146	20	3	64	33	13	73	27
153/132/105	56	1	49	50	11	72	28
174	15	3	35	62	3	80	20
187/182	18	0.4	25	75	4	70	30
phenanthrene	3600	5	59	36	1400	99	1
pyrene	900	8	17	75	400	98	2
chrysene	1900	0.1	22	78	40	69	31
benzo[b]fluoranthene	1400	0.09	25	75	15	24	76
benzo[k]fluoranthene	710	0.03	23	77	5	19	81
indeno[1,2,3-cd]pyrene	750	0.03	32	68	6	3	97
benzo[ghi]perylene	1000	<0.1	31	69	8	<2	98

¹ biphenyl compounds; IUPAC congener groups

² air sampled during rain storm; particle-associated concentrations of PCBs estimated from relationships of the gaseous concentration to subcooled liquid vapor pressures (ref 3) at 30°C

sediments and soils cannot accurately predict HOC speciation in precipitation.

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IVA.7 Figure Captions

- IV.A-1. Mean concentrations (N = 4 samples) of PCB congeners (A) and PAHs (B) in air collected during five storm events that occurred adjacent to Chesapeake Bay, late Summer, 1992. The total PCB concentration is given for the 61 PCB congeners shown in A. The standard deviation is given in parentheses. Note the scale difference between the lighter molecular weight PAHs and those that are less volatile (B).
- IV.A-2. Concentrations and speciation of PCB congeners measured in five storm events that occurred adjacent to Chesapeake Bay in 1992.
- IV.A-3. Concentrations and speciation of PAHs measured in five storm events. * = the compound was not quantifiable above the analytical blank.
- IV.A-4. Particle-associated PAH concentrations ($\mu\text{g/g}$) on a weight basis of particulate material in the atmosphere (A) and in two sequential rain events (B and C).
- IV.A-5. Observed enrichment of HOCs in rainwater relative to that predicted by gas exchange.

IV.A-6. Filter retained (*A*) and non-filter retained (*B*) particle partition coefficients of PAHs and PCBs in rain, as calculated by equations [2] and [3], which both take into account the presence of non-filter retained particles in the measured dissolved phase. (Log K_{ow} values from references 64,65).

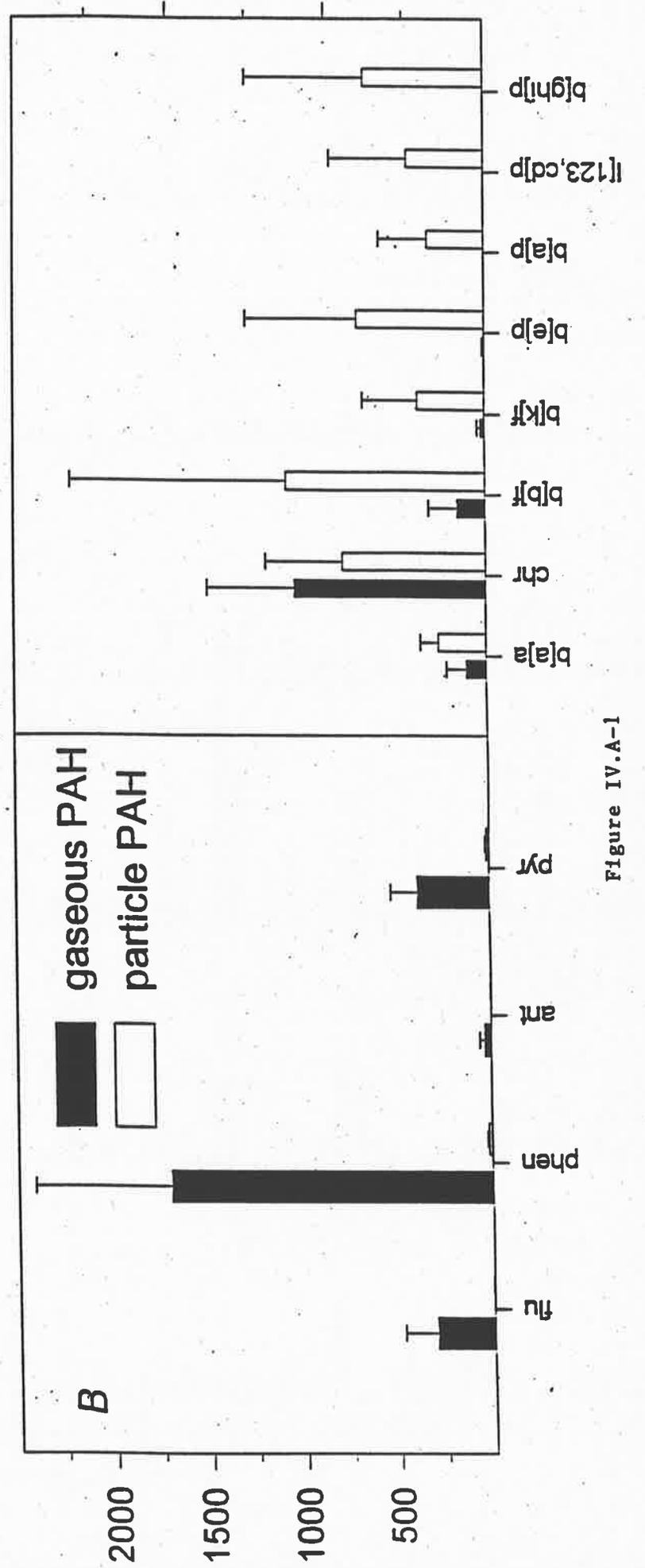
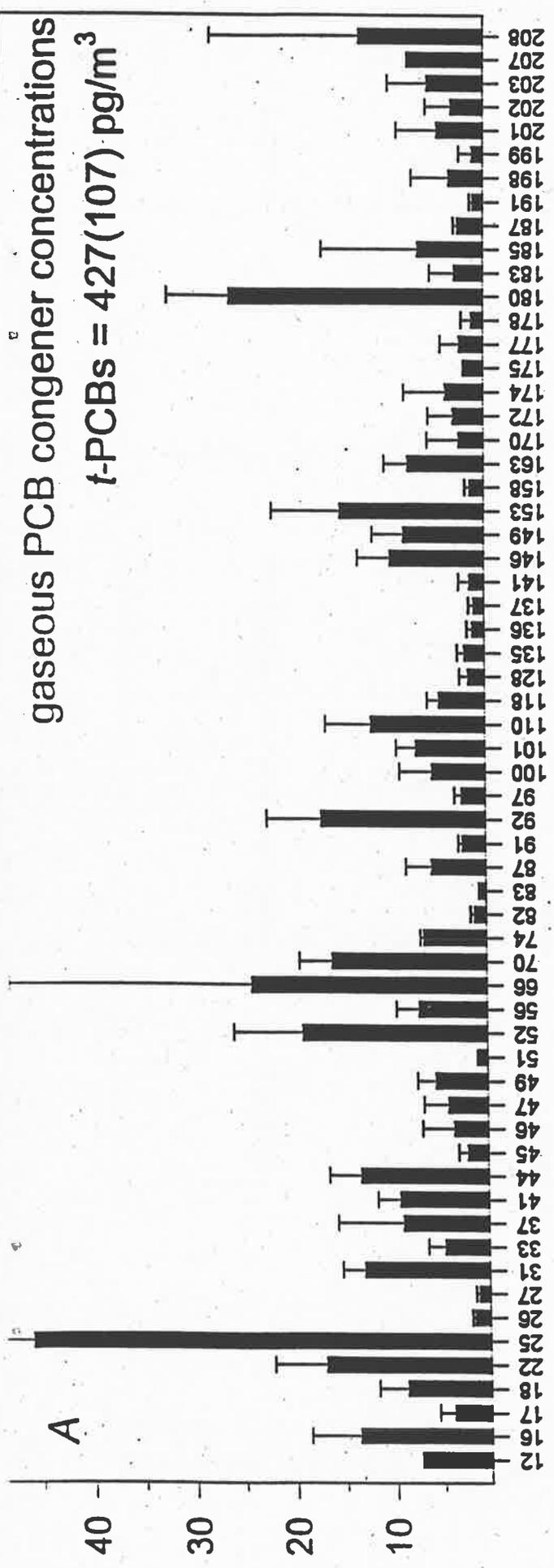


Figure IV.A-1

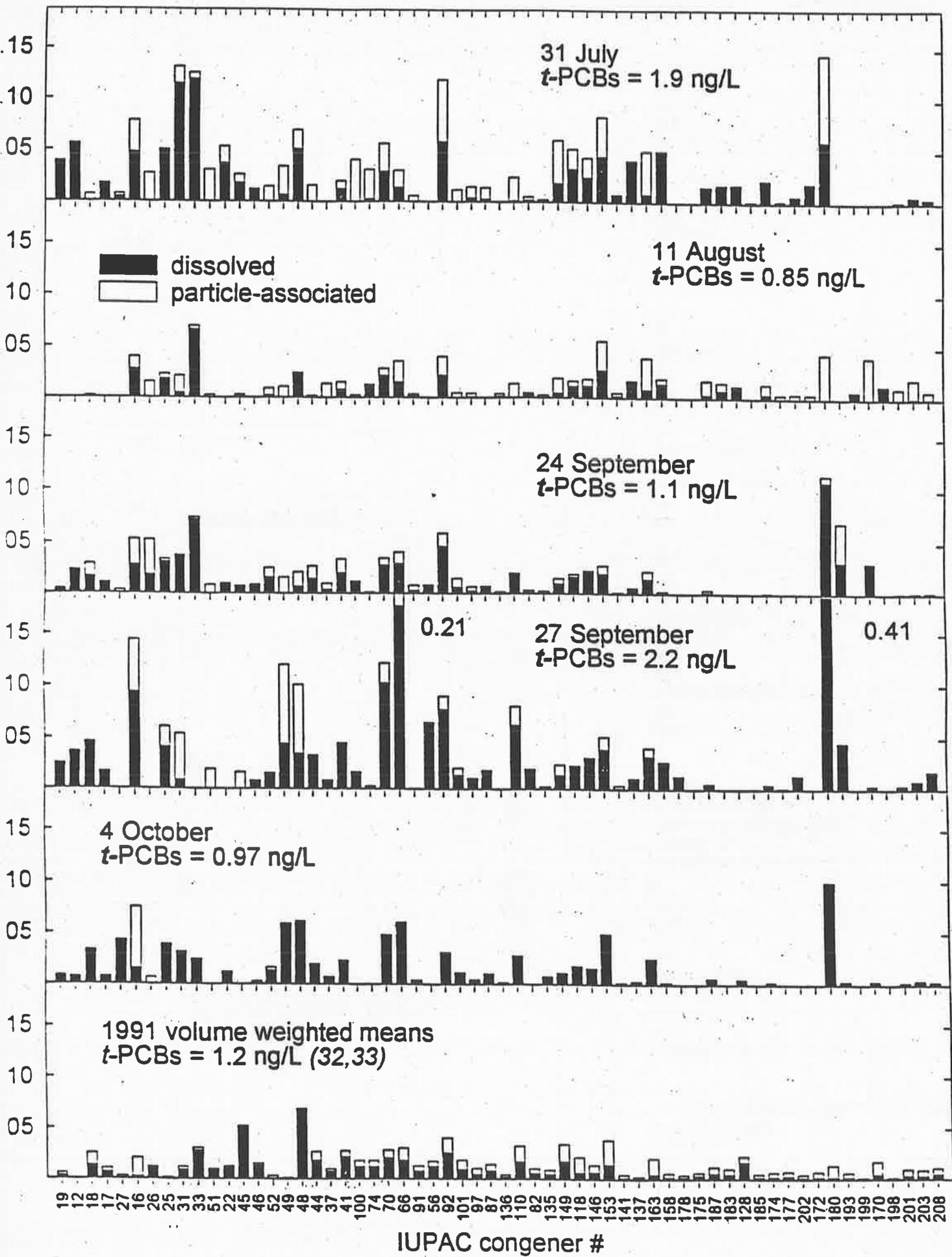


Figure IV.A-2

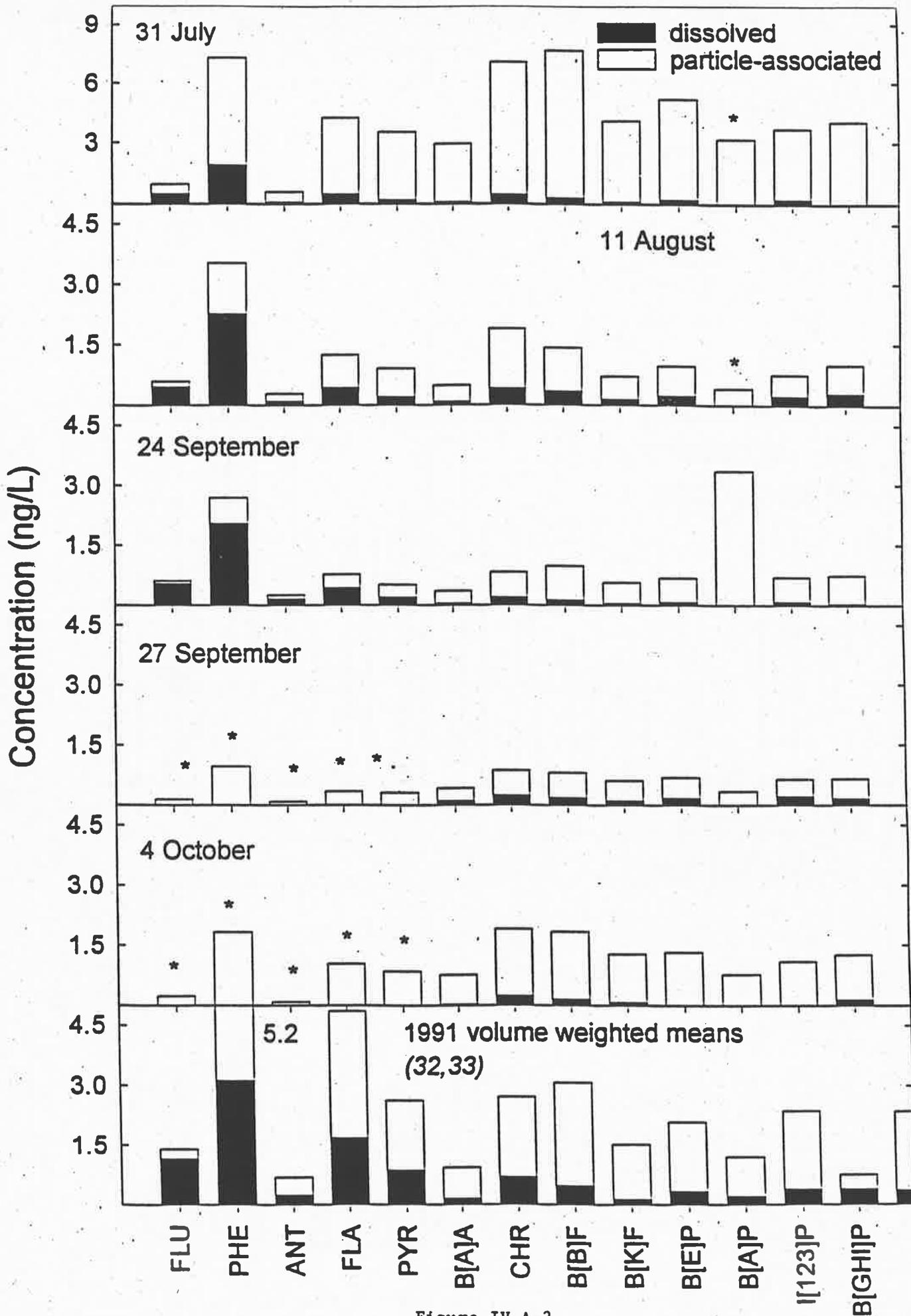


Figure IV.A-3

Concentration $\mu\text{g}/\text{g}$

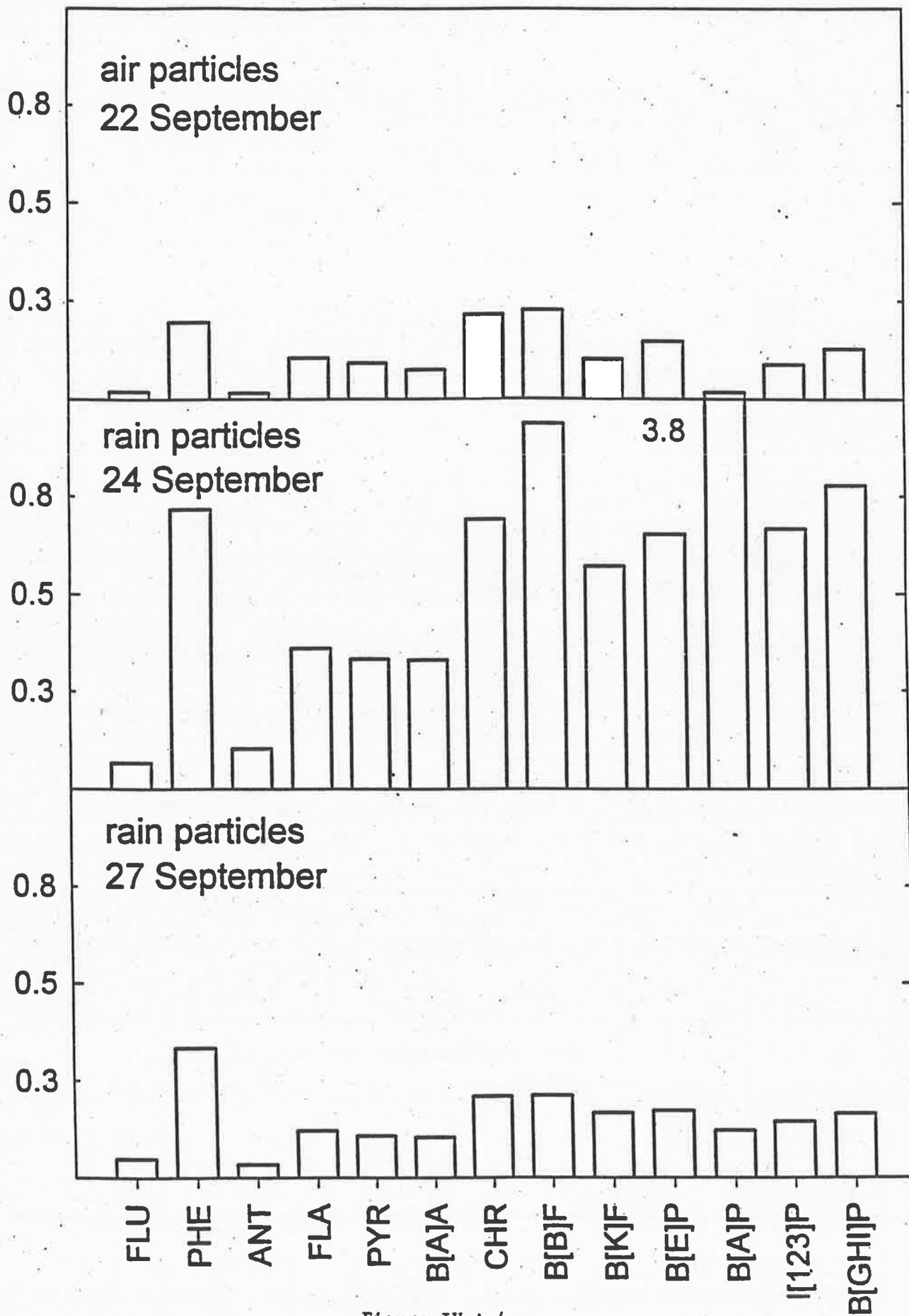


Figure IV.A-4

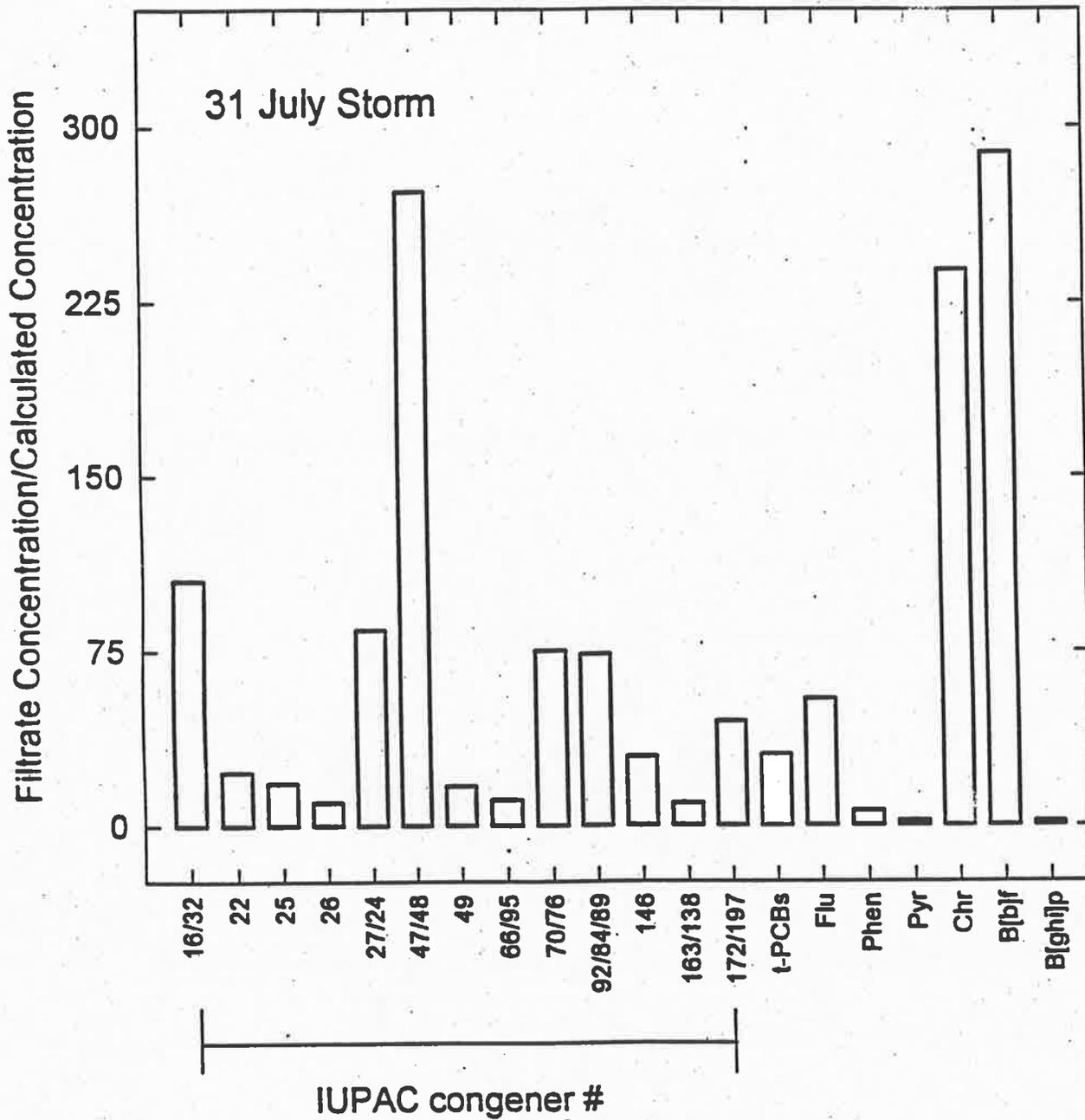


Figure IV.A-5

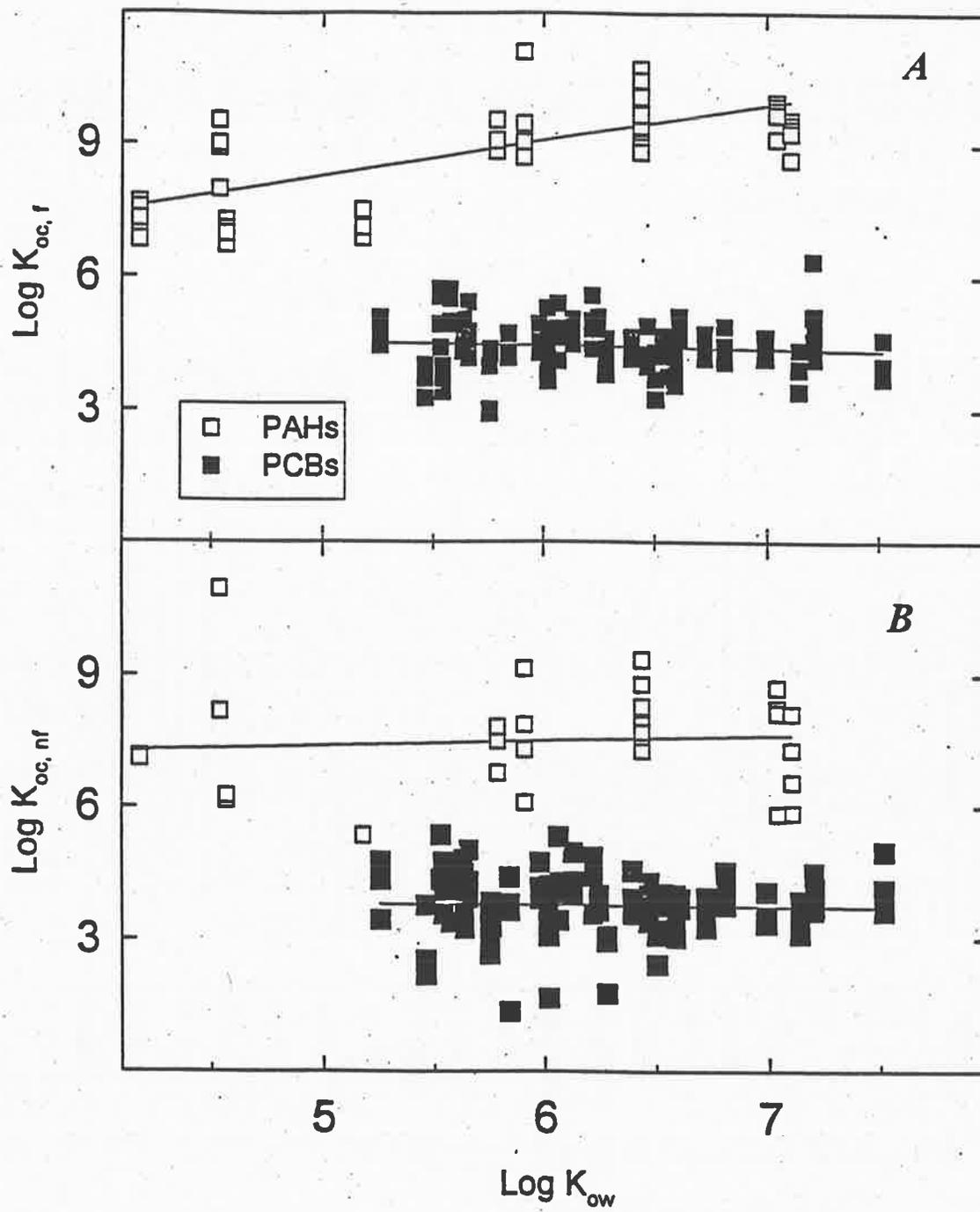


Figure IV.A-6



Part IV.B. The Influence of Submicron Particles on Hydrophobic Organic Contaminants in Precipitation. 2. Scavenging of Polycyclic Aromatic Hydrocarbons by Rain

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IVB.1 Abstract

In the second of two papers documenting the importance of submicron particles in controlling the concentrations and distributions of organic contaminants in precipitation, we examine the precipitation scavenging of polycyclic aromatic hydrocarbons (PAHs) as a function of the size distribution of atmospheric particles. We argue that particle scavenging rather than gas scavenging is the dominant removal mechanism for PAHs from the atmosphere by precipitation, and that previously reported PAH gas scavenging ratios were overestimated due to the presence of submicron particles in the operationally-defined dissolved phase of precipitation samples. In a series of five storms in the Chesapeake Bay region during the Summer, 1992, dimensionless PAH scavenging ratios for submicron particles range from 10^4 - 10^5 while those for larger particles ($>0.5 \mu\text{m}$) range from 10^5 - 10^6 . In contrast, gas scavenging ratios, as predicted from temperature-corrected Henry's Law constants, range from 10 to 500. The particle scavenging ratios of PAHs associated with large and small particles are similar to those measured for a variety of inorganic species and with those predicted for in-cloud scavenging of trace atmospheric species. Relative standard deviations of PAH particle scavenging ratios measured during the five storms are nearly 100%, suggesting that precipitation scavenging mechanisms are highly variable among the storms. This variability should be considered when using gas and particle scavenging ratios to model wet depositional fluxes of organic contaminants from the atmosphere.

IVB.2 Introduction

Polycyclic aromatic hydrocarbons (PAHs), often resulting from the incomplete combustion of fossil fuels and biomass, are emitted into the atmosphere either as vapors or associated with primary aerosol particles. Once in the atmosphere, the residence times and ultimate fates of these semi-volatile chemicals depend upon their distributions among vapor, particle, and droplet phases. This partitioning is in turn controlled by the vapor pressures, Henry's Law constants, and aqueous solubilities of the compounds, and by the concentrations and size distributions of particles and droplets in the atmosphere (1-4). Chemical transformations and wet and dry deposition remove gaseous and particle-associated PAHs from the atmosphere (5). Models used to estimate wet deposition of organic compounds generally distinguish between the scavenging of gaseous and particle-associated compounds (6,7). Atmospheric gases dissolve into droplets within clouds and into falling raindrops at rates driven by the diffusive gradient (*e.g.*, as determined by the equilibrium condition *via* Henry's Law) and by the mass transfer and mixing characteristics (*e.g.*, molecular diffusion and within-drop mixing). In contrast, particle scavenging results not from equilibrium partitioning, but rather from physical processes controlled by cloud microphysics, meteorological conditions, and the solubility, number density, and size of ambient aerosol particles (7-9). Particles greater than $1 \mu\text{m}$ are incorporated into falling drops by collisional capture and particles that escape capture *via* frontal collisions may become entrapped in standing eddies developed in the lee of falling water drops (10). On the other hand, submicron particles ($< 0.05 \mu\text{m}$) collide with water drops by Brownian motion and particles with radii in the range 0.05 and $1 \mu\text{m}$ are influenced by phoretic (*e.g.*, thermal) and coulombic forces (11,12). Slinn *et*

al. (6) demonstrated that particle scavenging, as described by the collision efficiency of a rain drop with a particle, is a sensitive function of particle size for particles with radii between 0.1 and 10 μm , with a minimum near 0.1 μm and maxima at much less than 0.1 μm and greater than 10 μm , respectively.

Although scavenging of atmospheric particles by precipitation is size-dependent, models describing wet deposition of organic contaminants have not incorporated size-dependent scavenging ratios because few reliable organic contaminant size distribution data were available. Using data gathered from a field study of the removal of PAHs from the atmosphere during five discrete rain events that occurred adjacent to the Chesapeake Bay during late Summer, 1992 (13), in this paper we present precipitation scavenging ratios for PAHs associated with small ($<0.5\mu\text{m}$) and large ($>0.5\mu\text{m}$) aerosol particles. Using temperature-corrected Henry's Law constants, measured concentrations and phase distributions of PAHs in the ambient air and in precipitation (13), and the size distribution of aerosol particle-associated PAHs (14-16), we suggest that gaseous scavenging coefficients previously reported in the literature (17,18) are likely overestimated and that the scavenging of large and small particles by rain is very efficient.

IVB.3 Background

Previous investigators have modeled the overall removal of semivolatile organic contaminants from the atmosphere resulting from the scavenging of gases and particles by precipitation by relating the measured total concentration of contaminant in precipitation (C_{rain}) to the contaminant concentration in the surrounding air (C_{air}) (17,18):

$$W_T = C_{\text{rain}}/C_{\text{air}} = W_g(1-\phi) + W_p \phi \quad [1]$$

where W_T , W_g , and W_p are the total, gas, and particle scavenging ratios, respectively, and ϕ is the fraction of contaminant bound to particles in the atmosphere. The gas scavenging ratio (W_g) is defined as $[C_d] / [C_g]$, where C_d is the dissolved organic contaminant concentration in rainwater and C_g is the gaseous contaminant concentration in the interstitial air. Similarly, W_p in Equation [1] is defined as $[C_p] / [C_{p,\text{air}}]$, where C_p is the particle-associated concentration in rainwater and $C_{p,\text{air}}$ is the concentration of contaminant associated with aerosol particles. The fraction of contaminant bound to particles (ϕ) is defined as $C_{p,\text{air}} / [C_g + C_{p,\text{air}}]$ (17,18). Atmospheric gases in true equilibrium with water droplets will be scavenged according to Henry's Law, and the gas scavenging ratio at equilibrium ($W_{g,\text{eq}}$) therefore equals RT/H where R is the universal gas constant ($8.31 \times 10^5 \text{ atm}\cdot\text{m}^3\cdot\text{mol}^{-1}\cdot\text{K}^{-1}$), T is absolute temperature, and H is the Henry's Law constant ($\text{atm}\cdot\text{m}^3\cdot\text{mol}^{-1}$). The relative importance of gas and particle scavenging in removing organic contaminants from the atmosphere during rainfall depends upon the relative magnitudes of $W_g(1-\phi)$ and $W_p\phi$ (3). Experimental values of W_p for organic contaminants range from 2000 to 10^6 (3 and references within) and W_g values range from about 900 to 31,000 for neutral organic compounds (18). Particle scavenging is a complex and highly variable process affected both by meteorology and by the size spectra of particles and rain drops (7,19,20). Vapor scavenging of HOCs is influenced by their vapor-particle distribution, which is directly related to temperature,

and the surface area and chemical characteristics of the aerosol particles (1,3).

Accurate measurements of organic contaminant scavenging coefficients by precipitation are limited by the difficulties of separating particle-associated compounds from those that are truly dissolved in precipitation. Submicron-sized particles (generally less than 0.2 - 0.7 μm , refs 17, 21) may not be collected quantitatively on glass fiber filters due to slow rates of particle diffusion in water and the presence of double layer charges (17). Observed enhancements of organic contaminants in the filtrate of precipitation relative to that supported by equilibrium absorption of ambient gaseous contaminants have been attributed to the presence of contaminant-laden submicron particles in the filtrate (21-26). Murphy and Rzeszutko (22) report that about 66% of the total polychlorinated biphenyls in rain collected in Chicago passed through a glass-fiber filter. Murray and Andren (21) report similar findings for rain samples collected in Madison, Wisconsin. Schomburg *et al.* (23) and others (24-26) conclude that the enrichment of organic contaminants in filtered fogwater is due to the presence of submicron particle-associated contaminants. We also conclude from our measurements of the speciation of PCBs and PAHs in rain collected adjacent to the Chesapeake Bay that as much as 80% of these compounds are bound to particles not retained by glass fiber filters, and that less than 9% of these hydrophobic chemicals are truly dissolved in rainwater (13). Therefore, reported estimates of particle and gas scavenging of organic compounds may be under- and overestimated, respectively, due to filtration inefficiencies.

IVB.4 Methods and Model Development

The distribution of PAHs between the gas and particulate phases in the atmosphere and between operationally-defined dissolved and particulate phases in precipitation were measured in five consecutive storms at a site on the shore-line of the Chesapeake Bay. Sampling and analytical details are described in the preceding paper (13), and estimates of wet and dry depositional fluxes of PAHs and other organic and inorganic species at this site are described elsewhere (27-29). Of particular relevance to the use of these field measurements to develop a PAH scavenging model are the possible artifacts inherent in any sampling procedures. PAH distributions between the gas and aerosol particle phases were determined by drawing ambient air through a dry glass fiber filter and a polyurethane foam (PUF) plug in series. While it has been suggested that adsorption of gases onto filters may bias this separation (30), PAHs were not detected in significant quantities on clean glass fiber filters placed in the sampling stream between the first filter and the PUF plug (27). Also, we evaluated the efficiency with which the glass fiber filters trapped submicron particles from the air by placing a five stage Berner impactor downstream of the filter (14). After sampling both urban and rural air for 12 hours each, no visible submicron particles were collected on the impactor stages, demonstrating the glass fiber filters efficiently retain atmospheric particles much smaller than their nominal pore sizes. We conclude from these studies that the PAHs measured on the filters and PUF plugs are unbiased estimates of particulate and gaseous concentrations, respectively.

Unlike air, glass fibers are relatively inefficient at separating small particles from rain water. In our studies, the rainwater was filtered during each precipitation event by drawing the

sample through a glass fiber filter and an Amberlite XAD-2 resin column in series. Evidence of small particles passing through the filter include a brown-black discoloration of the head of the resin column and unusually high dissolved organic carbon concentrations in the rainwater (13). Although small particles certainly passed through these wetted filters, there was no evidence that dissolved PAHs adsorbed to the filters, as determined by placing a second filter between the first filter and the resin column in the rain sampling train (31). Minor amounts of high molecular weight PAHs detected on the back-up filter in the precipitation sampler likely resulted from further trapping of very small particles which passed through the first filter.

Equation [1] can be modified to include scavenging of submicron particles present in the filtrate of precipitation samples:

$$W_T = C_{\text{rain}}/C_{\text{air}} = W_{g,\text{eq}}(1-\phi_T) + (W_{p,l}\phi_p) + (W_{p,sm}\phi_{sm}) \quad [2]$$

where W_T is the total scavenging ratio, $W_{g,\text{eq}}$ is the equilibrium gas scavenging ratio, ϕ_T , ϕ_p , and ϕ_{sm} are the fractions of contaminant associated with all particles, with large, filter-retained particles, and with small particles not retained by filters, respectively. For the purposes of the analyses in this paper, we assume a particle size cut-off of $0.5 \mu\text{m}$. Although this is somewhat arbitrary, the following discussion is rather insensitive to the exact cut-off. Murray and Andren (21) report a similar particle cut-off for the glass fiber filters used in their study of PCB distributions in Wisconsin precipitation. The particle scavenging ratio for particles greater than $0.5 \mu\text{m}$ ($W_{p,l}$) is defined as $[C_{p,r, > 0.5 \mu\text{m}}]/[C_{p,\text{air}, > 0.5 \mu\text{m}}]$, and the submicron particle scavenging ratio ($W_{p,sm}$) is defined as $[C_{p,r, < 0.5 \mu\text{m}}]/[C_{p,\text{air}, < 0.5 \mu\text{m}}]$. $C_{p,r, < 0.5 \mu\text{m}}$ can be estimated by taking the difference between measured precipitation filtrate concentrations and that supported by equilibrium gas exchange (13). $C_{p,r, > 0.5 \mu\text{m}}$ is the concentration of contaminant associated with particles retained by a glass fiber filter. Equation [2] shows that the relative contribution of gaseous or particle-associated compounds to rain depends upon the fraction of contaminant associated with large and small particles and upon the equilibrium distribution between the gaseous and dissolved phases. Equation [2] is similar to that suggested by Atlas and Giam (32), but here we modify their model to include ϕ as a function of large and small particles.

To estimate the fraction of each PAH associated with large and small atmospheric particles (ϕ_l and ϕ_{sm}), the total particulate PAH concentration measured on each date was proportioned according to PAH size distributions measured elsewhere. Recently, we have developed and employed a method using a Berner low pressure 5-stage particle impactor to measure PAH size distributions in rural and urban atmospheres (14). For the calculations here, the PAH size distributions measured in Egbert, Ontario during February 1993 were used to estimate the relative amount of each PAH on large ($>0.5 \mu\text{m}$) and small ($<0.5 \mu\text{m}$) particles for the five air samples collected in concert with precipitation sampling at the Chesapeake Bay station (13). For the samples taken at Egbert, Ontario, between 83 and 98% of the total particle-associated PAH concentration occurred on particles less than $0.5 \mu\text{m}$. These distributions are very similar to those previously reported for PAHs and other organic compounds in a variety of atmospheres (15,16,33,34). While it would have been preferable to have measured PAH particle size

distributions coincident with our precipitation sampling, the relatively uniformity in organic contaminant size distributions reported in the literature supports this extrapolation. Mean values of ϕ_r and ϕ_{nr} , calculated utilizing our size distribution data and the ambient gaseous and total particle-associated PAH concentrations measured during our Chesapeake Bay field study, are listed in Table IV.B-1. As expected based on their properties, PAHs with vapor pressures $> 10^{-8}$ atm exist primarily as gases in the atmosphere and have particle fractions less than 0.5. Approximately 90% of the PAH which were associated with atmospheric particles occurred on particles $< 0.5 \mu\text{m}$, and there was no systematic difference in the relative amount of dependence of the PAH size distribution on vapor pressure (Table IV.B-1).

Equilibrium gas scavenging ratios were calculated for each PAH using the temperature-corrected Henry's Law constant and the measured gaseous concentration to estimate the truly dissolved PAH concentrations in the precipitation. Henry's Law constants for PAHs calculated at 25°C using published solubility and vapor pressure data (35-37) were corrected to the ambient air temperatures during sampling using the relationship developed by Tateya *et al.* (38) and compound-specific intercepts (39). The resulting Henry's Law constants are significantly higher than those calculated using an alternate relationship reported by ten Hulscher *et al.* (40). However, the ten Hulscher *et al.* relationship yields PAH Henry's Law constants at 25°C which differ greatly from values calculated from published PAH aqueous solubilities and vapor pressures (35-37). Based on this discrepancy and the limited number of PAHs studied by ten Hulscher *et al.* (40), we chose to correct the PAH Henry's Law constants for temperature using the Tateya *et al.* relationships. Using the ten Hulscher *et al.* temperature correction increases our calculated PAH gas scavenging ratios by a factor of 70. Nonetheless, the contribution of gas equilibration to the overall PAH scavenging is negligible (see below).

IVB.5 Results and Discussion

Measured concentrations and speciations (*i.e.*, filterable particles *versus* filtrate) of PAHs in five discrete rain events at a shoreline station near the Chesapeake Bay (see companion paper; 13) were used to calculate PAH wet scavenging ratios. Total concentrations of PAHs and PCB congeners and their distributions between dissolved, gaseous, and particulate phases are presented in the preceding paper (13). Atmospheric gaseous concentrations of PAHs range from 0.2 pg/m^3 (benzo[ghi]perylene) to 2700 pg/m^3 (phenanthrene). Particle-associated PAHs in ambient air range from $< 0.01 \text{ pg}/\text{m}^3$ (fluorene) to 65 pg/m^3 (benzo[b]fluoranthene). Filter-retained particle associated PAH concentrations range from 0.08 ng/L (anthracene) to 5.4 ng/L (phenanthrene). Concentrations of PAHs in the dissolved phase of rain, calculated as that supported by equilibrium gas exchange, range from 0.0025 ng/L (anthracene) to 0.31 ng/L (phenanthrene). Submicron particle associated PAH concentrations, calculated as the difference between the measured filtrate concentrations and the truly dissolved concentrations predicted by Henry's Law, range from 0.06 ng/L (benz[a]anthracene) to 2.1 ng/L (phenanthrene).

Mean equilibrium gas scavenging ratios ($W_{g,eq}$) at the ambient sampling temperatures range from 13 (standard deviation (SD) = 11) for anthracene to 540 (SD = 420) for benz[a]anthracene

Table IV.B-1. Mean Values of σ for Total, Large, and Small Particles¹

	Particle Fractions					
	σ T ²	STD DEV	$\sigma > 0.5 \mu\text{m}^3$	STD DEV	$\sigma < 0.5 \mu\text{m}^4$	STD DEV
fluorene	0.0028	0.0017	0.0003	0.0002	0.0025	0.0016
phenanthrene	0.013	0.0070	0.0011	0.0006	0.011	0.0064
anthracene	0.28	0.4755	0.020	0.0335	0.26	0.4421
pyrene	0.045	0.0358	0.0037	0.0029	0.041	0.0328
benz[a]anthracene	0.78	0.2862	0.051	0.0188	0.73	0.2674
chrysene	0.43	0.2243	0.030	0.0154	0.40	0.2088
benzo[b]fluoranthene	0.79	0.2048	0.059	0.0153	0.73	0.1895
benzo[k]fluoranthene	0.91	0.0771	0.061	0.0052	0.85	0.0719
indeno[1,2,3,cd]pyrene	0.97	0.0212	0.050	0.0011	0.92	0.0201
benzo[ghi]perylene	0.98	0.0160	0.053	0.0009	0.92	0.0151

¹ $\sigma = C_p / [C_p + C_g]$ where C_p and C_g are the particle and gaseous PAH concentrations, respectively

² for all particles (large and small)

³ for particles $> 0.5 \mu\text{m}$

⁴ for particles $< 0.5 \mu\text{m}$

Table IV.B-2. Mean Precipitation Scavenging Ratios¹ of PAHs, Chesapeake Bay, MD 1992

PAH	Equilibrium Gas Scavenging		Large Particle Scavenging		Submicron Particle Scavenging	
	Ratio: Wg eq	STD DEV	Ratio: Wp > 0.5	STD DEV	Ratio: Wp < 0.5	STD DEV
fluorene	64	46	920000	480000	750000	na
phenanthrene	250	180	1100000	660000	150000	78000
anthracene	13	11	1400000	650000	130000	140000
pyrene	390	280	850000	490000	14000	7900
benz[a]anthracene	540	420	1900000	2100000	21000	14000
chrysene	130	95	1500000	1400000	25000	10000
benzo[b]fluoranthene	460	100	1700000	2600000	20000	11000
benzo[k]fluoranthene	250	92	2600000	3500000	24000	14000
indeno[1,2,3,cd]pyrene	**2	**	3700000	5600000	50000	16000
benzo[ghi]perylene	**	**	2300000	2600000	22000	18000

¹ N = five storms; ratios given to two significant figures

² ** = gaseous compound not detected

(Table IV.B-2). Scavenging ratios for particles with radii greater than $0.5 \mu\text{m}$ ($W_{p,r}$) range from 850000 (SD = 490000) (pyrene) to 3700000 (SD = 5600000) (indeno[123-cd]pyrene). Particle scavenging ratios for particles less than $0.5 \mu\text{m}$ ($W_{p,nr}$) range from 20000 (SD = 11000) (benzo[b]fluoranthene) to 750000 (fluorene). In general, with the exception of fluorene, submicron particle scavenging ratios are much lower than large particle scavenging ratios. This is consistent with observations reported for fine-particle size elements such as Pb, Zn, As, and V, which have lower particle scavenging ratios than do the crustal elements (Fe, Al, Mn, and Mg) which have larger mass median effective diameters (19). Interestingly, large particle scavenging ratios are greater for the less volatile PAHs, while the submicron particle scavenging ratios decrease with volatility. This may imply subtle shifts in the particle-size distributions of the various PAHs. The large standard deviations associated with the PAH scavenging ratios demonstrate the variability in scavenging mechanisms among the storms.

Gas scavenging ratios observed in this work are ten to 100 times less than those reported previously for rain sampled in Portland, Oregon (18, Table IV.B-3). In that study, W_g was calculated as the ratio of the PAH concentrations measured in the filtrate (*e.g.*, truly dissolved and bound to submicron particles) and measured in the gas phase. Although Ligocki *et al.* conclude that their measured W_g values at 8°C are consistent with equilibrium gas scavenging and that gas scavenging predominates total PAH washout, the reported washout values are likely overestimated due to the inclusion of submicron particles in the dissolved phase. We recalculated $W_{g,eq}$ (RT/H) values from the original Portland data (41), using Henry's Law constants at 8°C estimated by the method used in this study (13). Using these values and the air concentration data reported for Portland, we calculate that up to 80% of the PAH concentrations measured in filtered rainwater in Portland are due to PAHs associated with submicron particles

Particle scavenging ratios for large particles measured in this study are one to three orders of magnitude greater than the particle scavenging ratios reported by Hart *et al.* (42) and Ligocki *et al.* (17). Hart *et al.* report that W_p for Σ -PAHs ($N = 16$) in rain in Dübendorf, Switzerland ranges from 1000 to 200000 and Ligocki *et al.* (17) found that W_p values for individual PAHs range from 2200 to 17000 in Portland, Oregon (Table IV.B-3). Interestingly, the submicron particle scavenging ratios estimated from our data agree well with the filter-retained particle scavenging ratios of these earlier studies. Differences among studies may result either from operational differences in filtration techniques or from actual variations in the amount and type of particles present in precipitation. Rain collected in Dübendorf, Switzerland and Portland, Oregon was filtered through $0.2 \mu\text{m}$ silver and Teflon membrane filters, respectively, to separate the particle-associated compounds from the dissolved phase, and filtration artifacts were not considered in those studies.

PAH particle scavenging ratios should be similar to the ratios of other compounds associated with atmospheric particles (*e.g.* trace metals, radionuclides; 43). However, reported particle washout ratios for the below-cloud scavenging of clays, trace metals, and radionuclides range from 500 to 1500 (44,45). More recently, lower scavenging ratios for lead, cadmium and copper were reported (69, 129, and 67, respectively; ref. 46). Scavenging ratios of Chernobyl

Table IV.B-3. Precipitation Scavenging Ratios Compared to Other Measurements

PAH	Wg.eq	Wg.eq	Wp	Total Washout, Wt		
	this work Mean ¹	Portland Oregon ²	Portland Oregon ²	this study	Isle Royale ³	Portland, OR ²
fluorene	64	1500	15000	2200	n/a	1600
phenanthrene	248	3400	17000	3100	790	3500 ²
anthracene	13	1900	n/a	60000	6090	n/a
pyrene	390	5900	9300	4000	3300	6100
benz[a]anthracene	457	12000	1300	110000	51500	4000
chrysene	132	18000	2600	54000	32100	7000
benzo[b]fluoranthene	680	7400 ²	2200 ²	120000	n/a	2300 ²
benzo[k]fluoranthene	350	n/a	n/a	180000	n/a	n/a
indeno[1,2,3,cd]pyrene	2506	n/a	n/a	240000	174000	n/a
benzo[ghi]perylene	7218	n/a	3100	140000	248000	3100

¹ equilibrium gas scavenging coefficient; calculated as RT/H; N = 5; mean temperature 24.8°C

² rain at 8°C; ref 17, 18; benzo[b]-averaged with benzo[k]fluoranthene; phenanthrene + anthracene

³ ref 50; measured values

radionuclides associated with particles between 0.5 and 1 μm were about 400, which is in good agreement with expected scavenging ratios for continental aerosols in this size range (47). All of these literature values are much lower than those observed for PAHs in this work, suggesting that PAHs are associated with particles of different sizes and composition than are these inorganic species. Gatz (44) and Hewitt and Rashad (46) report that washout ratios of many atmospheric compounds increase with distance from the emission sources. Jafferezo and Colin (48) and Tschiersch *et al.* (49) conclude that scavenging ratios likely increase with decreasing particle size. Therefore, the PAH particle scavenging ratios reported here suggest that PAHs associated with smaller particles may have traveled further from the emission source relative to aerosol particles enriched with trace elements and radionuclides. Alternatively, Gatz (44) reported inorganic particle scavenging ratios on the order of 10^5 to 10^6 , which are consistent with particle scavenging ratios for in-cloud scavenging processes (7). Hence, the elevated ratios observed in this study may be evidence of in-cloud scavenging of PAHs rather than below cloud washout. In any event, the important observation is that particle-associated PAHs are removed from the atmosphere by rain scavenging with great efficiency for both large ($> 0.5 \mu\text{m}$) and small ($< 0.5 \mu\text{m}$) particles.

Total washout ratios (W_T) calculated from equation [2] of PAHs range from 2200 (SD = 1200) for fluorene to 240000 (SD = 290000) for indeno[123-*cd*]pyrene (Table IV.B-4). These values are larger than those reported by others who conducted similar field studies (Table IV.B-3; 17,50), again suggesting either large spatial variability in scavenging or differences in sampling methodologies. The dominant scavenging process for each PAH depends on the relative magnitude of each process term in equation [2]. In this work, the product of $W_{g,cs}(1-\phi_T)$ is much less than 10% of either $W_p(\phi_p)$ or $W_m(\phi_m)$, indicating that gas scavenging contribute a negligible amount of PAHs to precipitation, even for those PAHs that exist predominately in the gas phase in the atmosphere (Figure IV.B-1, Table IV.B-4). Submicron particle scavenging is more effective for removing fluorene, phenanthrene and anthracene from the atmosphere during these rain storms. The remaining PAHs are predominantly removed from the atmosphere by the rain scavenging of particles $> 0.5 \mu\text{m}$.

IVB.6 Summary

Based on the measured speciation of PAH in the atmosphere and in precipitation, we conclude that equilibrium gas scavenging is not an important removal processes for PAHs during precipitation events. In contrast, incorporation of both large (*i.e.*, filter-retained) and small (submicron) particles into rain drop is extremely efficient, leading to elevated concentration of particle-associated PAHs in precipitation. Because larger particles may be more efficiently scavenged by precipitation than small particles, subtle shifts in the particle size distributions of PAHs and other contaminants in the atmosphere may contribute to the exceptionally high interstorm variability in PAH levels in rainwater. Regardless of the mechanisms responsible, the observed variability in PAH scavenging observed during these five storms suggests that a single set of gas and particle scavenging coefficients is inadequate to accurately model the concentrations of organic contaminants in precipitation. Subsequent estimates of regional wet depositional fluxes of organic contaminants calculated by multiplying measured or modeled atmospheric contaminant

Table IV.B-4. Precipitation Scavenging of Particle-Associated and Gaseous PAHs

PAH	Total Scavenging Ratio: W_i	STD DEV	$W_{p>0.5} \cdot \sigma > 0.5$	STD DEV	$W_{p<0.5} \cdot \sigma < 0.5$	STD DEV	$W_{g,eq} \cdot (1-\sigma T)$	STD DEV	Dominant Scavenging Mechanism ¹
fluorene	2200	1200	280	222	1900	1163	63	46	s
phenanthrene	3100	1600	1200	994	1700	1289	242	175	s
anthracene	60000	82000	27000	47744	33000	67116	10	10	s
pyrene	4000	3100	3100	3088	580	562	368	267	p
benzo[a]anthracene	110000	110000	88000	113641	16000	11871	118	180	p
chrysene	54000	49000	44000	48353	10000	6694	74	61	p
benzo[b]fluoranthene	120000	160000	103000	155868	15000	9131	96	96	p
benzo[k]fluoranthene	180000	220000	160000	215132	20000	12007	21	21	p
indeno[1,2,3-cd]pyrene	240000	290000	190000	282340	46000	14826	**	na	p
benzo[ghi]perylene	140000	140000	120000	140403	21000	16951	**	na	p

¹ s = submicron particle scavenging; p = large particle scavenging

inventories by constant scavenging coefficients likely contain unacceptable levels of error. Further characterization of meteorological parameters, including the intensity of precipitation and the size distributions of the rain drops, as well better measurements of aerosol particle size distributions and compositions are required in order to more accurately describe the removal of organic contaminant from the atmosphere by precipitation.

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IVB.8 Figure Caption:

- IV.B-1. Relative scavenging (mean values) of submicron, filter-retained, and gaseous PAHs in rain. Data are presented in Table IV.B-4.

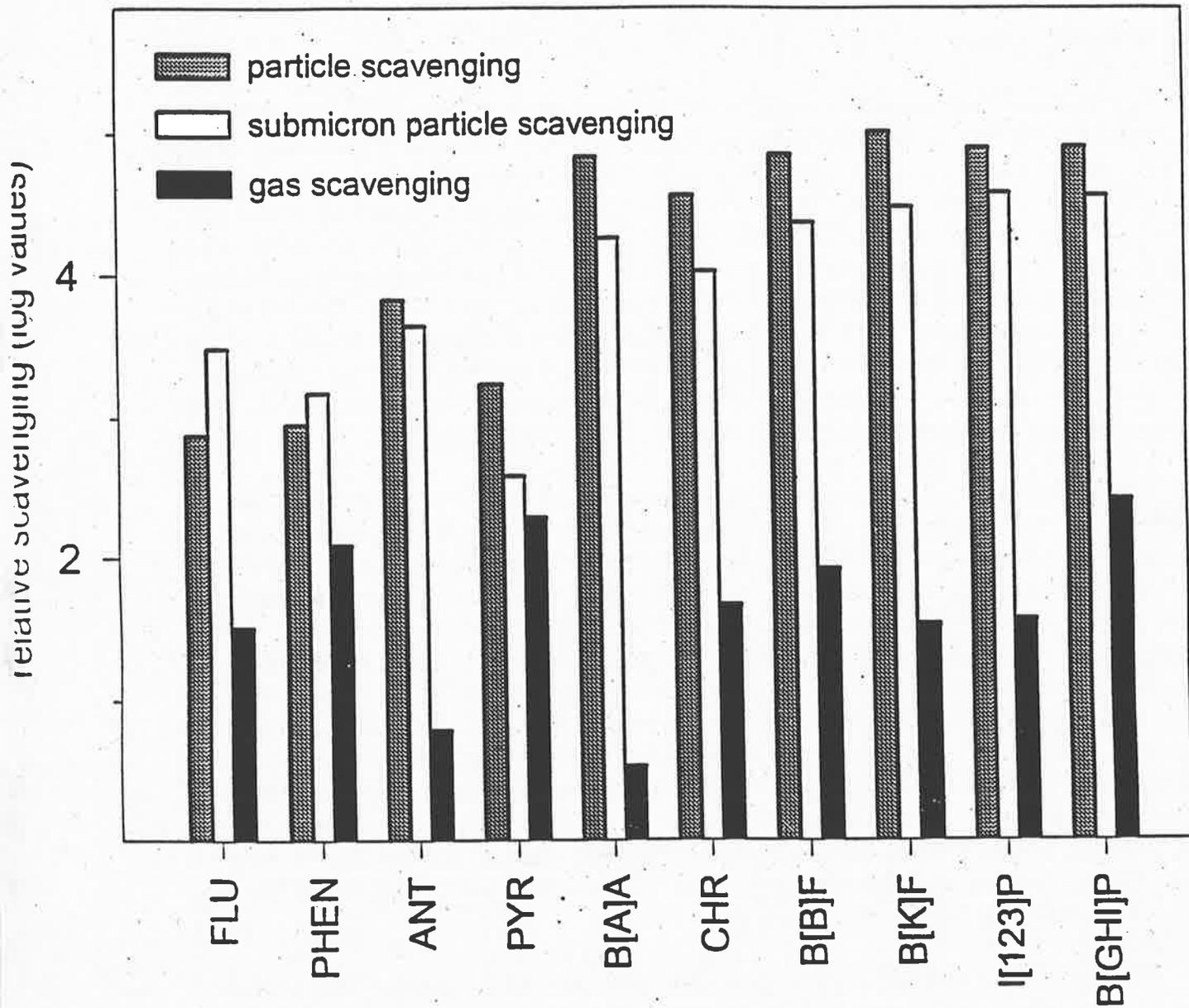


Figure IV.B-1.

V. Conclusions

During 1990-1993, atmospheric deposition of trace elements and organic contaminants to the Chesapeake Bay was monitored at three shore-line locations. Estimates of wet depositional fluxes at a given location are likely precise to within 20-30%; driven by the analytical uncertainty in the precipitation analysis. Extrapolated bay-wide wet deposition loadings are less certain and in this study did not include possible near-field influences of urban areas. Dry aerosol deposition fluxes are much less certain (precision perhaps within a factor of two or three) due to the inexact modeling of aerosol deposition velocities. Diffusive exchange of gaseous organic chemicals across the air-water interface, which may be the dominant transport process for many important species, is not included in this study. As a result of assuming a constant dry aerosol deposition velocity with time and not including gas exchange, the results presented here likely understate the actual seasonal variation in atmospheric deposition to the Chesapeake Bay.

Concentrations of trace elements and organic contaminants in air over the rural Chesapeake Bay are similar at the three sites, and annual averages are influenced by occasional spikes in concentration at each of the sites. Although the highest concentrations in rainwater coincide with small precipitation events, monthly wet depositional fluxes of trace elements and organic contaminants do not significantly vary throughout the year. Concentrations and fluxes of PAHs appeared to decrease between 1990 and 1993, while those of most trace elements remained relatively constant. Concentrations of organic contaminants in rainwater in the Chesapeake Bay region are lower than comparable measurements in the Great Lakes, despite higher atmospheric inventories. Whether this interregional difference reflects more efficient wet scavenging process in the colder, drier Great Lakes, or simply results from methodological difference between the two studies is unclear. Estimates of total atmospheric deposition are within a factor of two between the Chesapeake Bay and the Great Lakes, suggesting that both regions share a common 'regional background' atmospheric deposition signal. The potential for elevated deposition near urban areas is not considered in this comparison.

Comparing the magnitude of atmospheric deposition relative to other possible sources of trace elements and organic contaminants to the Chesapeake Bay is complicated both by inconsistencies in the methods and study periods, and by fundamental differences in the speciation and geochemical reactivity of materials carried by each source. Atmospheric deposition directly to the surface waters of the Chesapeake Bay supplies loads of polycyclic aromatic hydrocarbons which are comparable to or greater than the loads of dissolved PAHs delivered by the Susquehanna River. In contrast, dissolved total PCB loads from the river are approximately three times those from the atmosphere. Particulate-bound organic contaminants discharged from the Susquehanna River are considerably larger, especially during high flows. Atmospheric depositional fluxes of several elements, including lead, cadmium, and chromium, are within a factor of two of the dissolved load from the Susquehanna River. Again, particulate metal loads from the river dominate over both dissolved riverine loads and atmospheric deposition.

In intensive, event-based sampling at the mid-bay Elms site during Summer, 1992,

scavenging of semivolatile organic contaminants by precipitation was studied. By comparing the measured levels of PAHs and PCBs in filtered rainwater to those concentrations predicted to be in equilibrium with the surrounding gas phase, an apparent over enrichment of these contaminants in rainwater was observed. Based upon the high levels of non-filter-retained organic carbon in these rainwater samples, and upon the observed blackening of the adsorption resin in the precipitation sampling train, we conclude that the very efficient washout of submicron particles by raindrops contributes a substantial portion of the overall wet depositional flux. In contrast, gaseous contaminants are not effectively scavenged by raindrops under summertime conditions, and contribute little to the wet depositional fluxes measured here.

The Chesapeake Bay Atmospheric Deposition Study has established annual atmospheric deposition loads to the surface waters of the Chesapeake Bay during the early 1990's. Spatial and temporal variability in atmospheric concentrations and deposition fluxes to rural areas of the bay have been documented. Future studies should address the several important areas not investigated here, including (1) the influence of urban areas on enhanced localized atmospheric deposition, (2) diffusive exchange of gaseous pollutants, including mercury, across the air-water interface, (3) the influence of meteorology, surface conditions, and aerosol size distributions on the dry aerosol deposition rates, (4) deposition rates to and subsequent transmission through the various types of land uses in the watershed, and (5) the fate and reactivity of atmospherically-deposited materials in estuarine waters.

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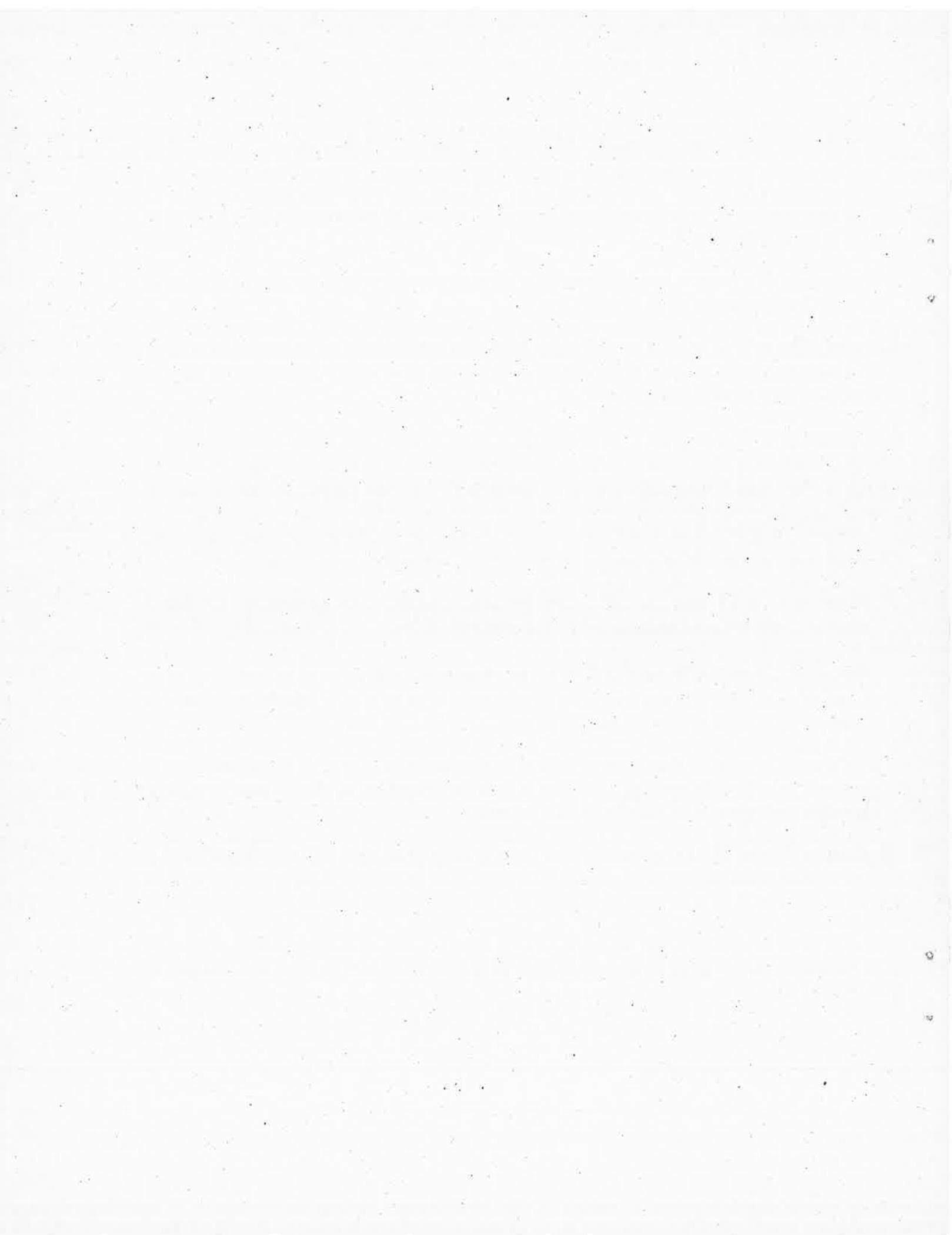
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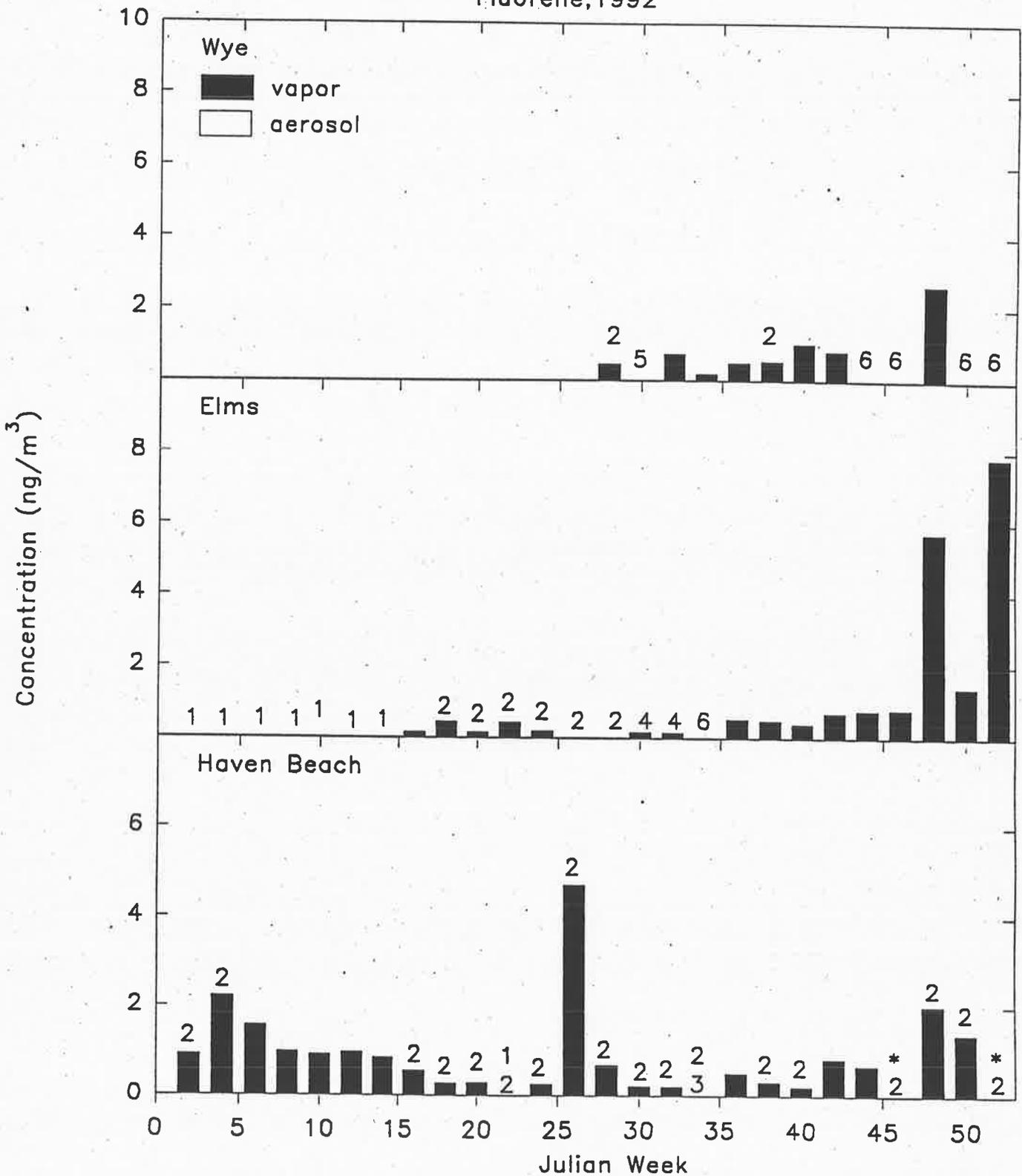
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Appendix A1

Fluorene, 1992

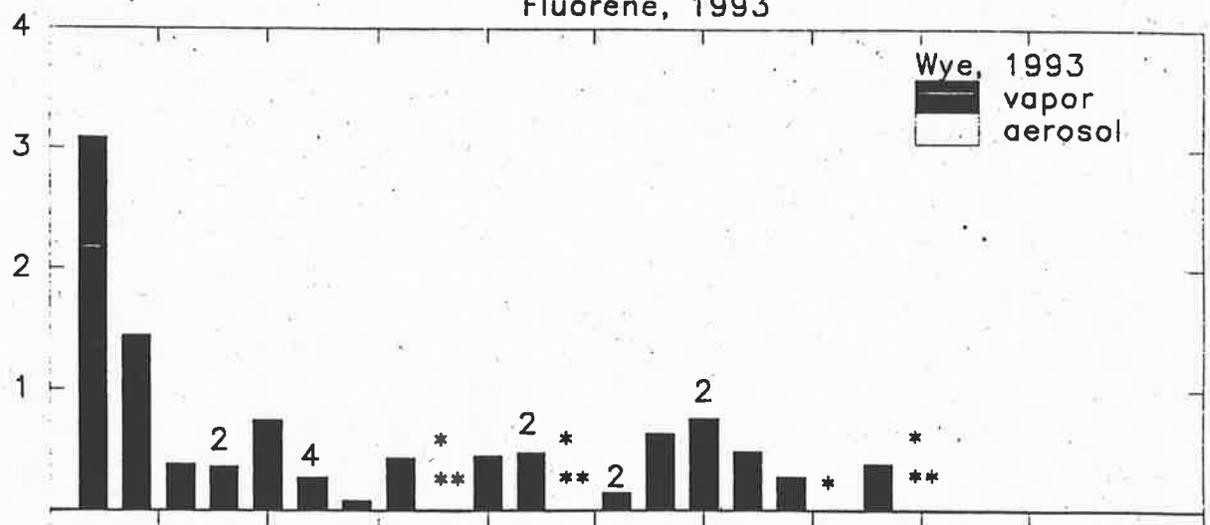


A1.1 Fluorene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Fluorene, 1993

Wye, 1993
 vapor
 aerosol

Concentration (ng/m³)



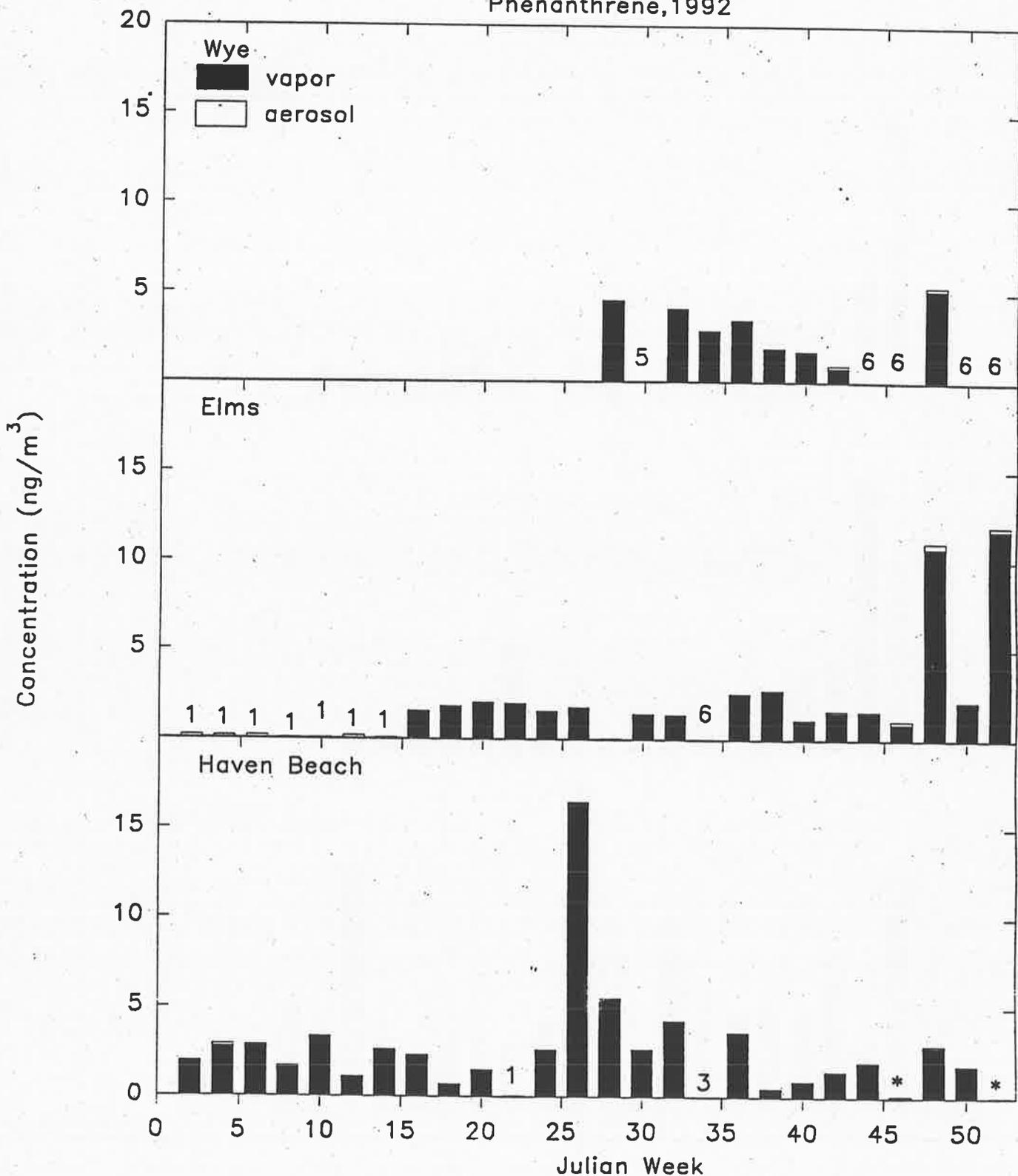
Elms



Haven Beach

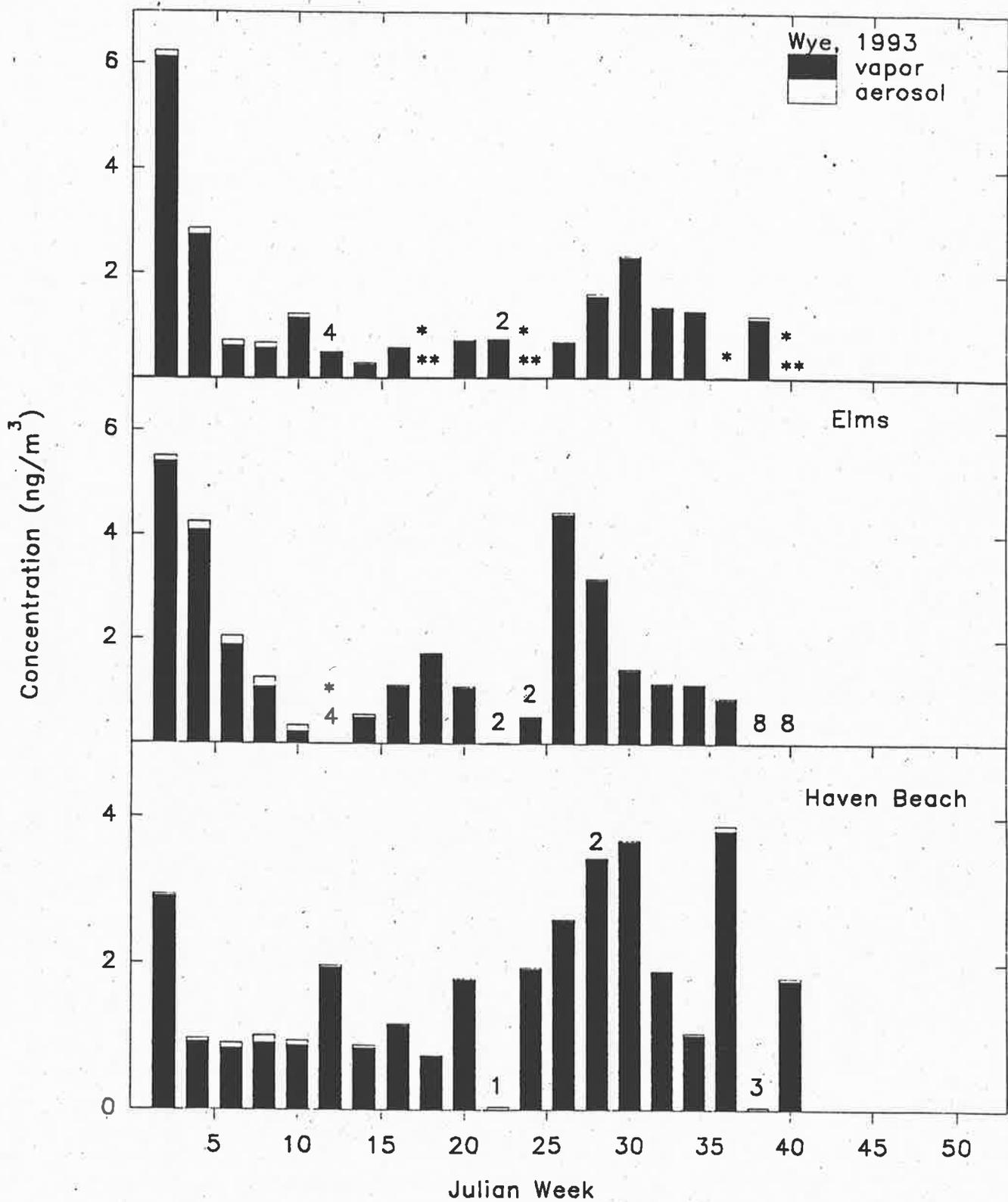
Julian Week

Phenanthrene, 1992

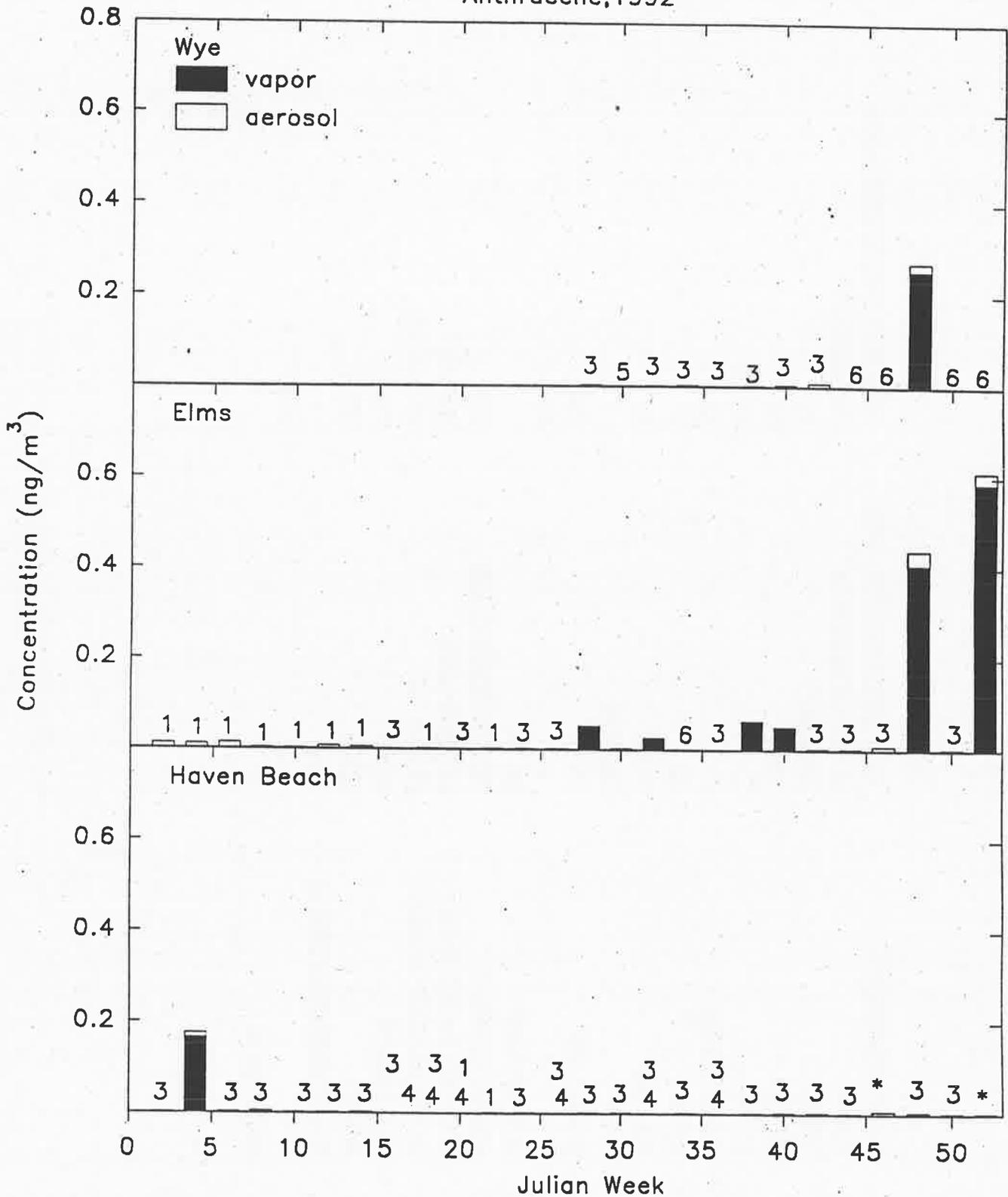


A1.2 Phenanthrene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Phenanthrene, 1993

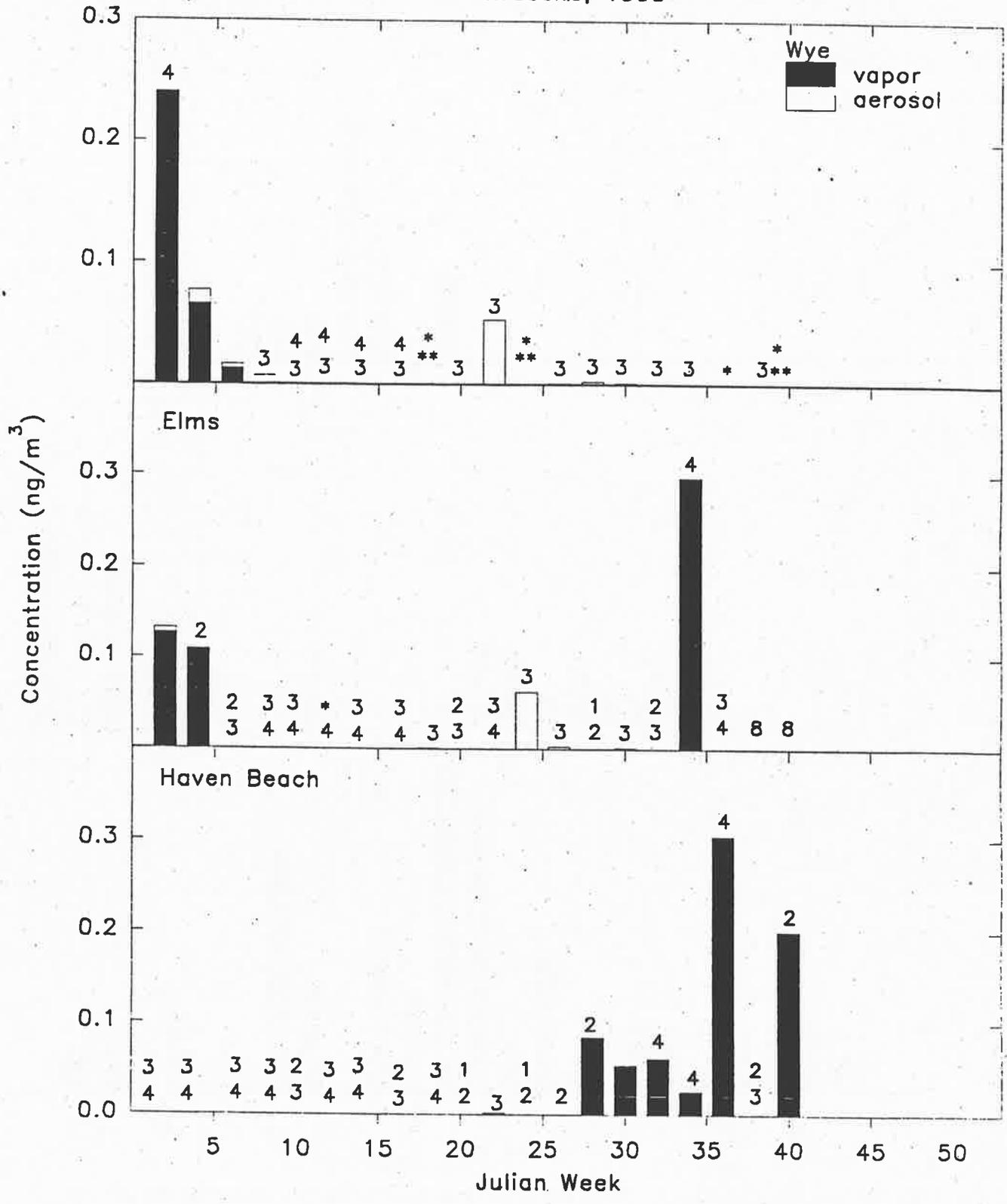


Anthracene, 1992

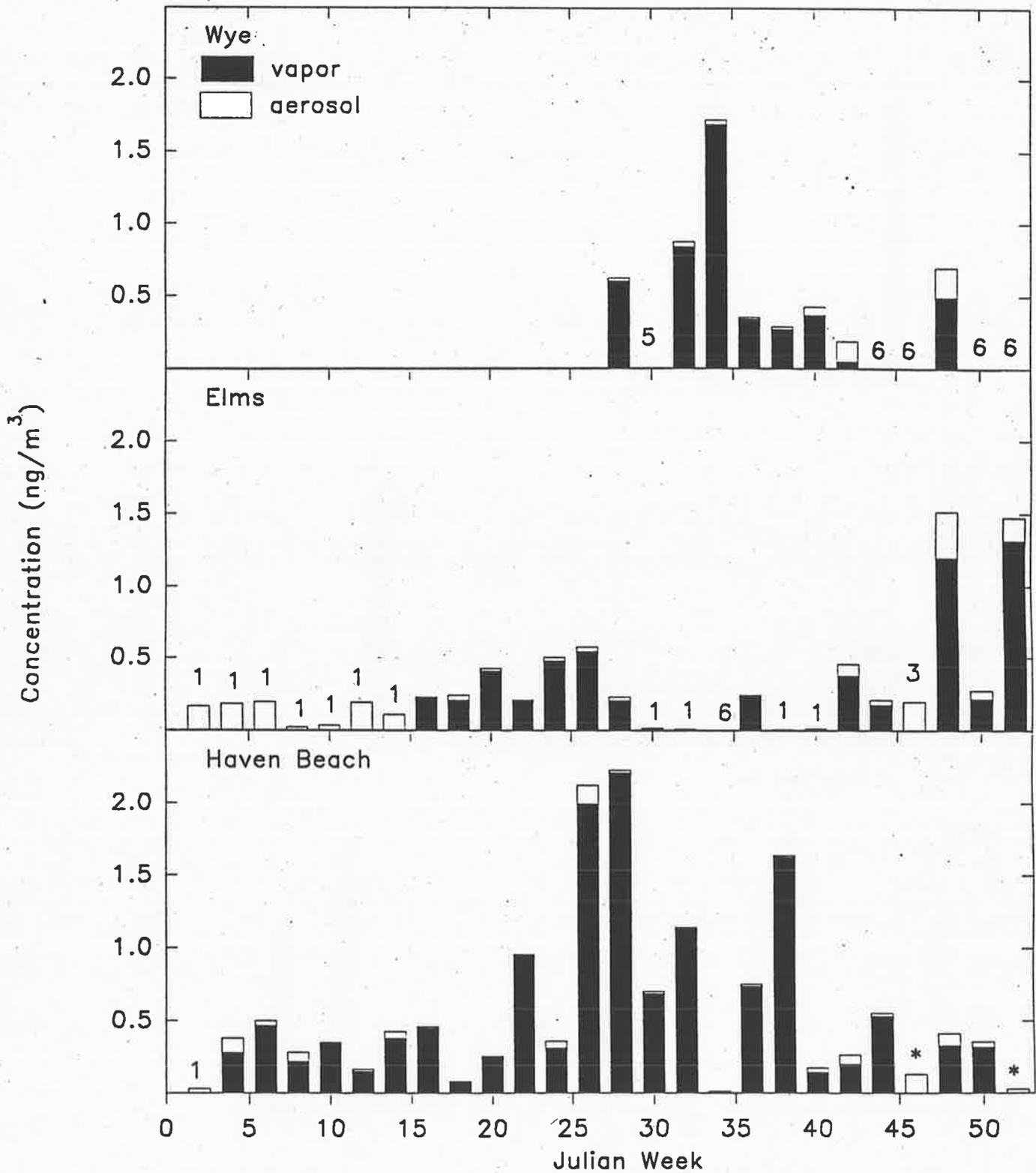


A1.3 Anthracene concentrations collected biweekly in air at Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Anthracene, 1993

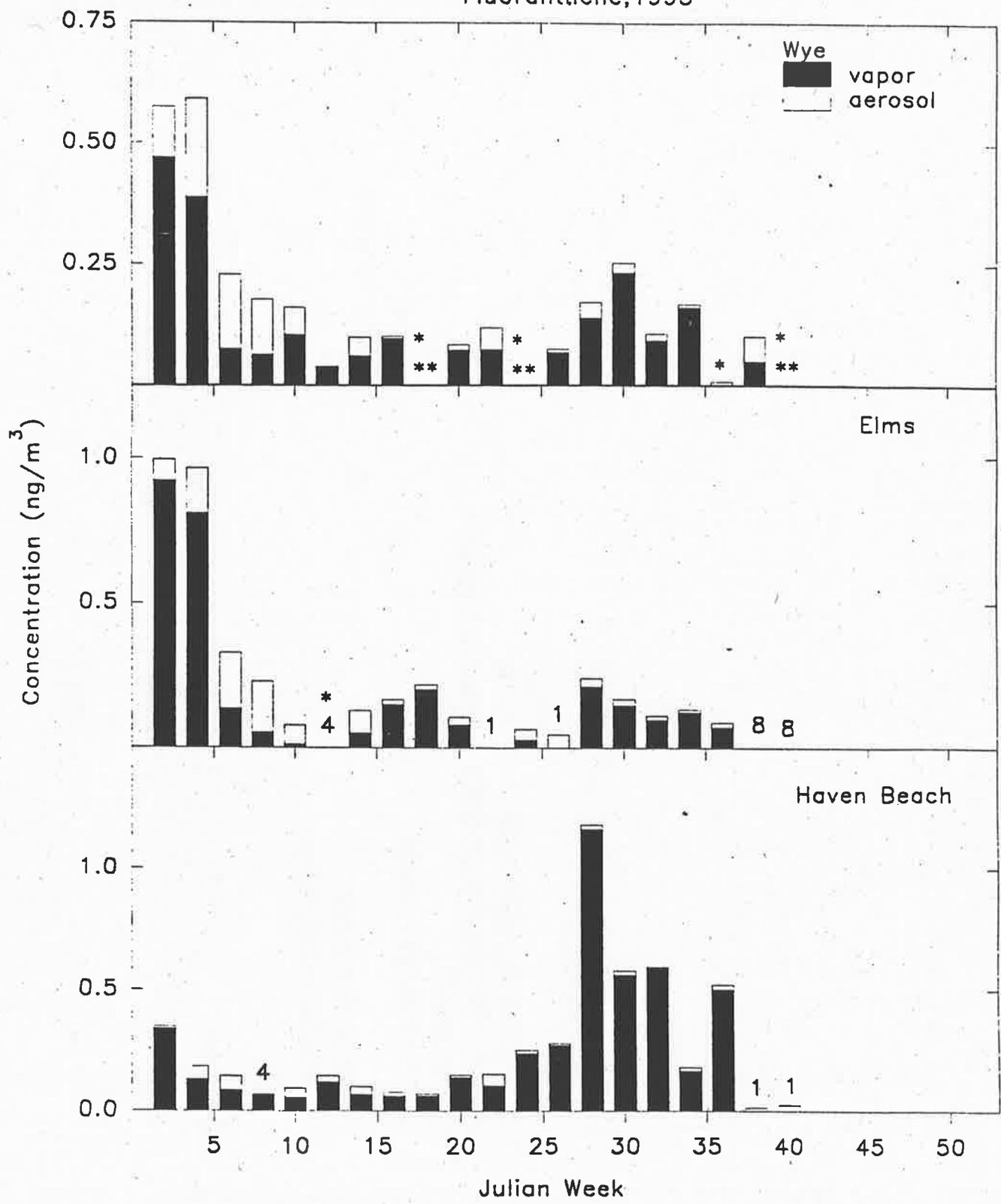


Fluoranthene, 1992

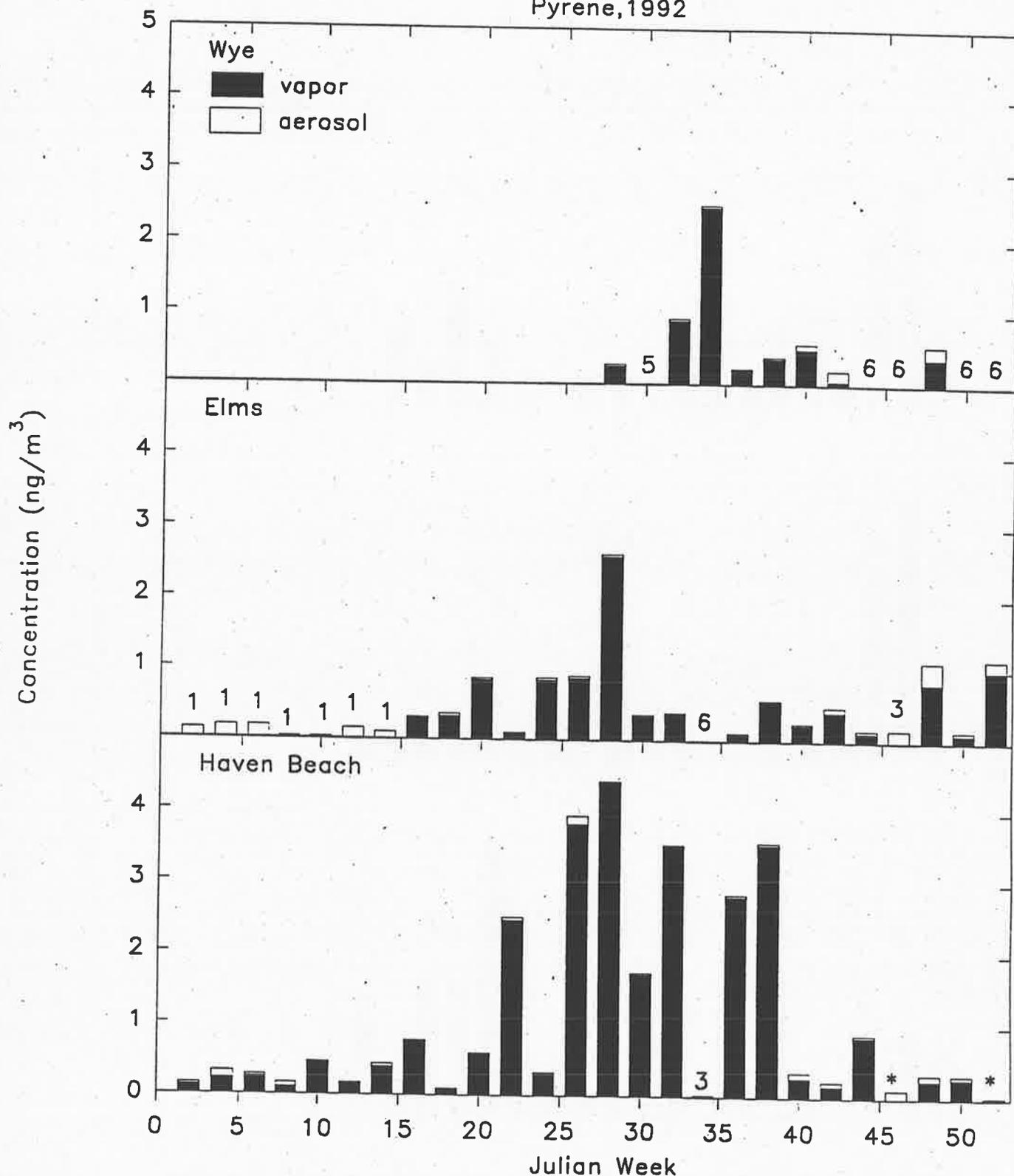


A1.4 Fluoranthene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Fluoranthene, 1993

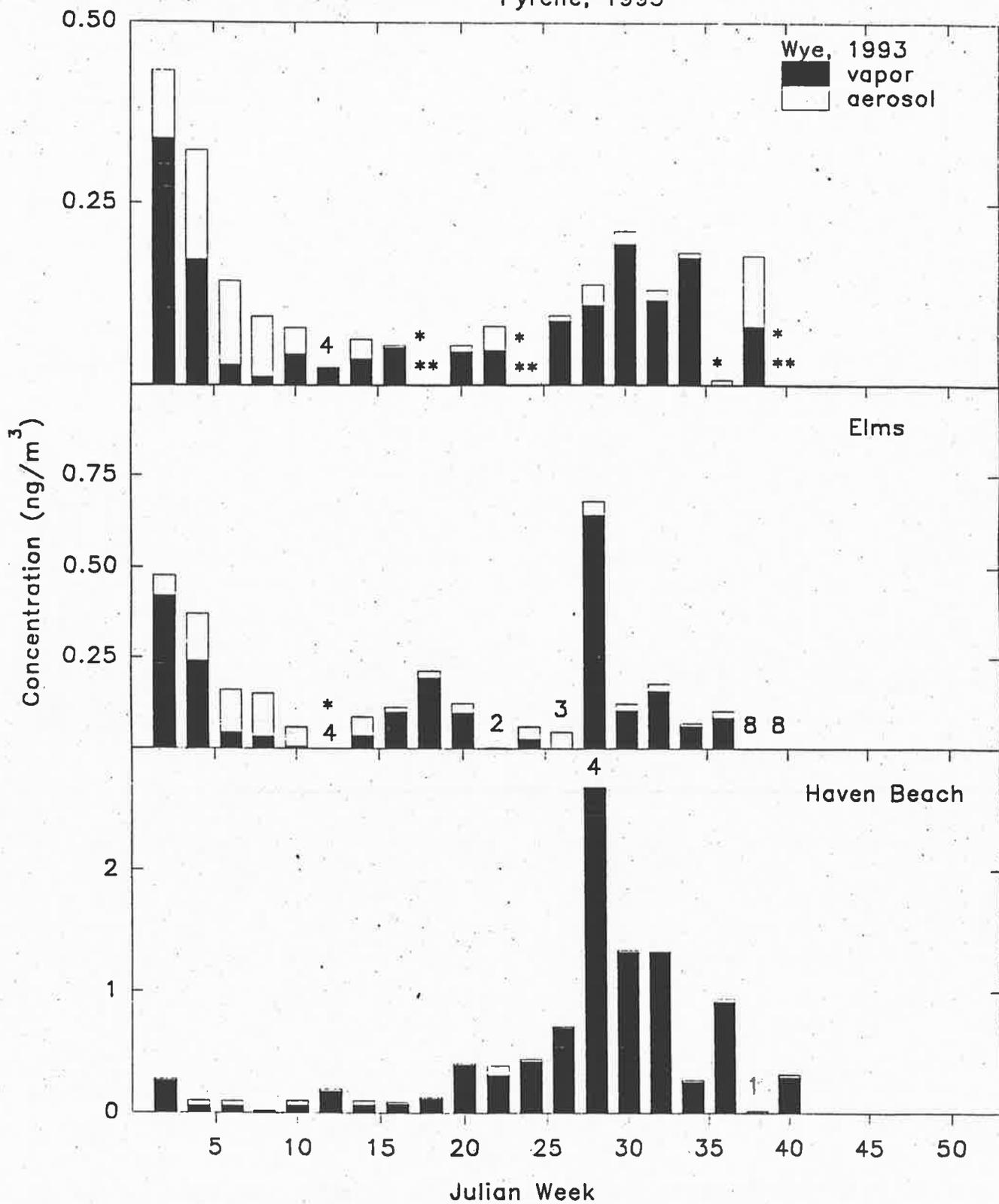


Pyrene, 1992

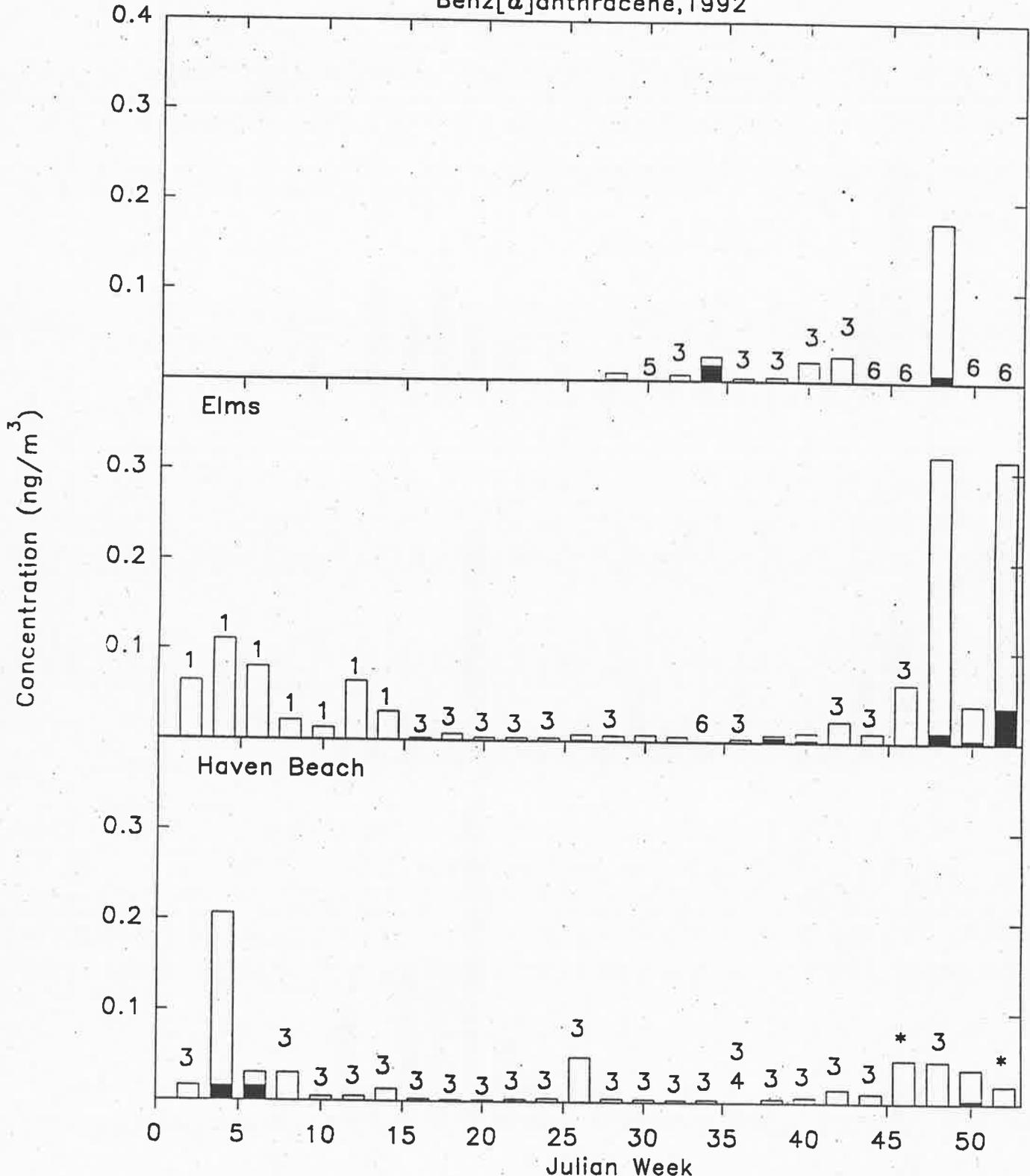


A1.5 Pyrene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank; 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Pyrene, 1993

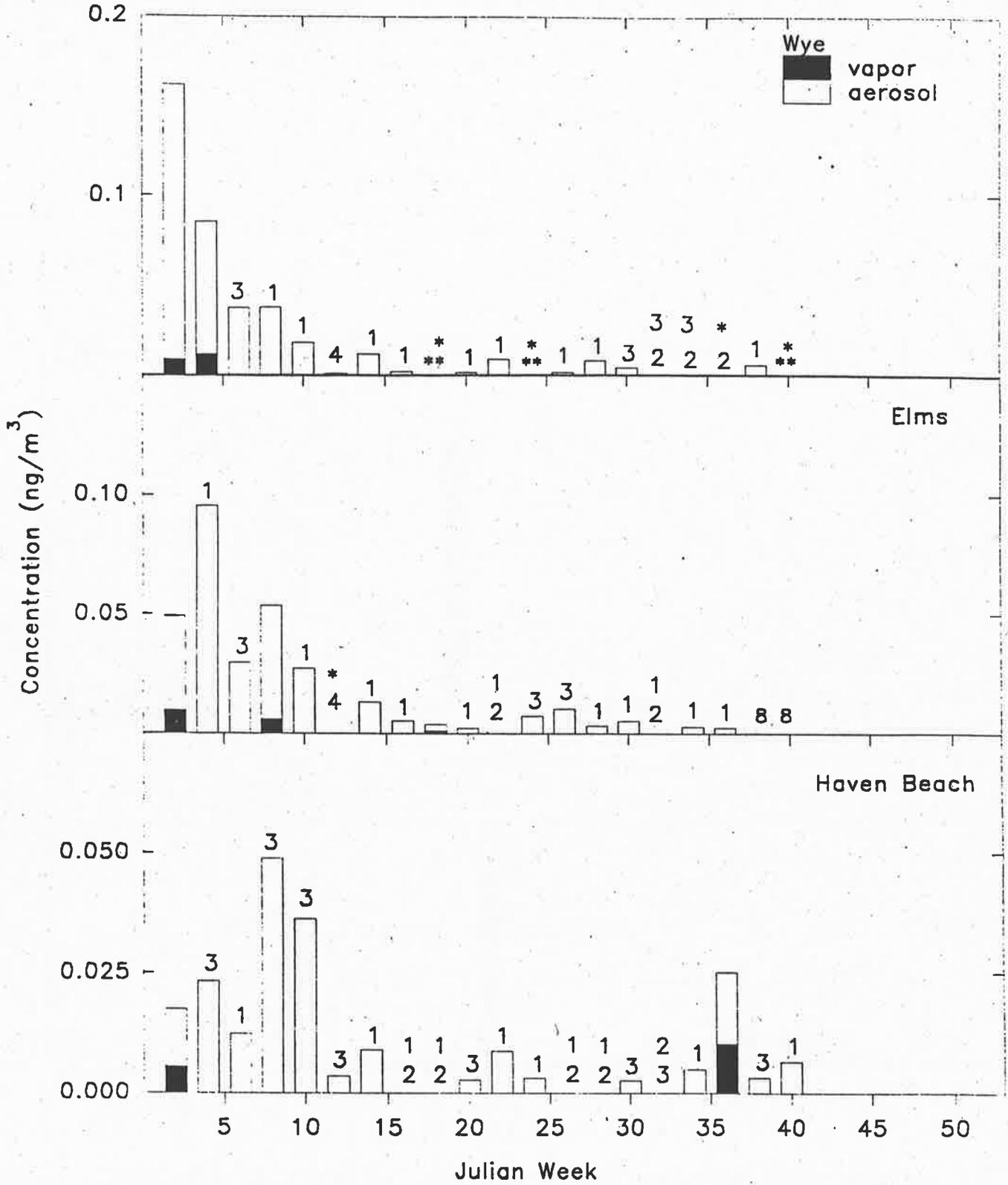


Benz[a]anthracene, 1992

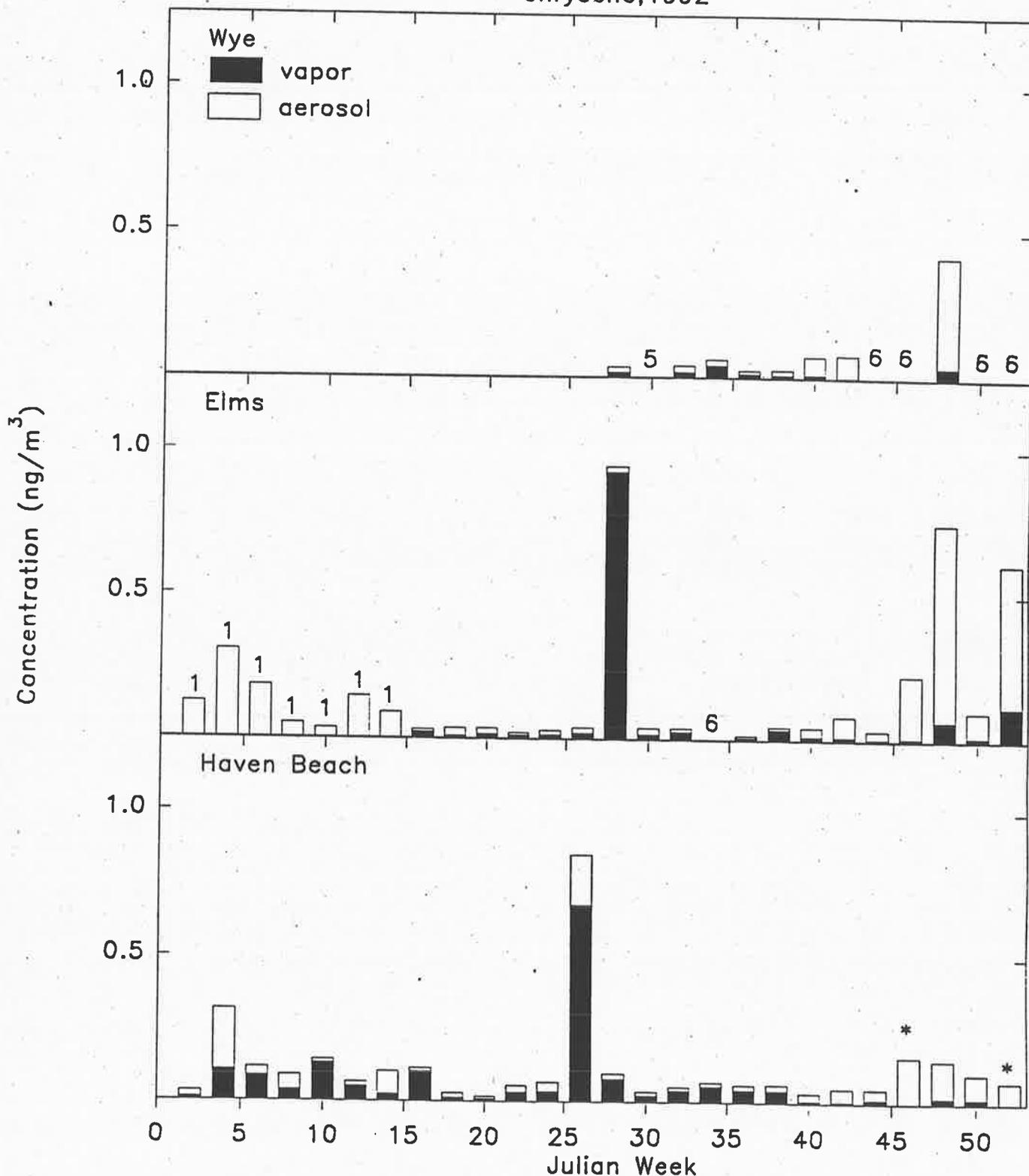


A1.6 Benz[a]anthracene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benz[a]anthracene, 1993

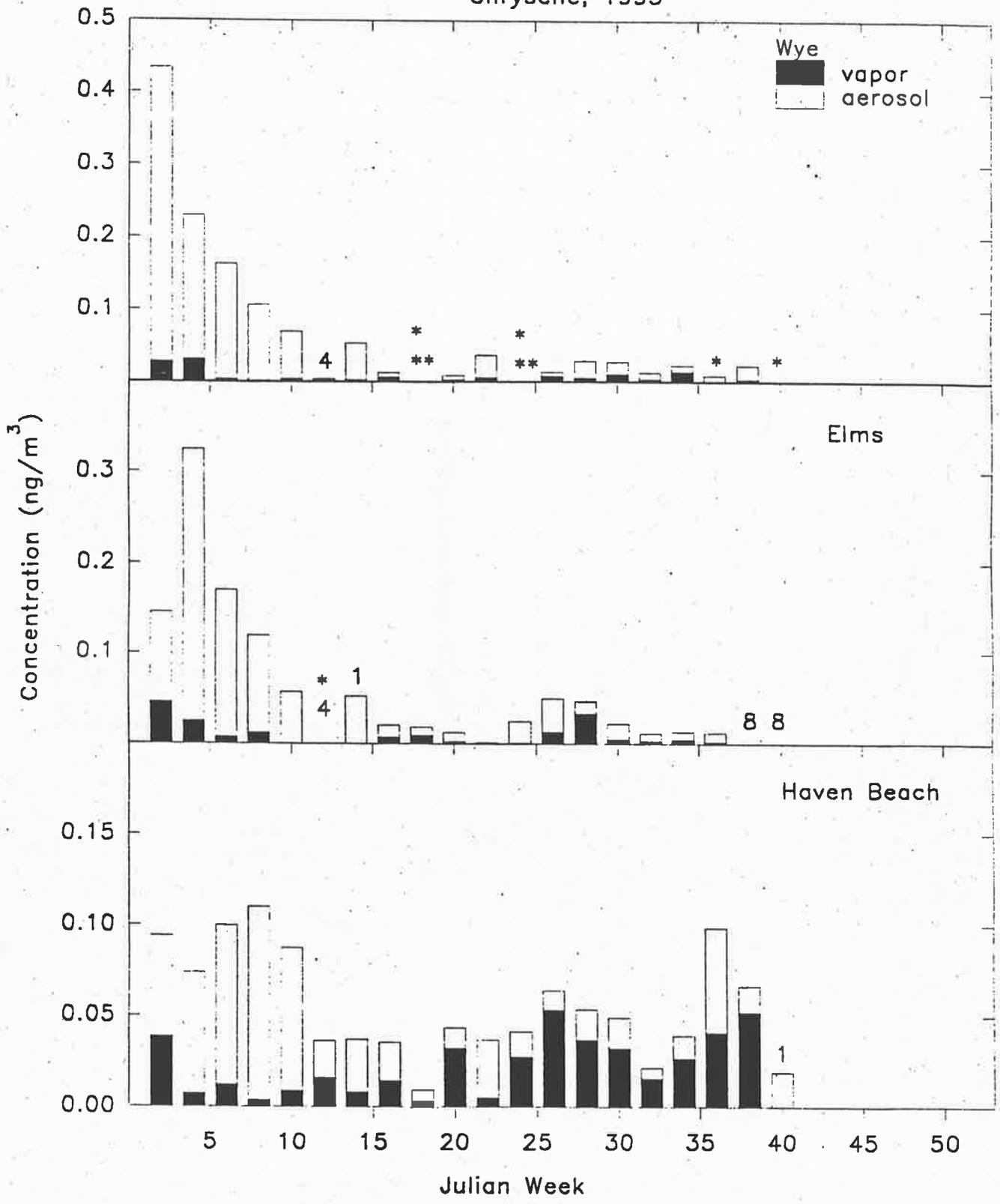


Chrysene, 1992

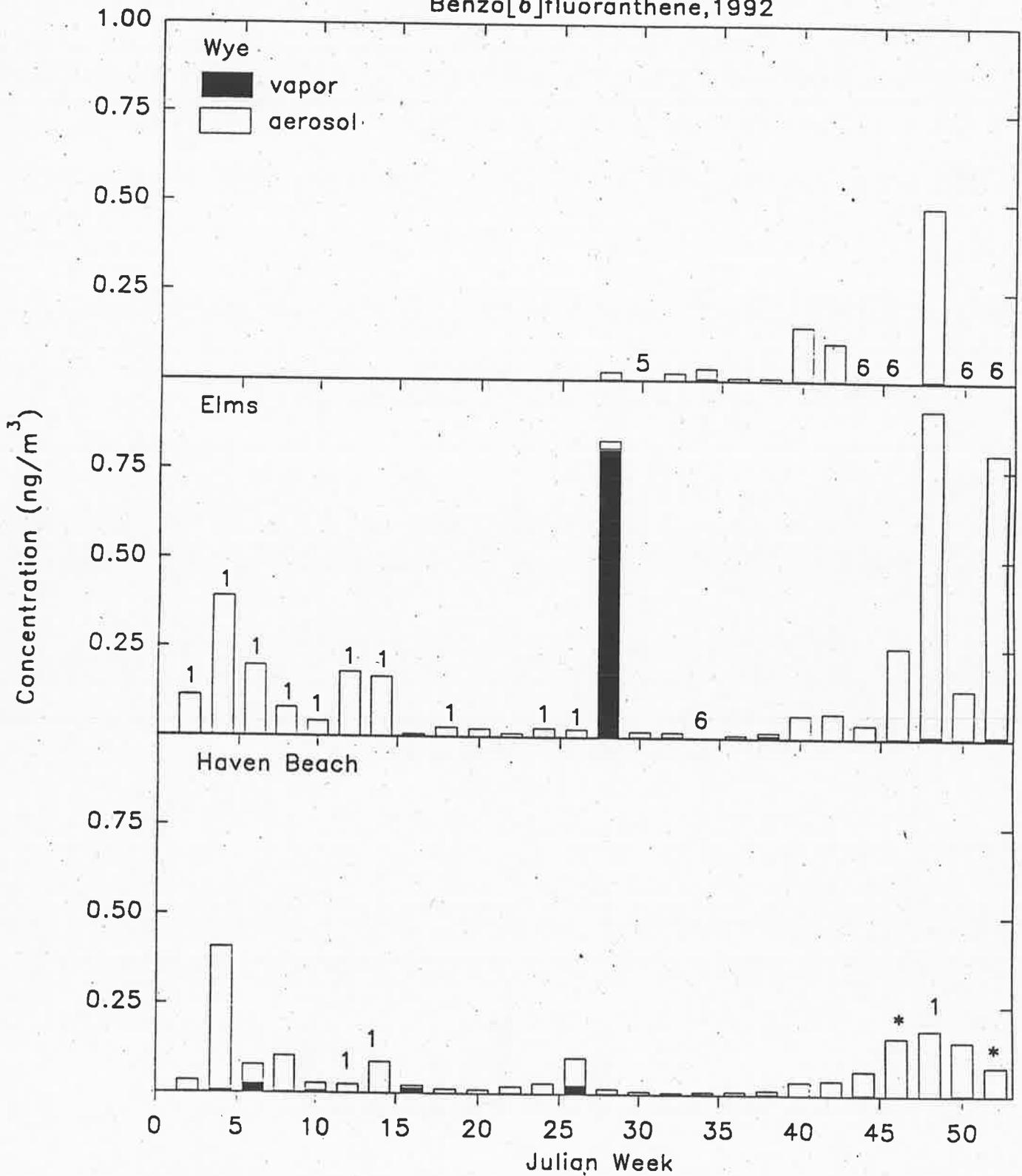


A1.7 Chrysene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Chrysene, 1993

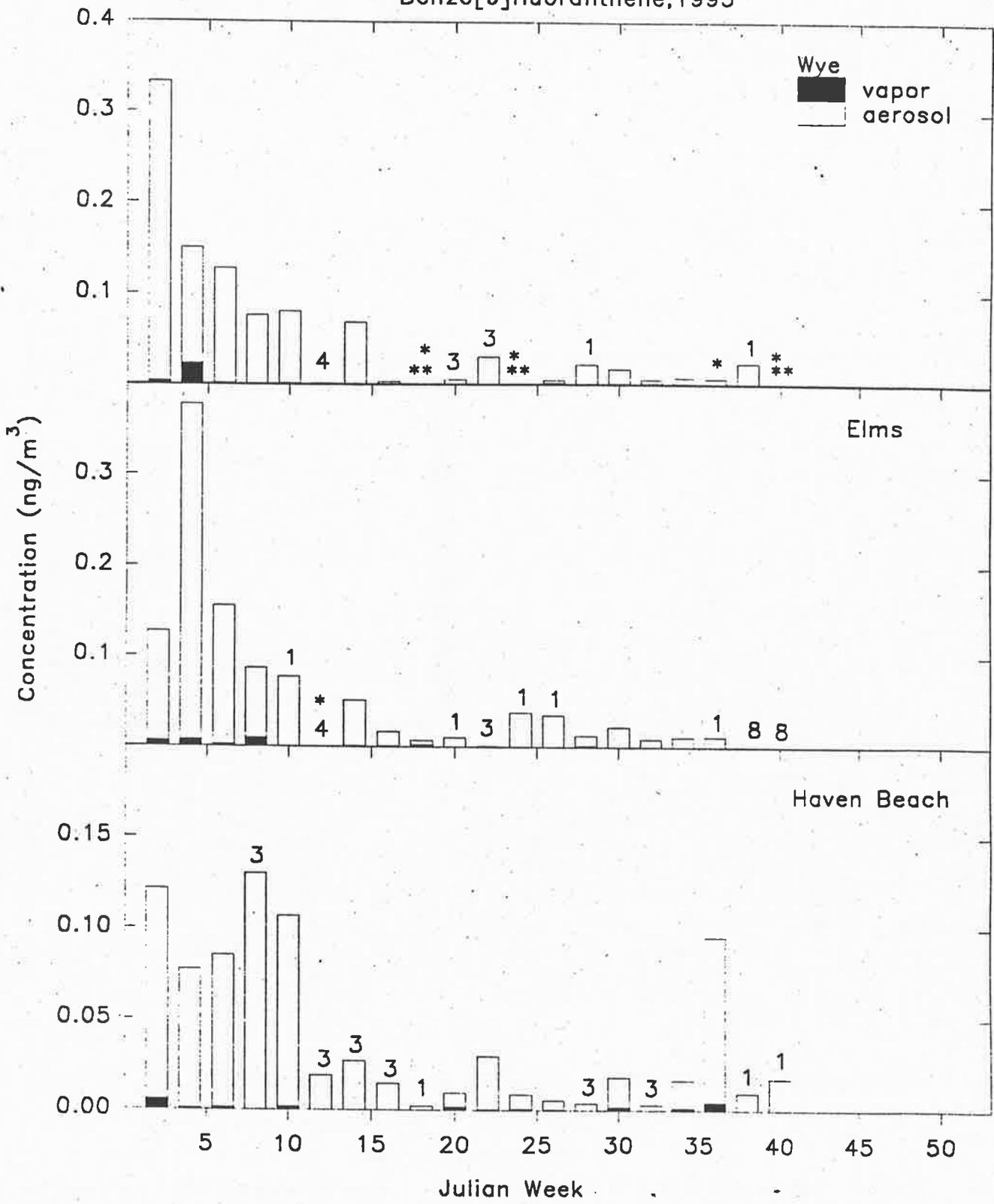


Benzo[b]fluoranthene, 1992

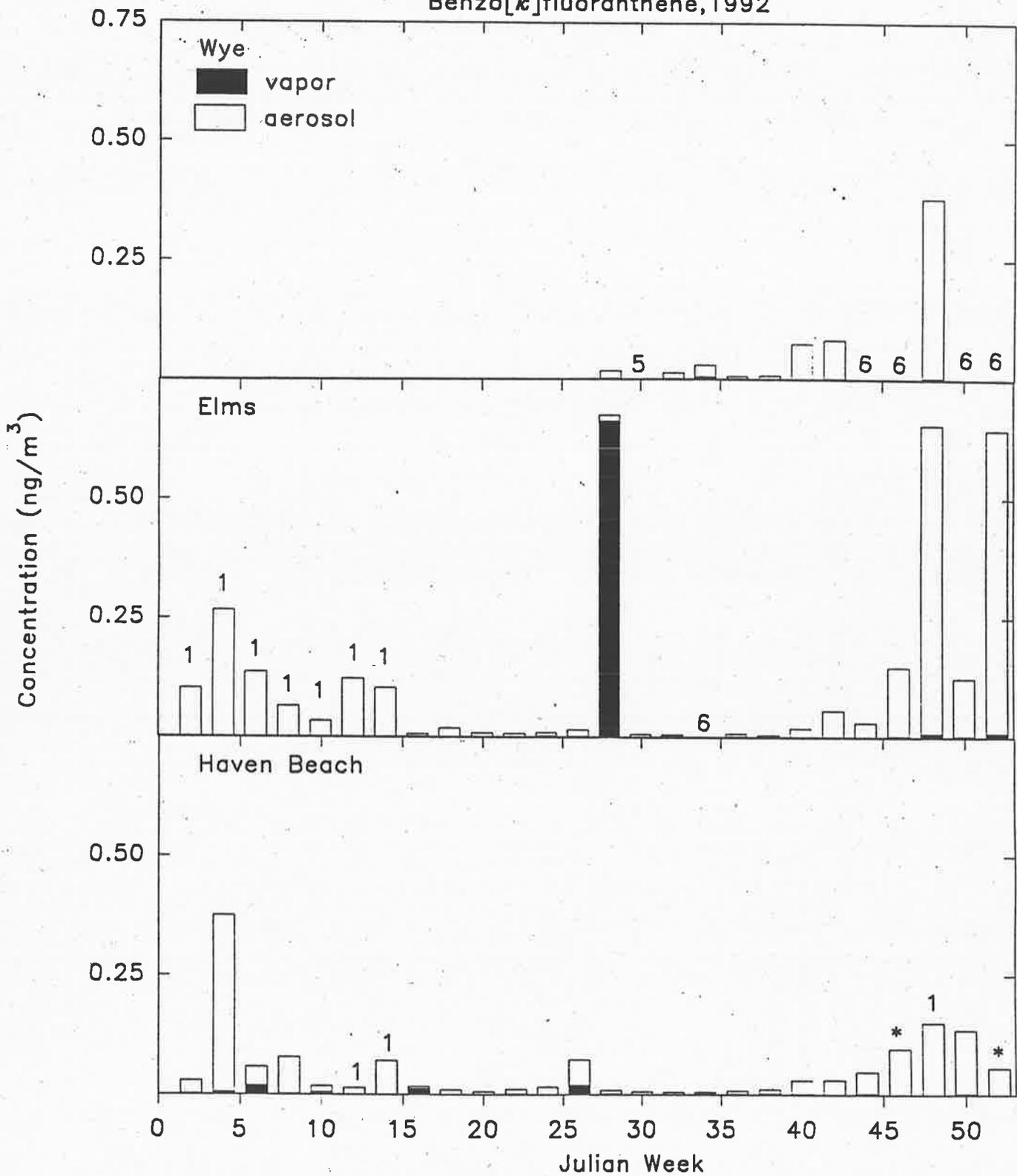


A1.8 Benzo[b]fluoranthene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benzo[b]fluoranthene, 1993

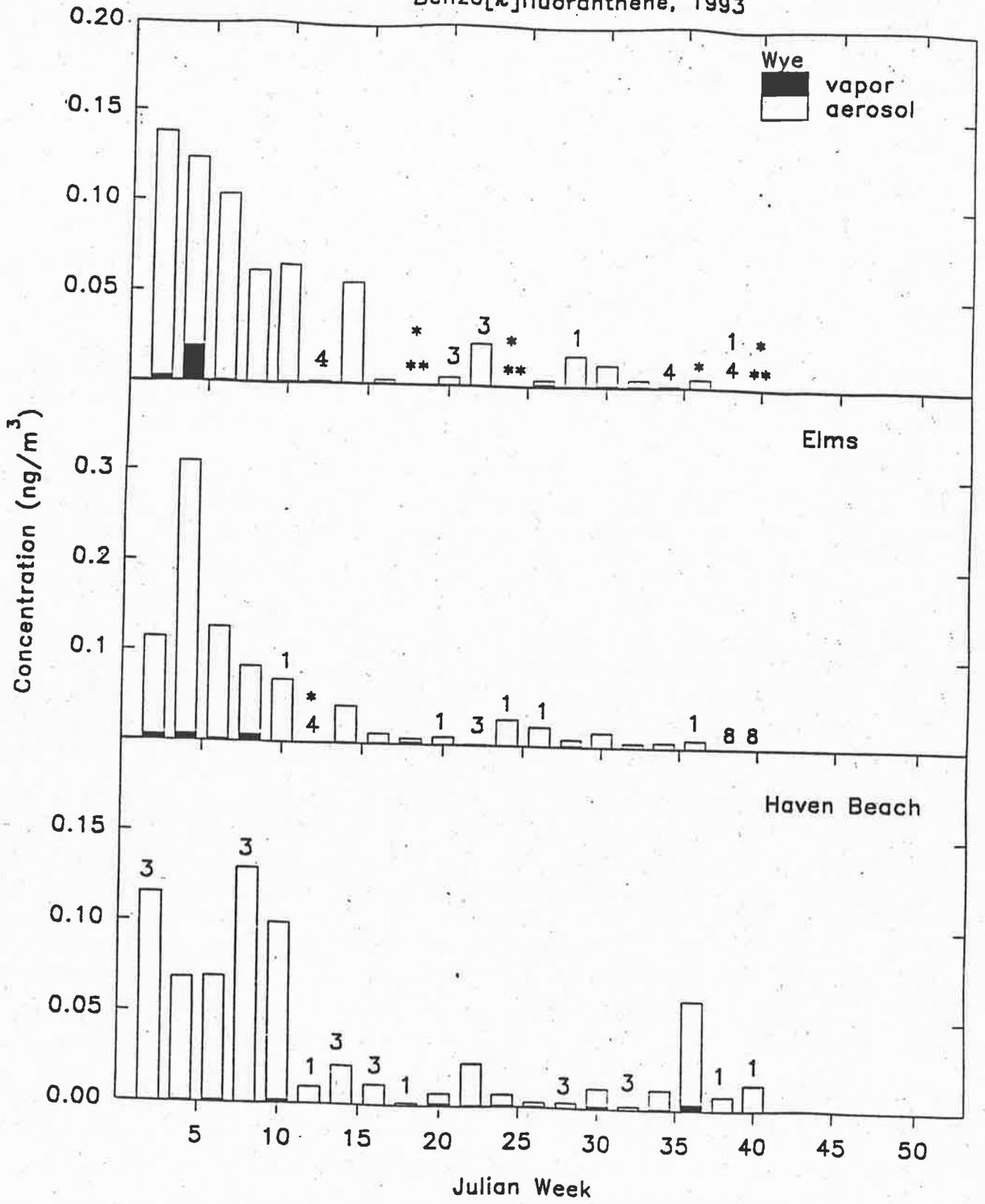


Benzo[k]fluoranthene, 1992

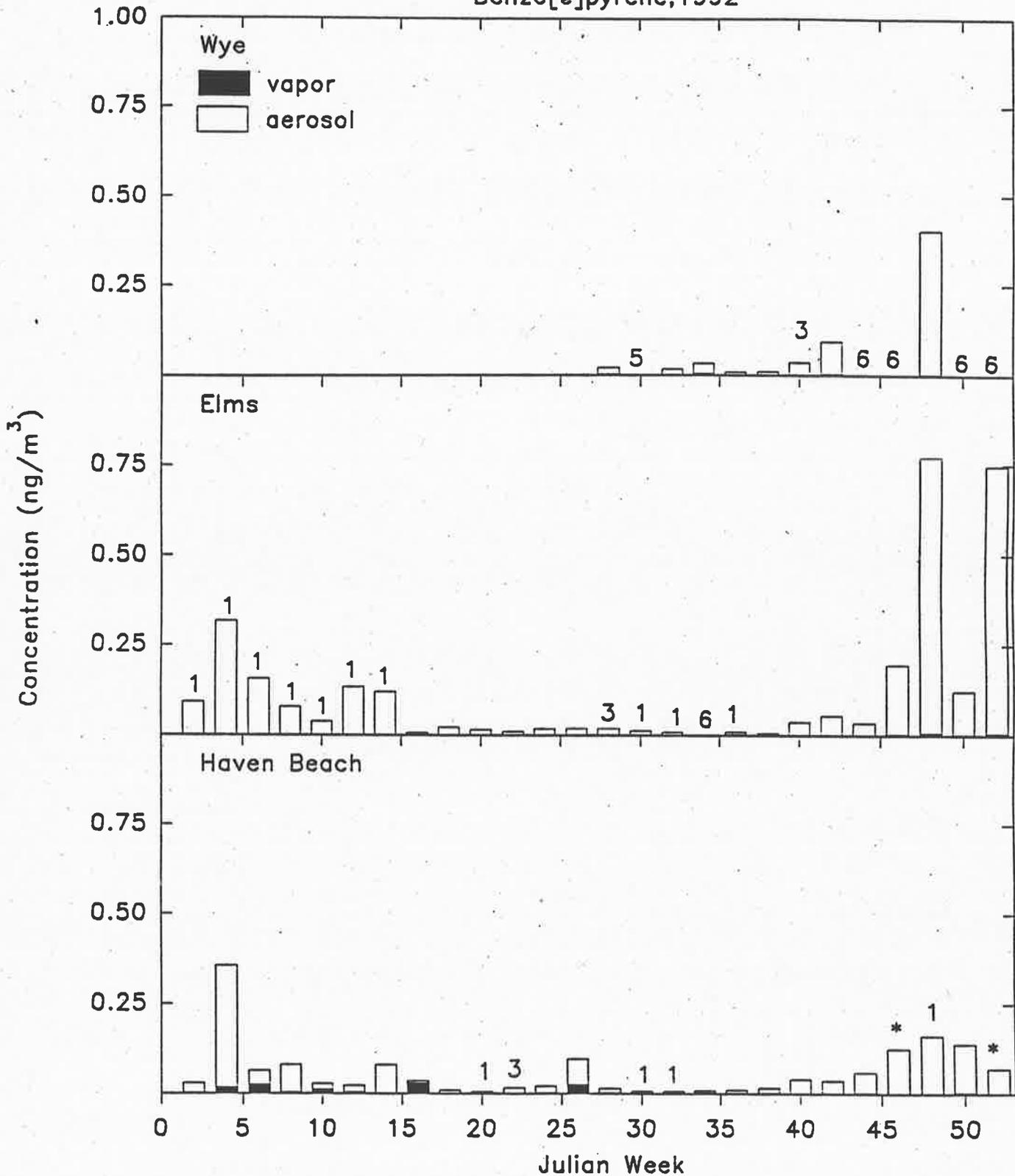


A1.9 Benzo[k]fluoranthene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benzo[k]fluoranthene, 1993

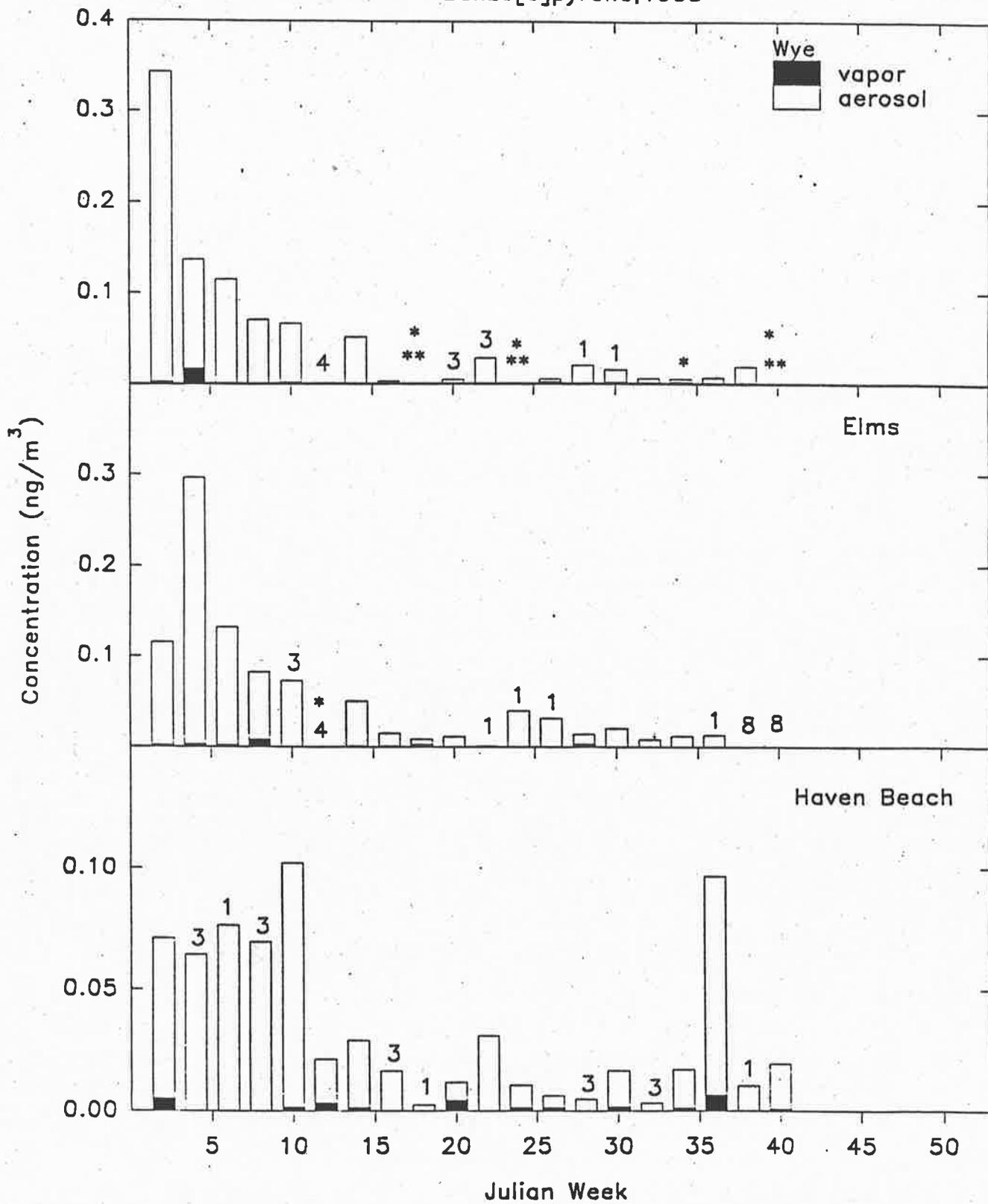


Benzo[e]pyrene, 1992

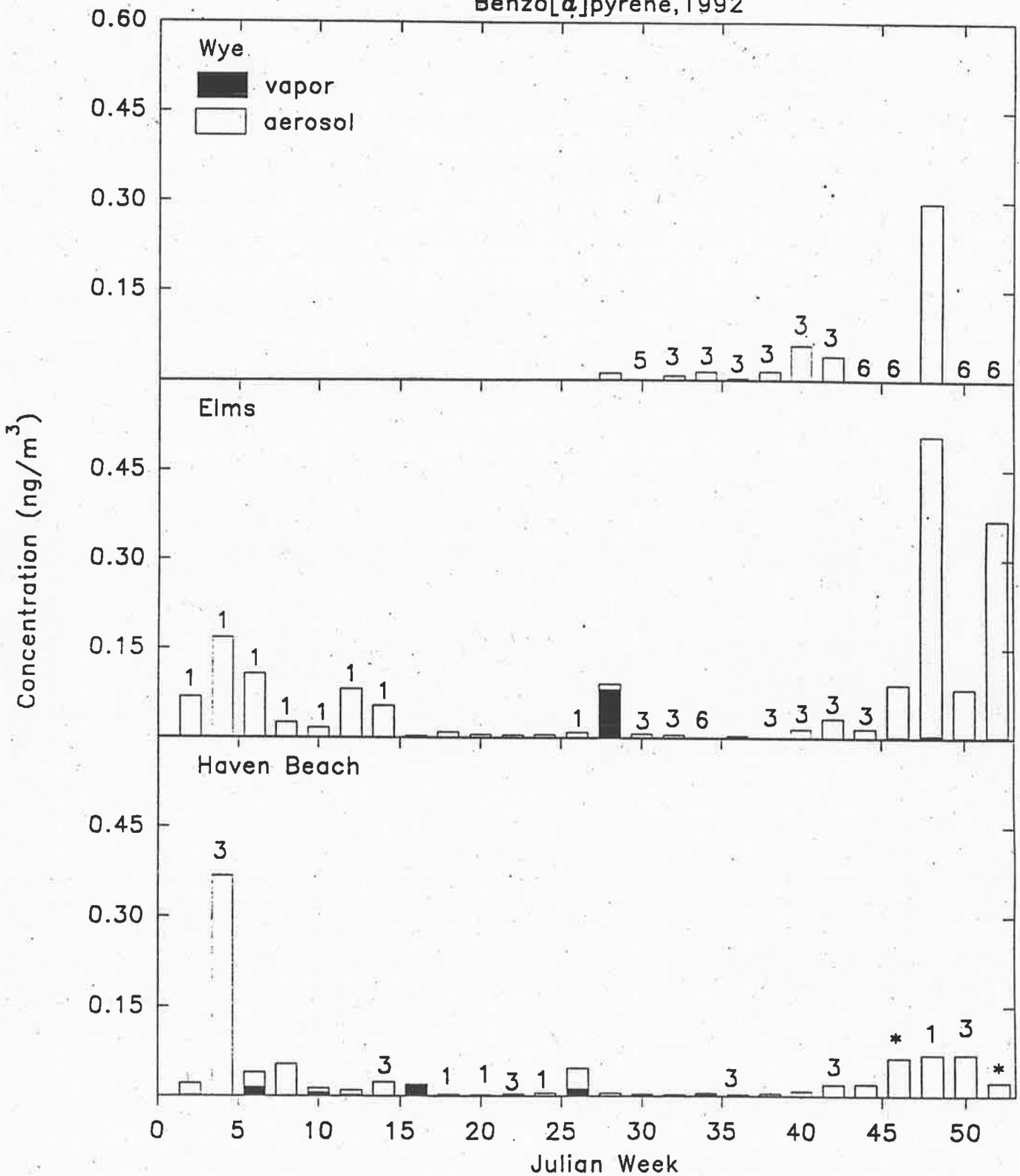


A1.10 Benzo[e]pyrene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benzo[e]pyrene, 1993

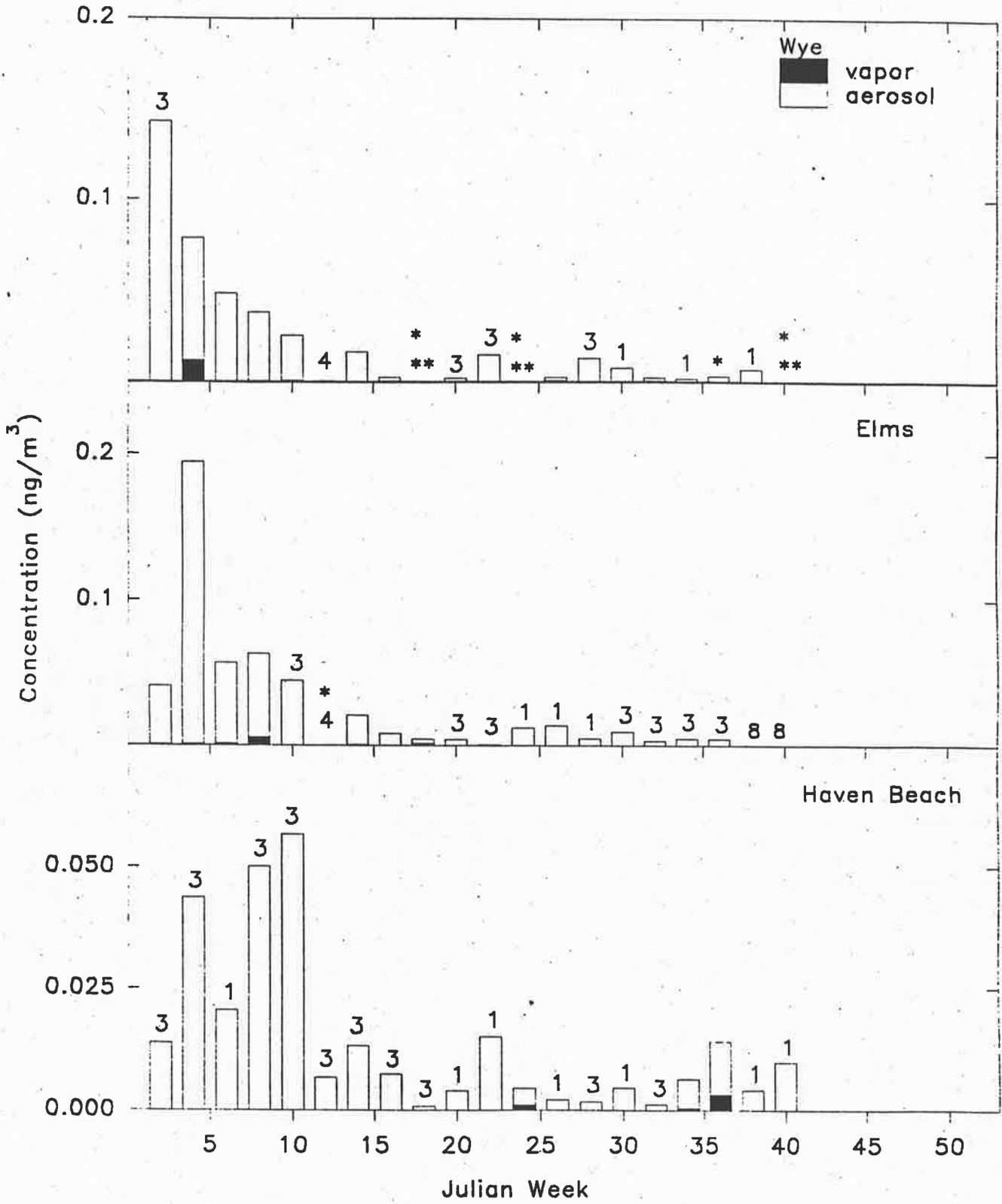


Benzo[*a*]pyrene, 1992

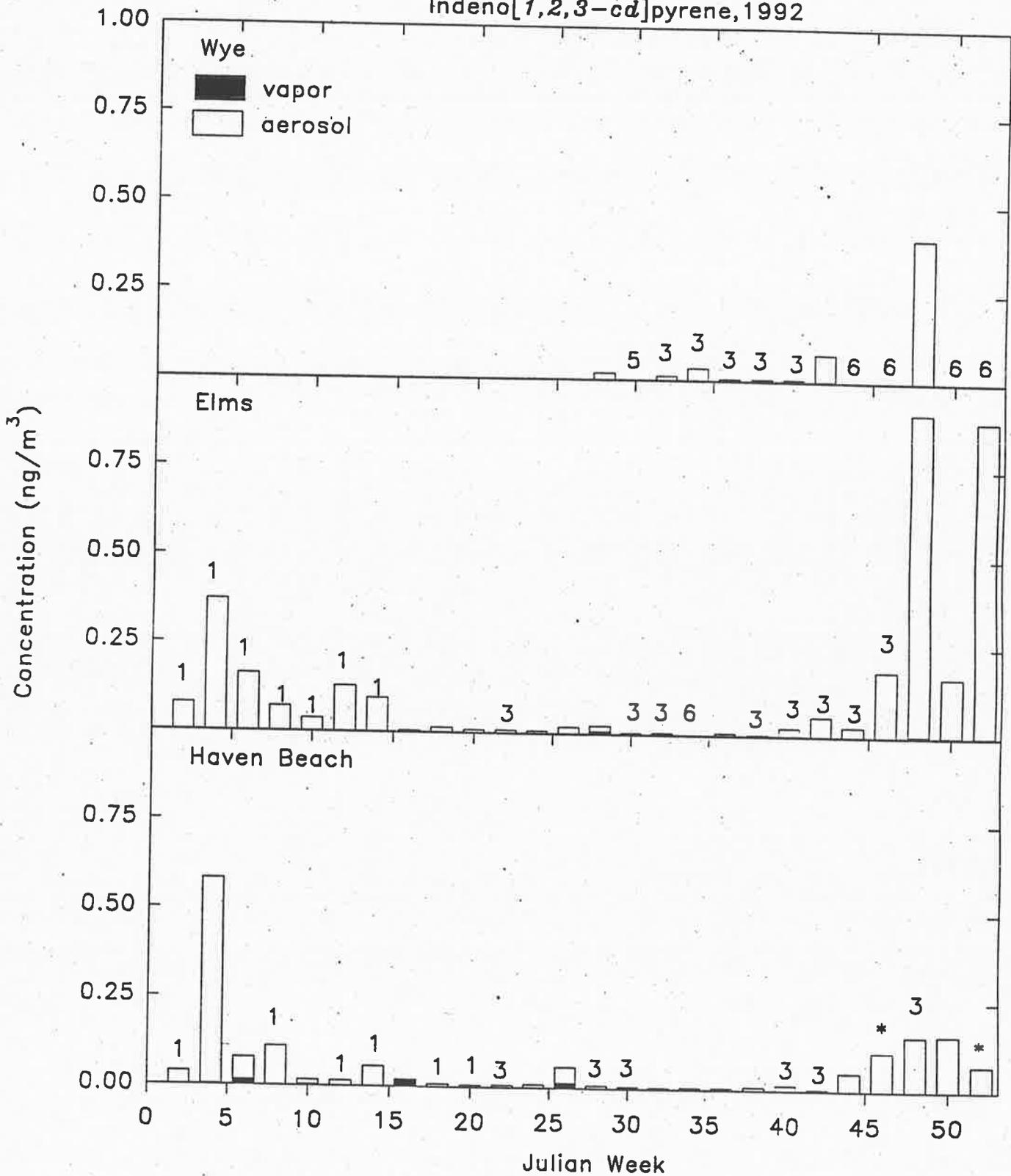


A1.11 Benzo[*a*]pyrene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benzo[a]pyrene, 1993

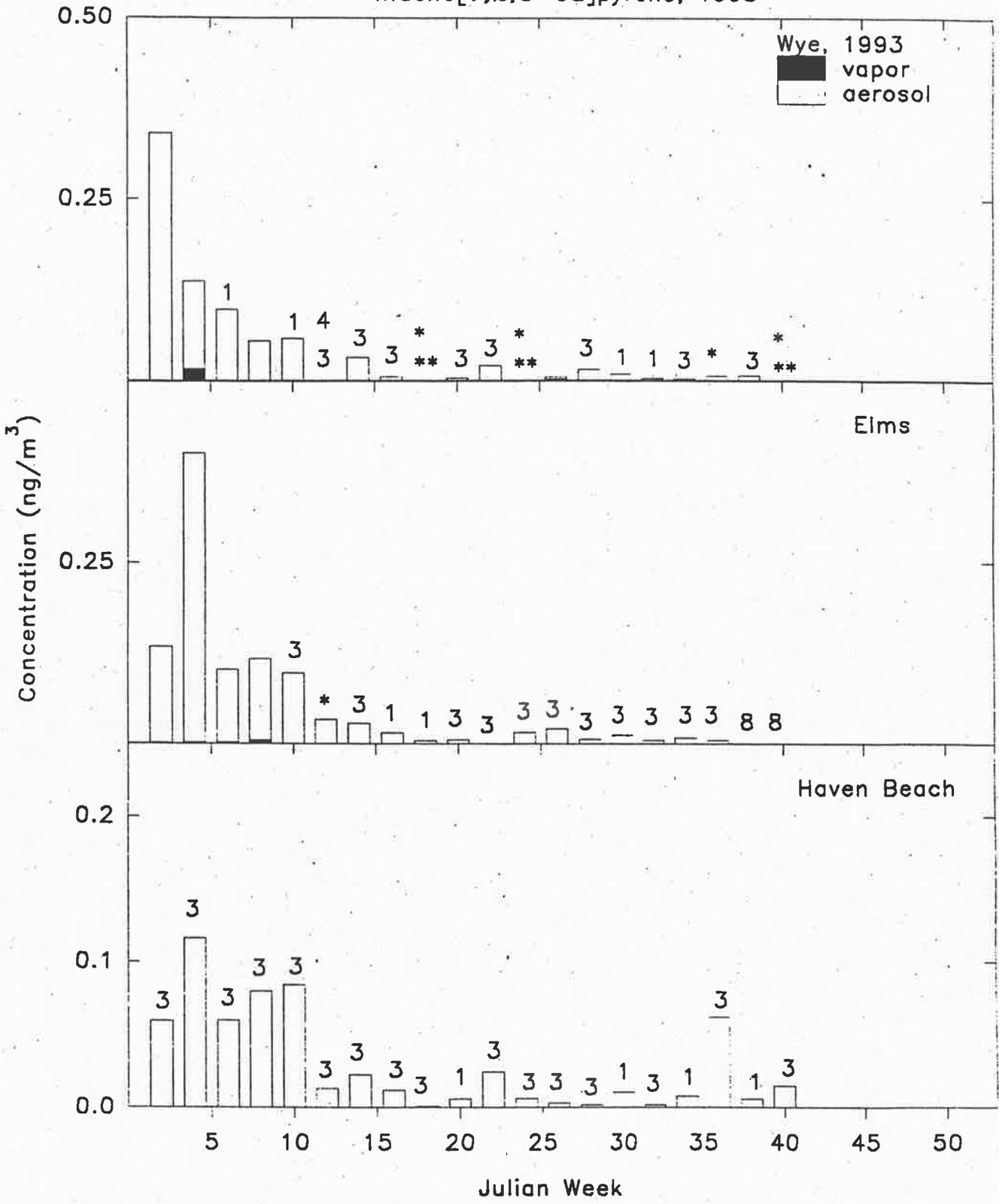


Indeno[1,2,3-cd]pyrene, 1992

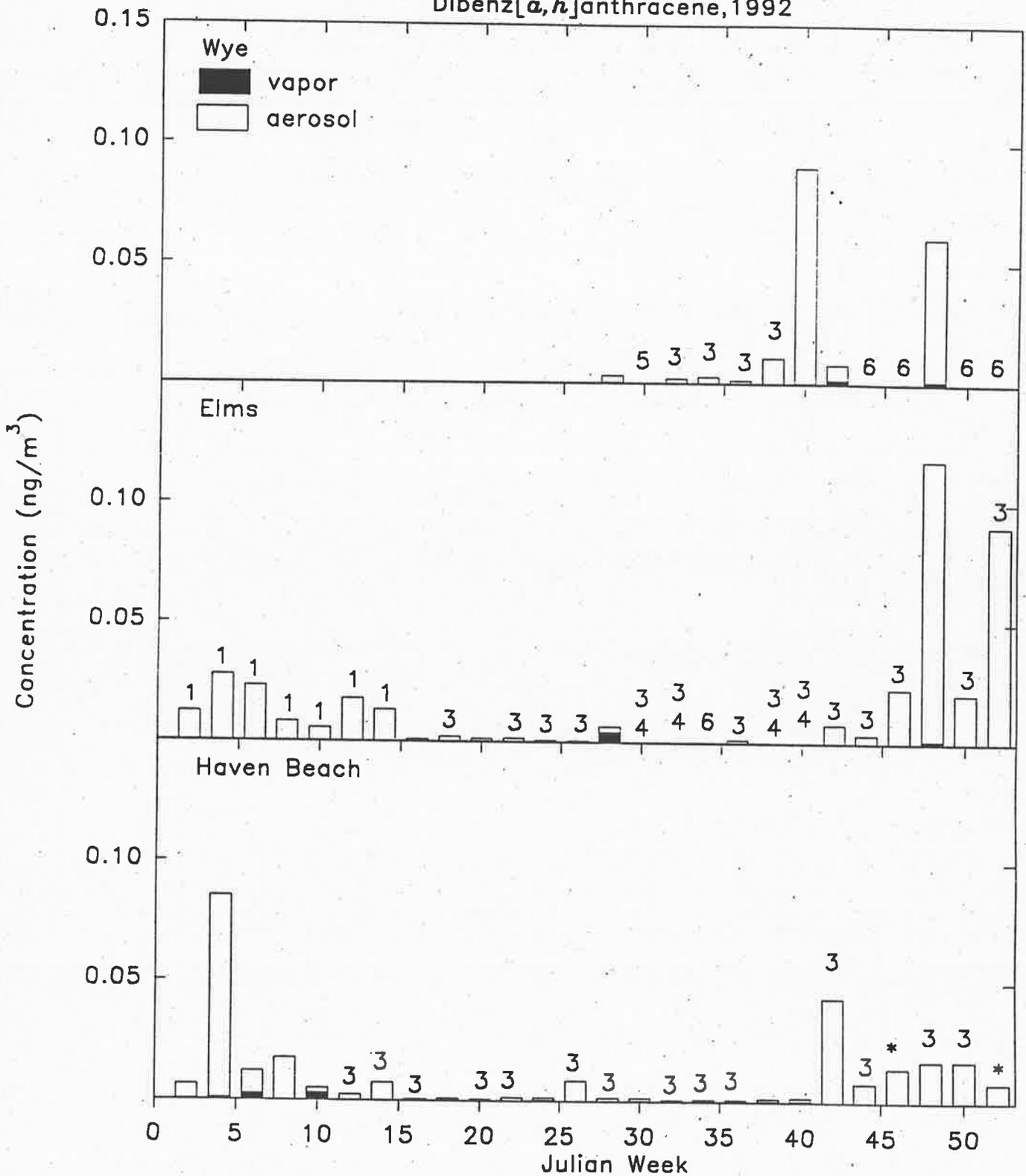


A1.12 Indeno[1,2,3-cd]pyrene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (.1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Indeno[1,2,3-cd]pyrene, 1993

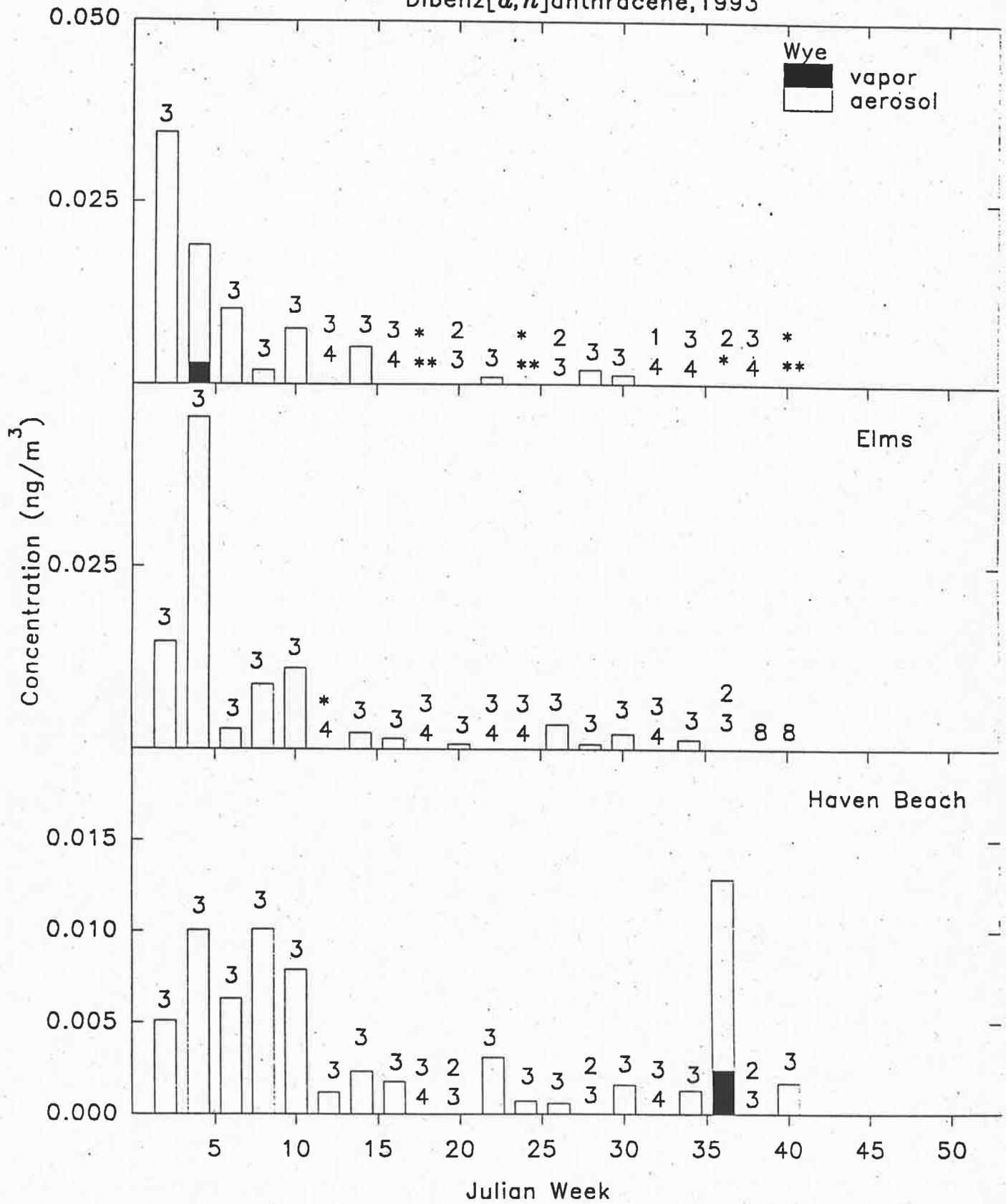


Dibenz[*a,h*]anthracene, 1992

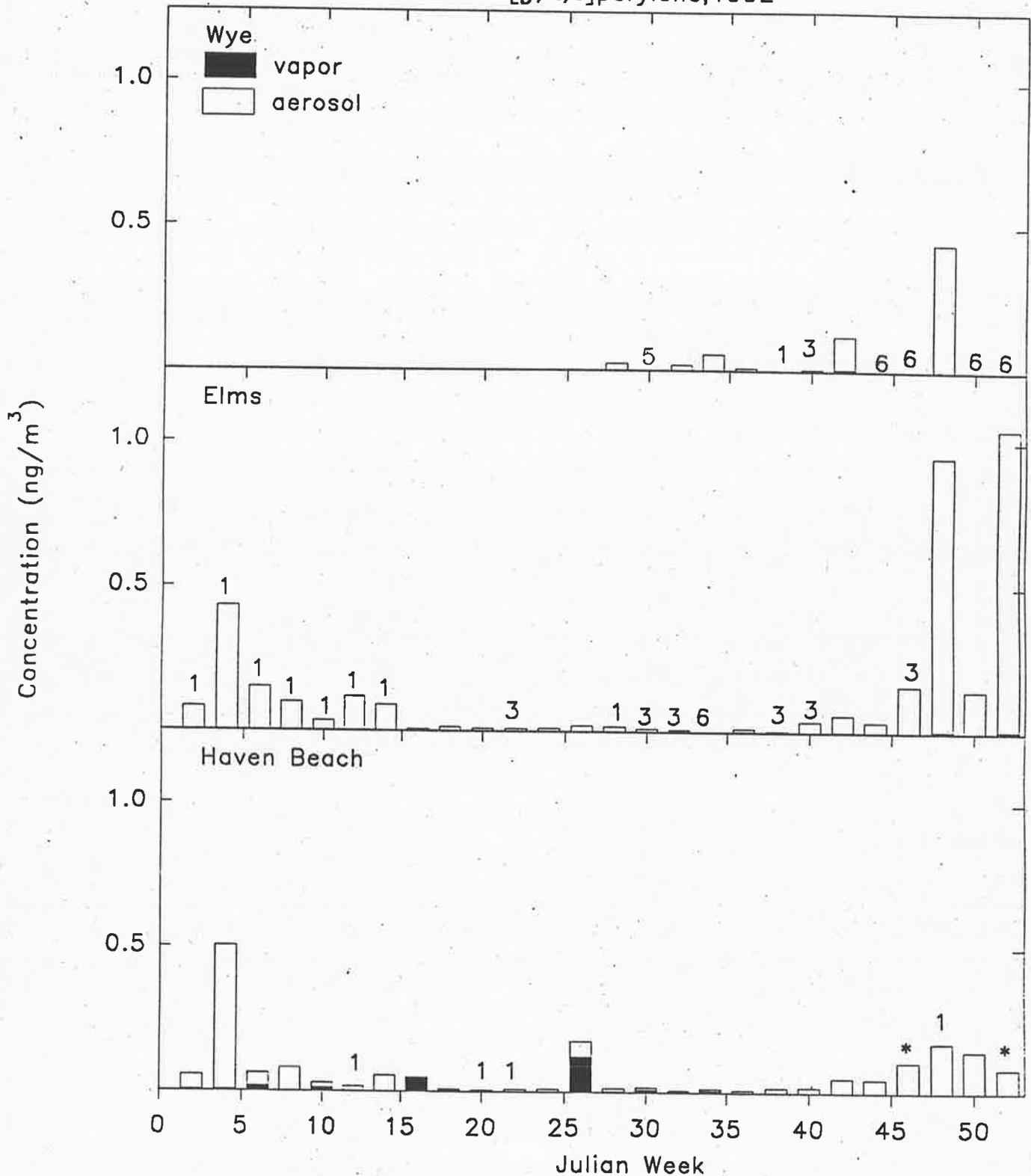


A1.13 Dibenz[*a,h*]anthracene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Dibenz[*a,h*]anthracene, 1993

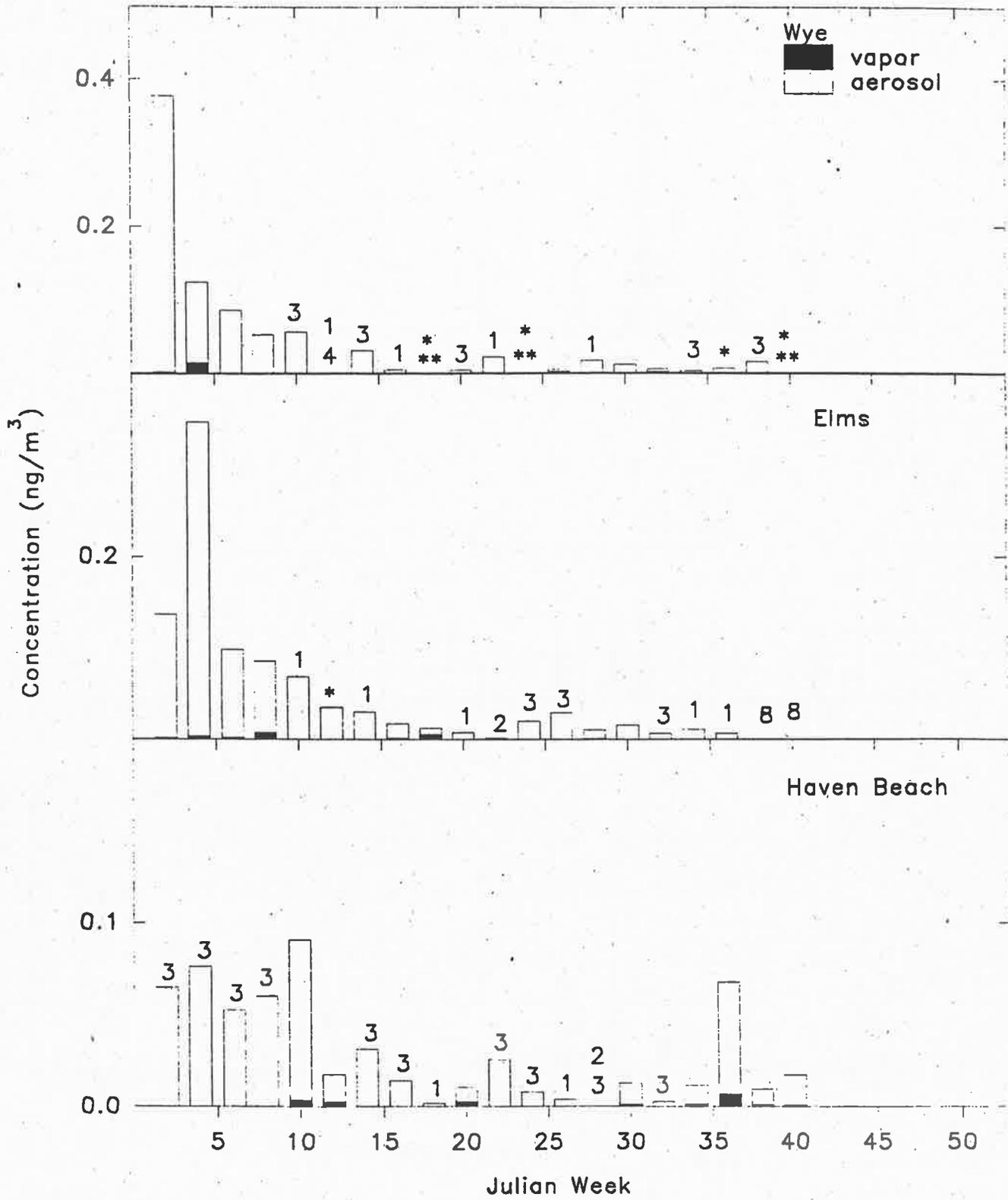


Benzo[*g,h,i*]perylene, 1992

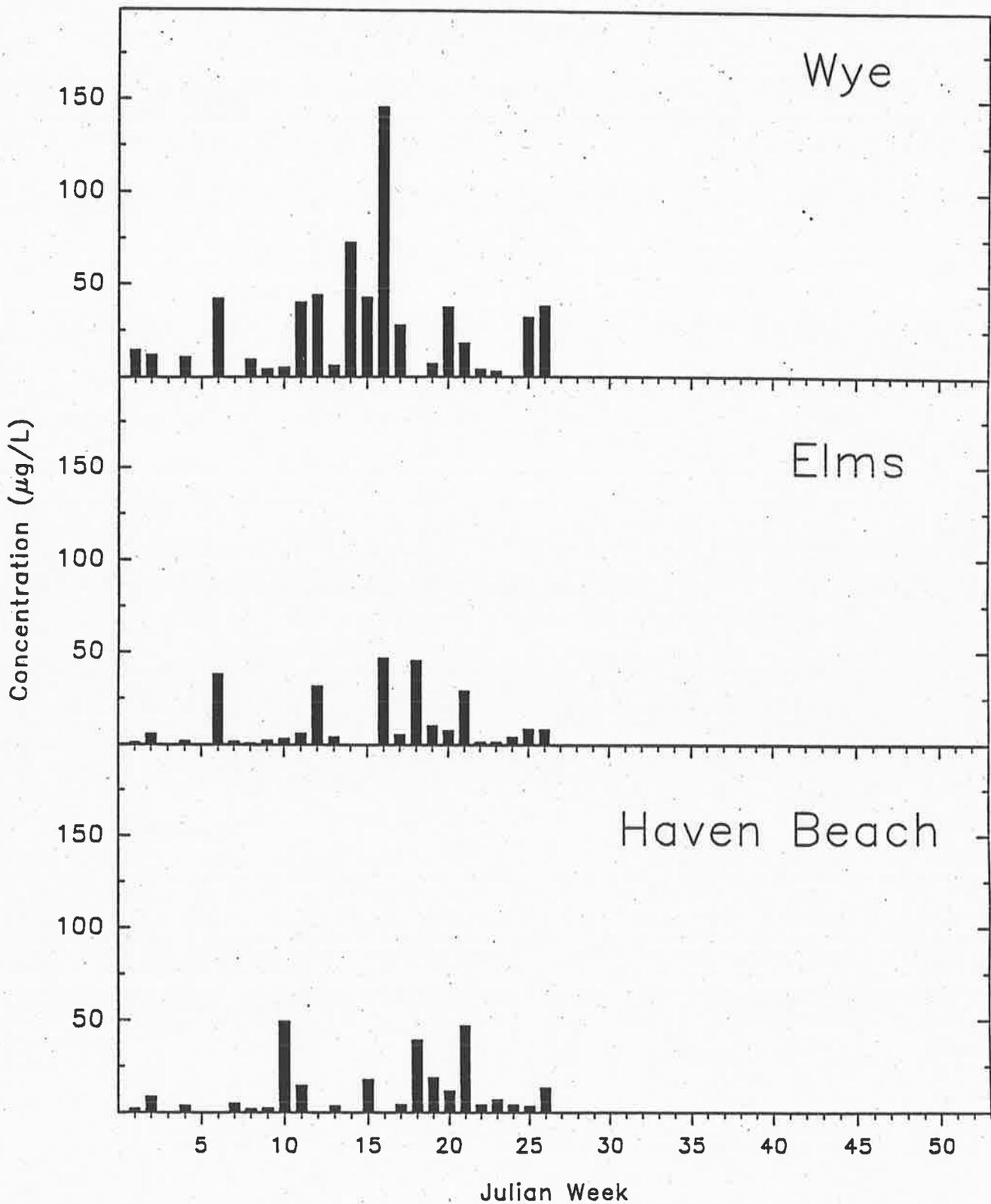


A1.14 Benzo[*g,h,i*]perylene concentrations collected biweekly in air at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in vapor phase, 2 = not quantifiable on filter, 3=not detected in vapor phase, 4 = not detected on filter, 5 =field blank, 6=no sample taken, *=lost vapor sample **= lost particulate sample).

Benzo[*g,h,i*]perylene, 1993

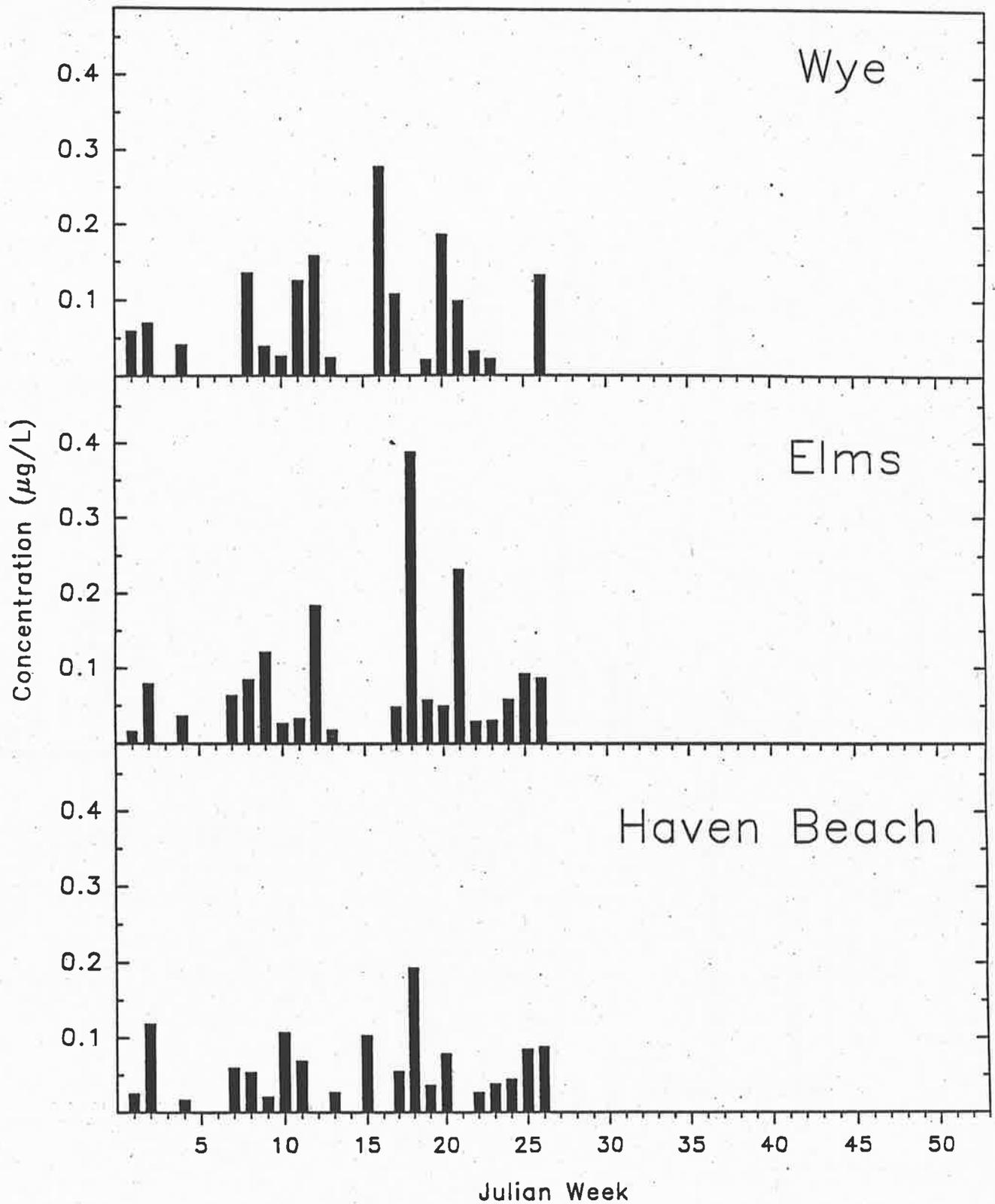


Aluminum, 1992



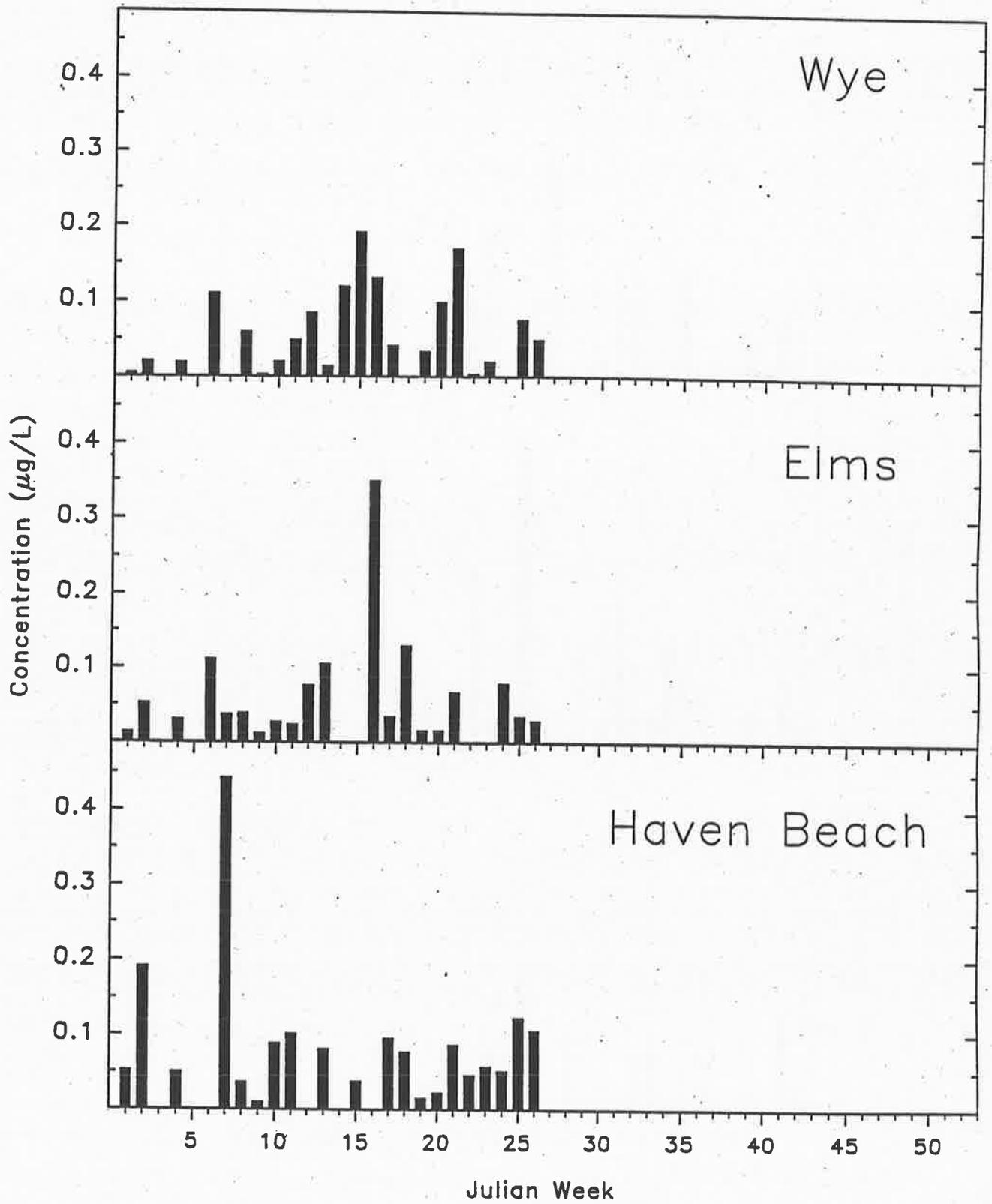
A1.15 Aluminum concentrations in precipitation, integrated at each CBAD site.

Arsenic, 1992



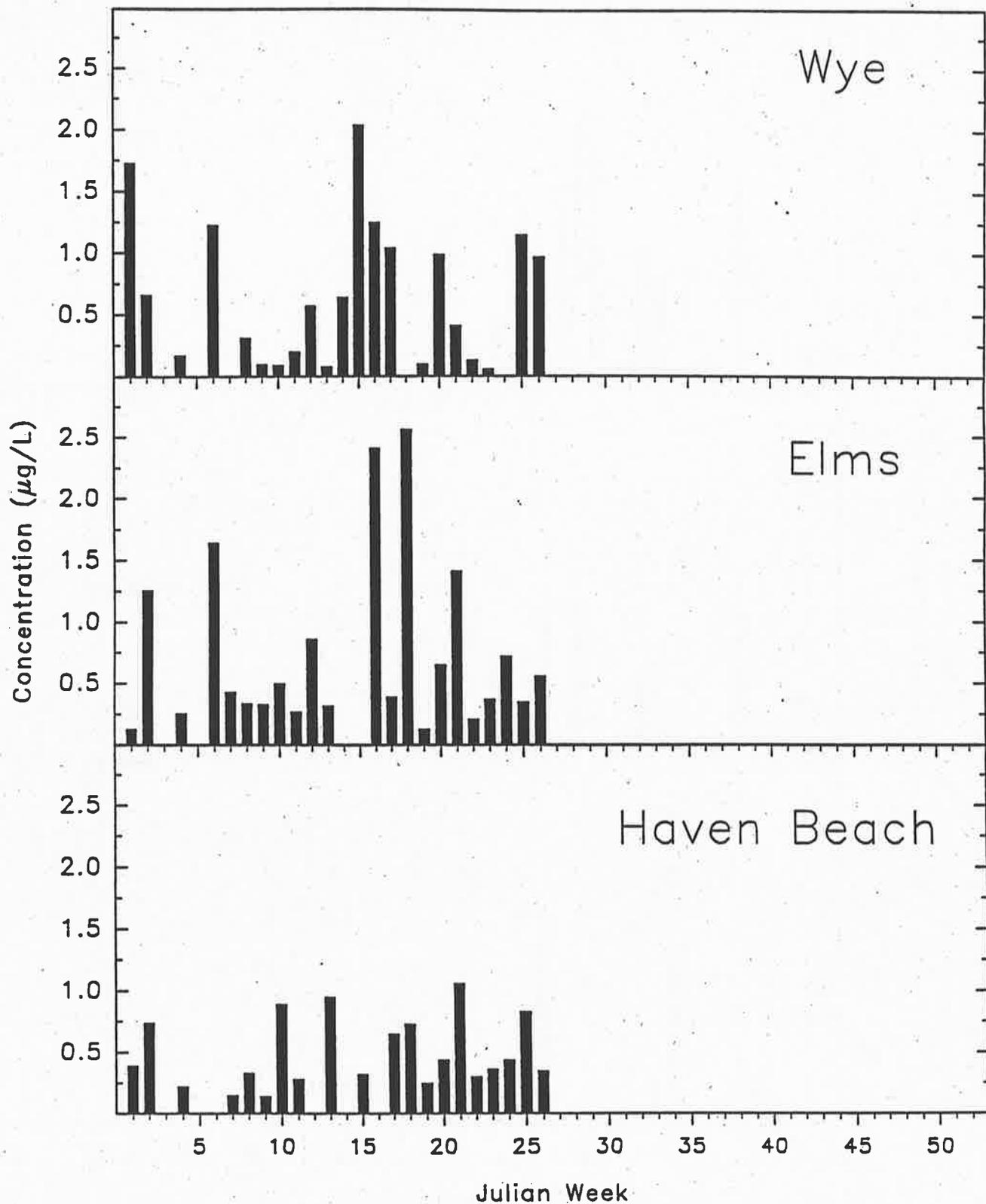
A1.16 Arsenic concentrations in precipitation, integrated at each CBAD site.

Cadmium, 1992



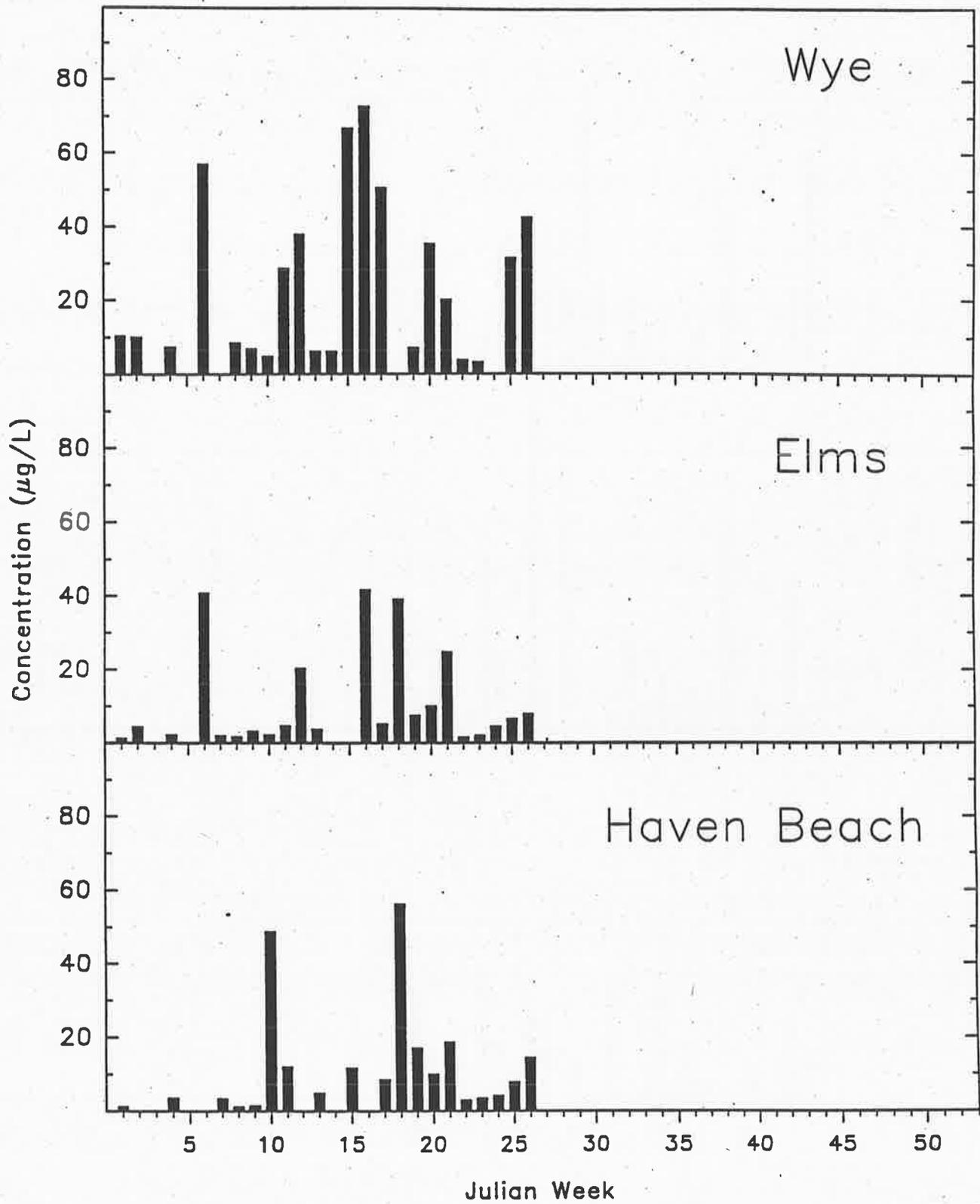
A1.17 Cadmium concentrations in precipitation, integrated at each CBAD site.

Copper, 1992



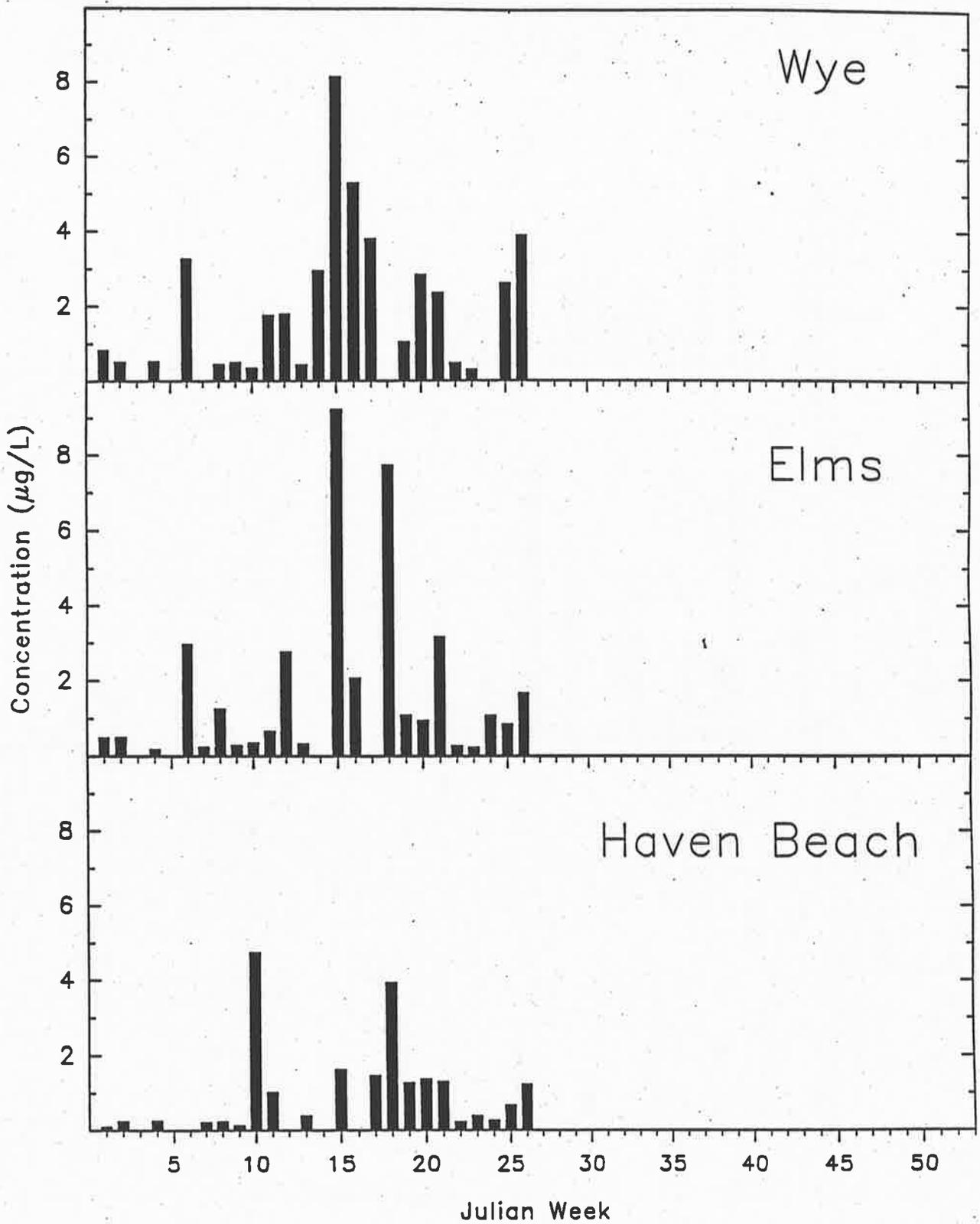
A1.18 Copper concentrations in precipitation, integrated at each CBAD site.

Iron, 1992



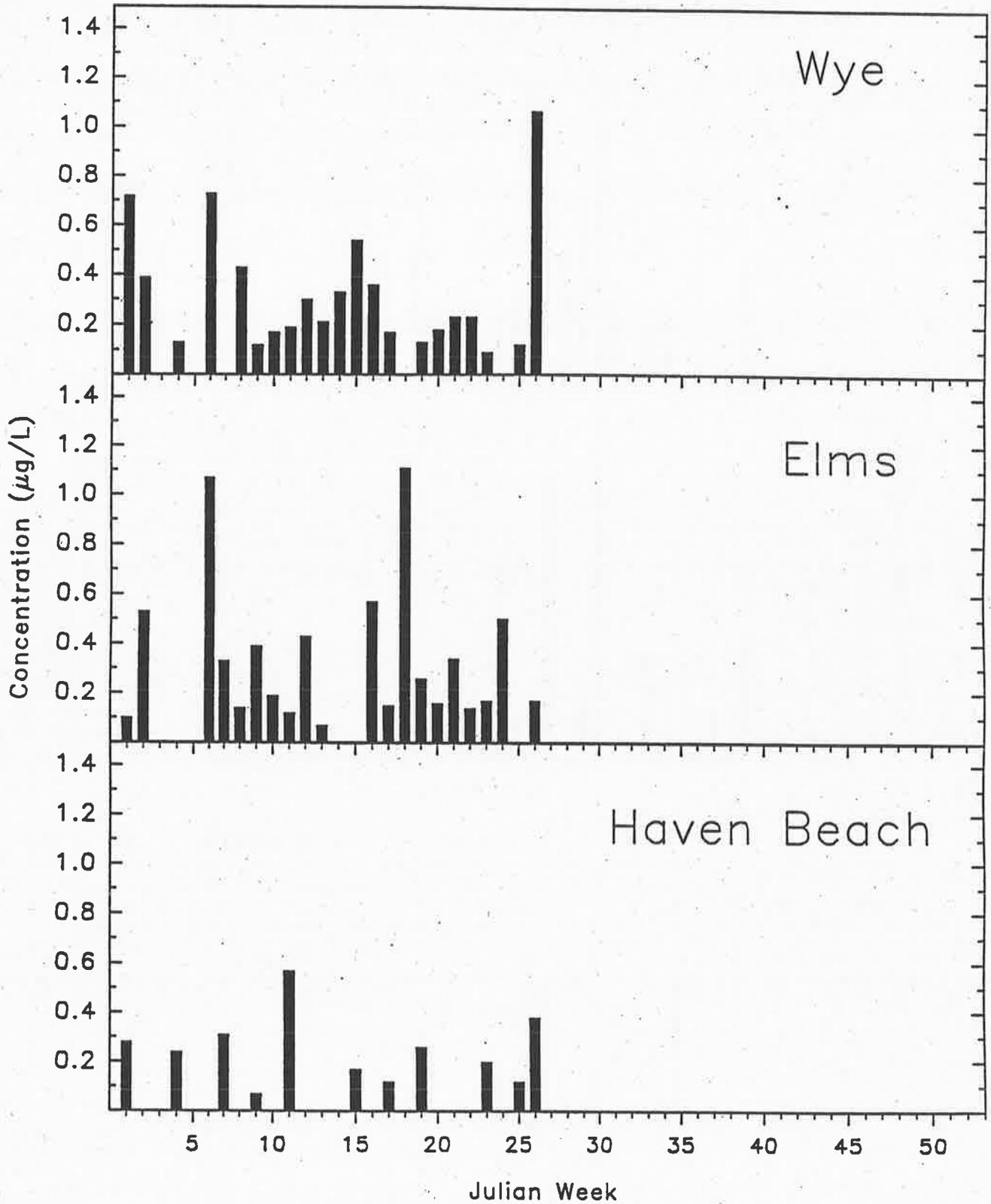
A1.19 Iron concentrations in precipitation, integrated at each CBAD site.

Manganese, 1992



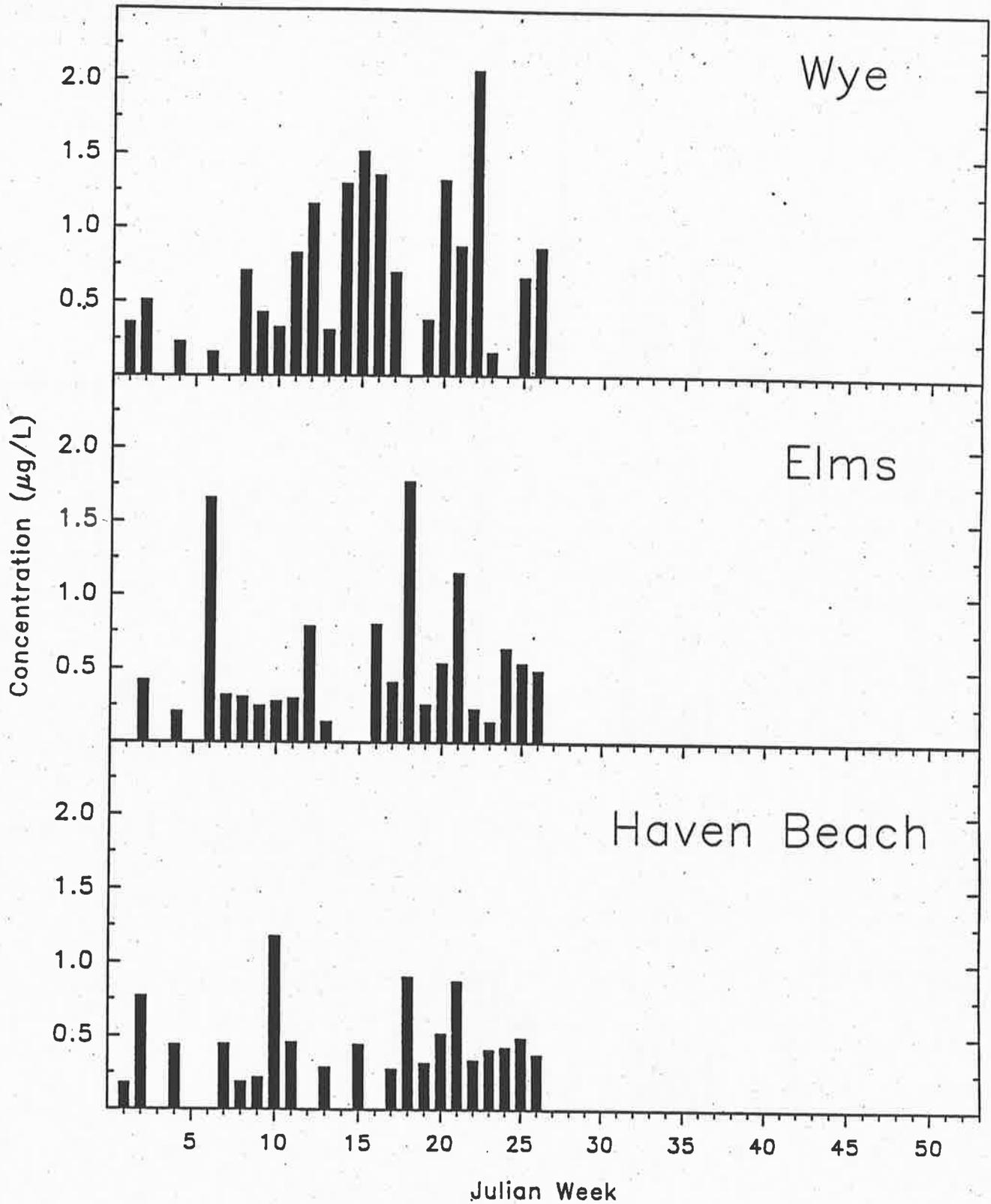
A1.20 Manganese concentrations in precipitation, integrated at each CBAD site.

Nickel, 1992



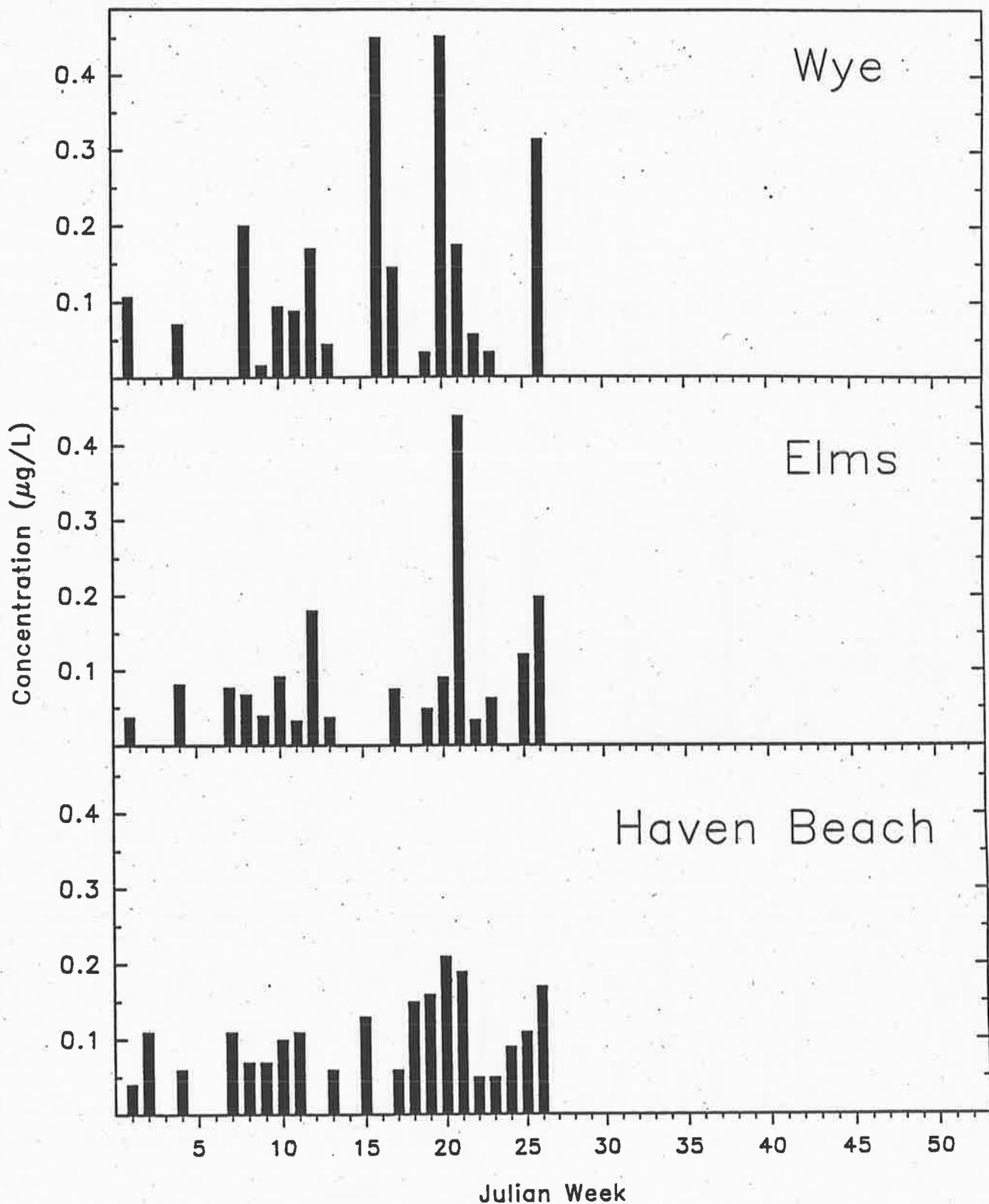
A1.21 Nickel concentrations in precipitation, integrated at each CBAD site.

Lead, 1992



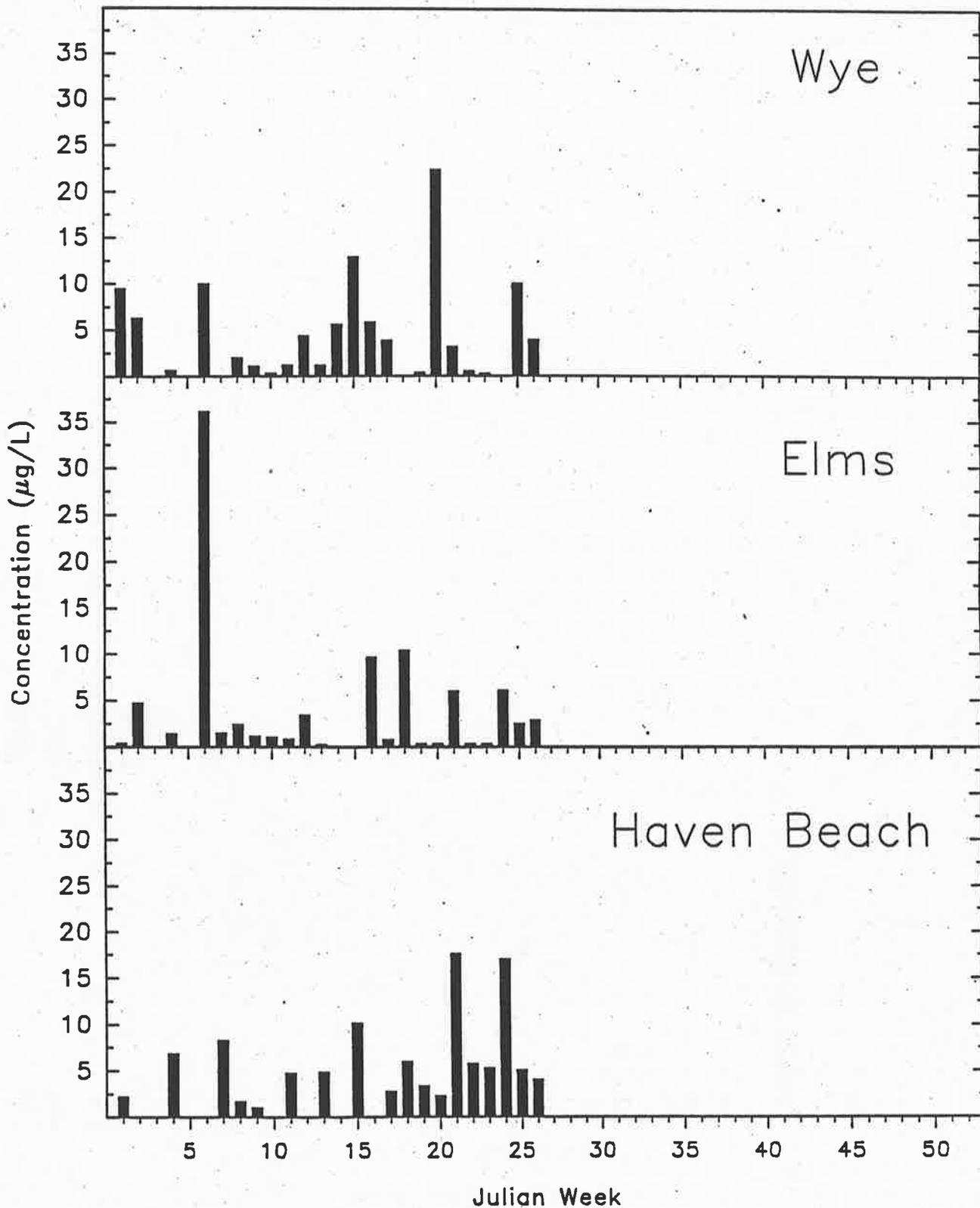
A1.22 Lead concentrations in precipitation, integrated at each CBAD site.

Selenium, 1992



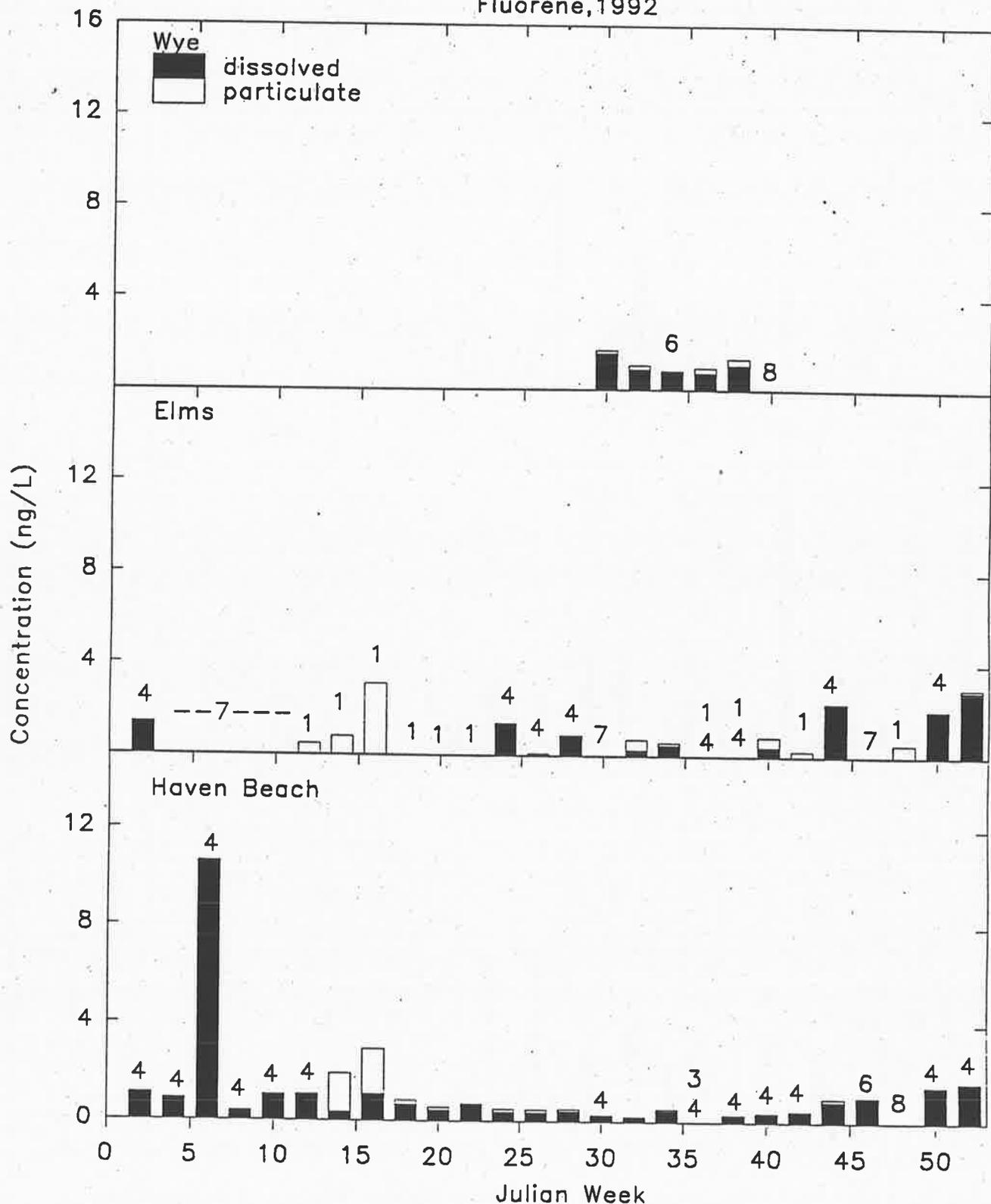
A1.23 Selenium concentrations in precipitation, integrated at each CBAD site.

Zinc, 1992

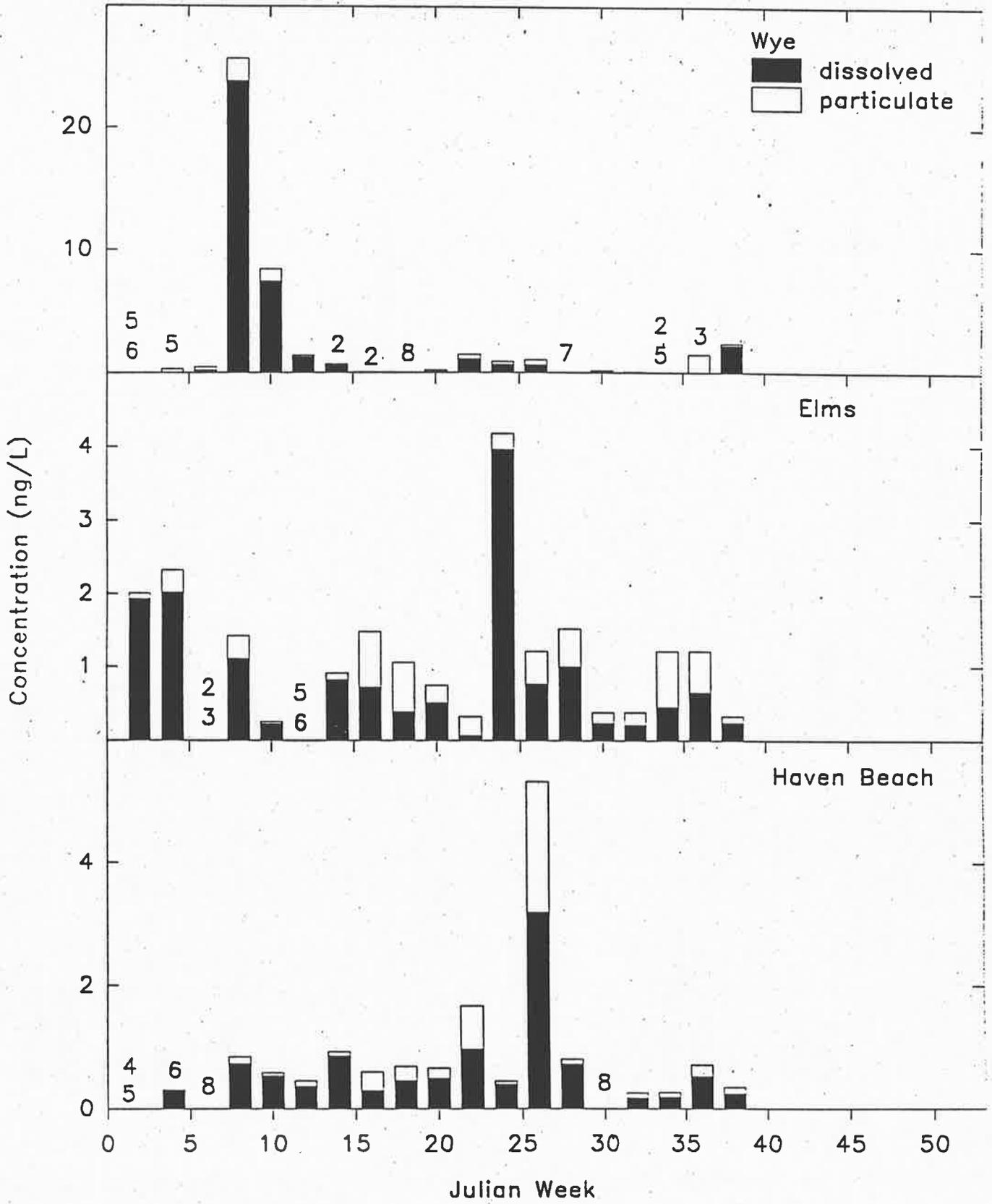


A1.24 Zinc concentrations in precipitation, integrated at each CBAD site.

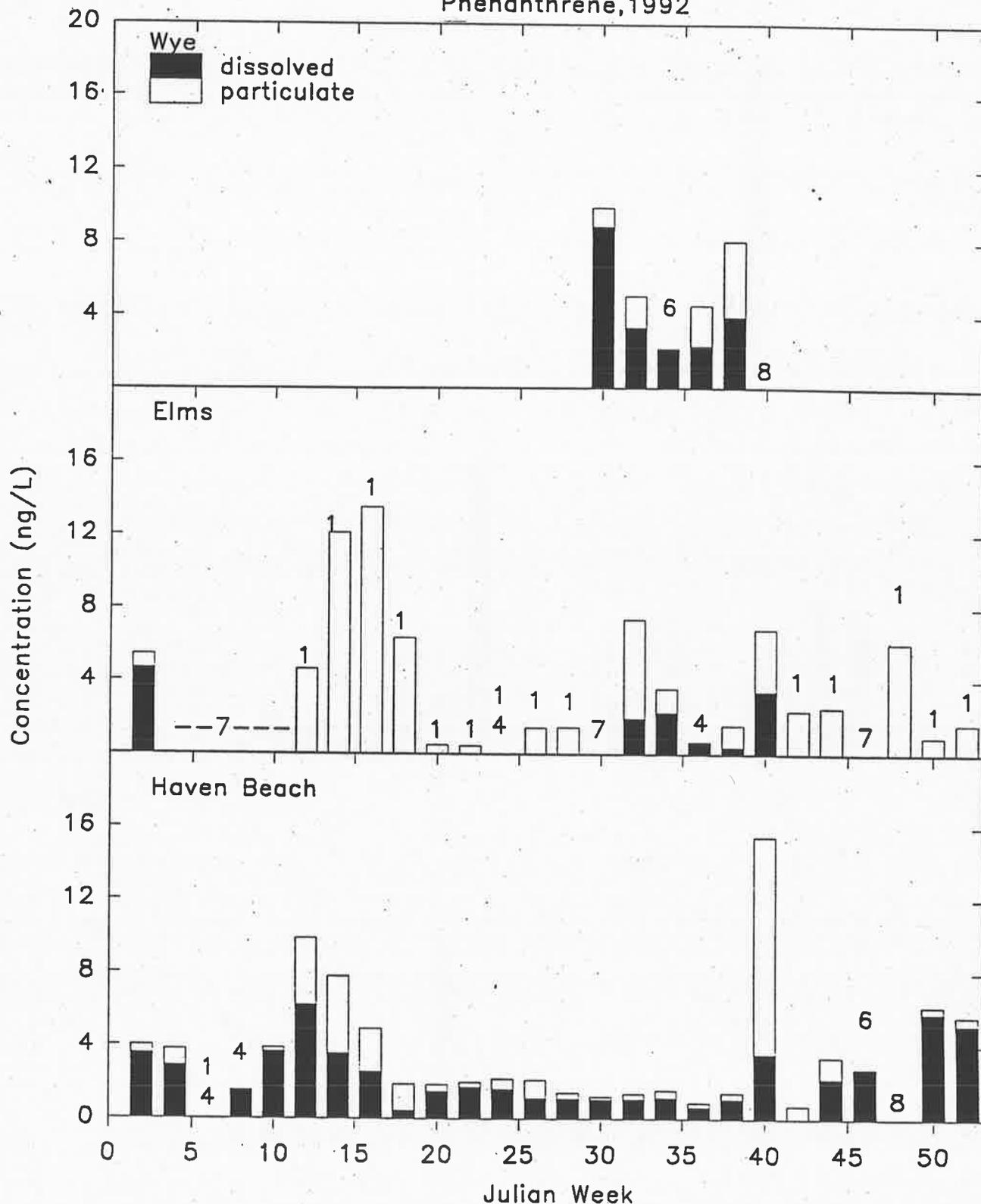
Fluorene, 1992



Fluorene, 1993

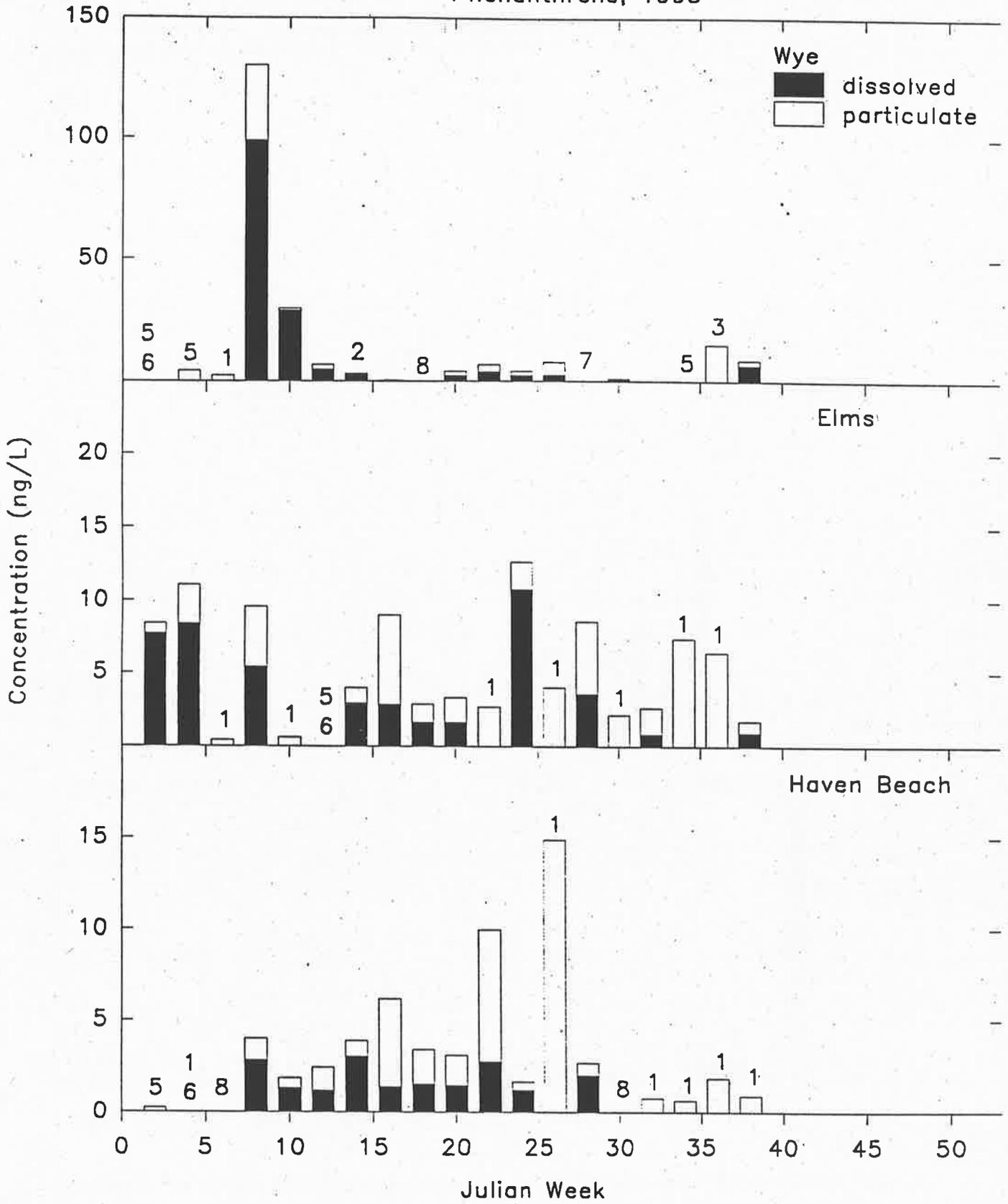


Phenanthrene, 1992

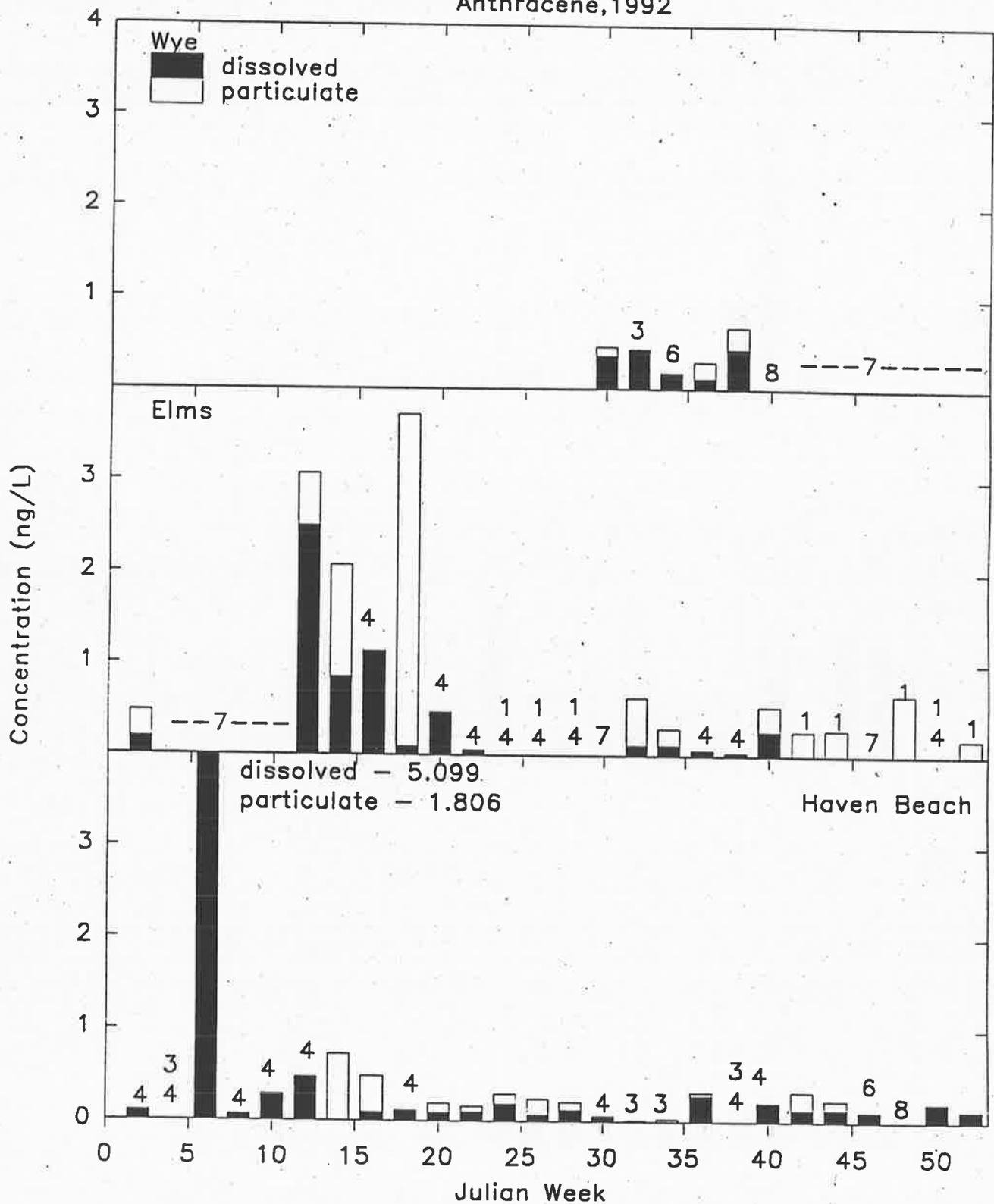


A1.26 Phenanthrene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8=field blank).

Phenanthrene, 1993

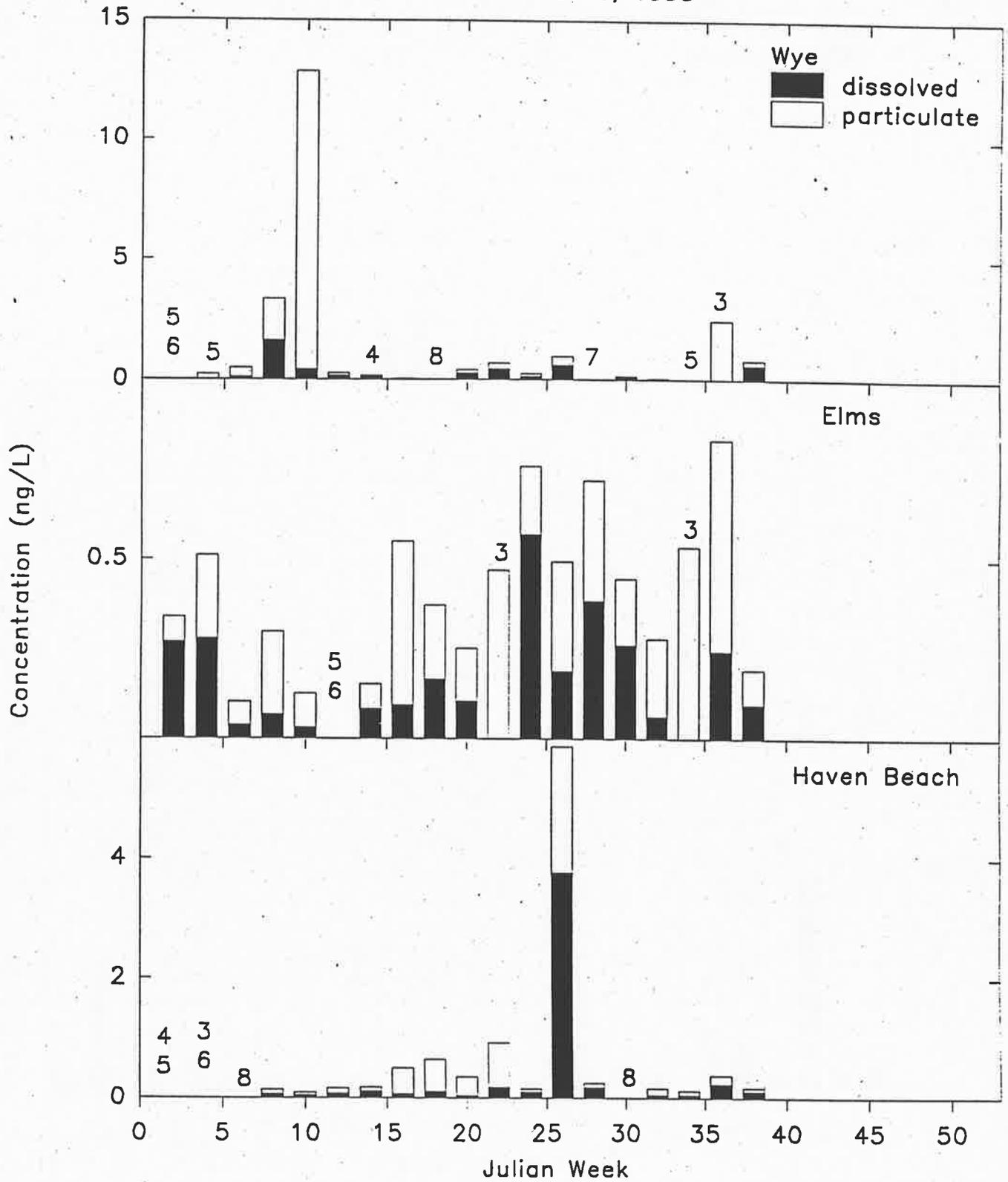


Anthracene, 1992

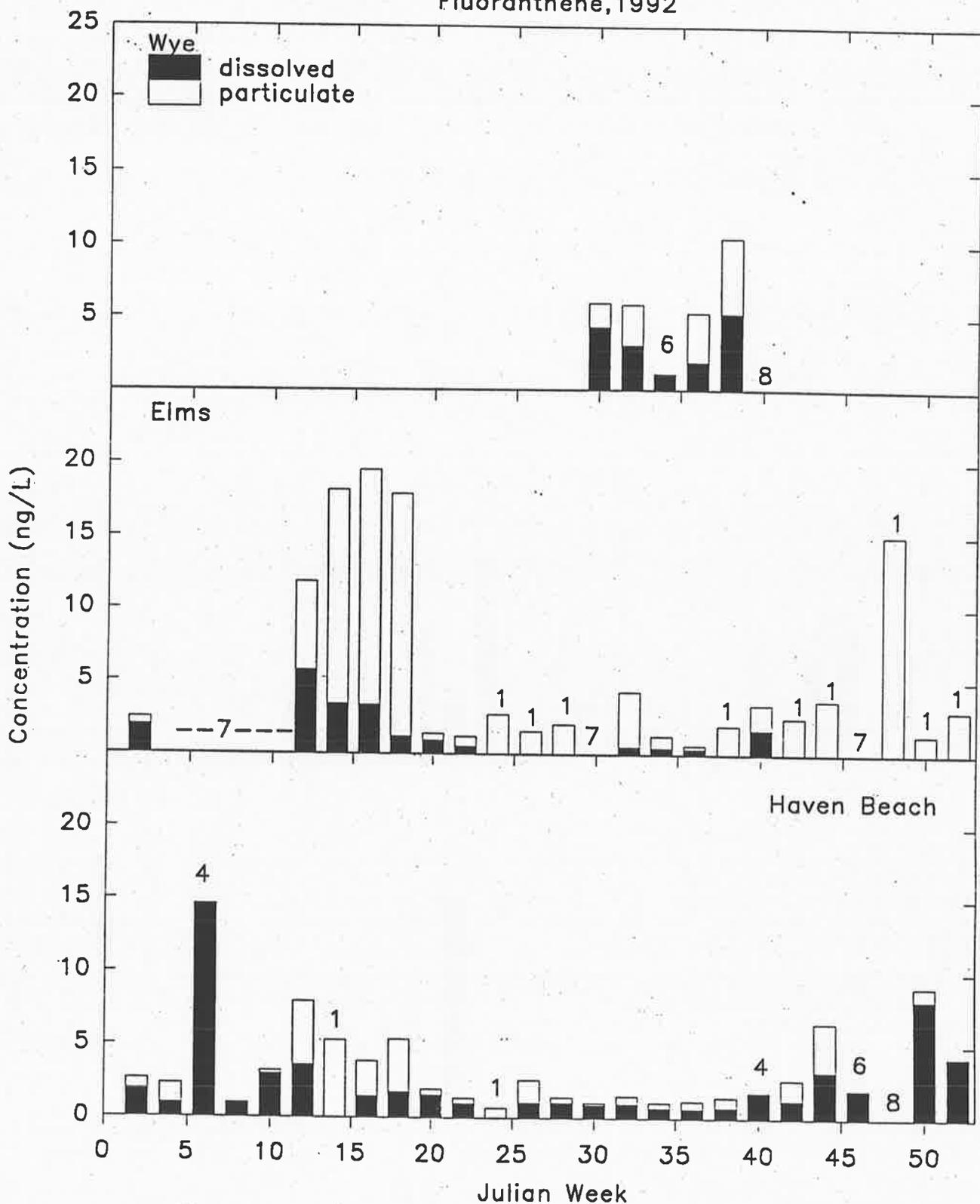


A1.27 Anthracene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Anthracene, 1993

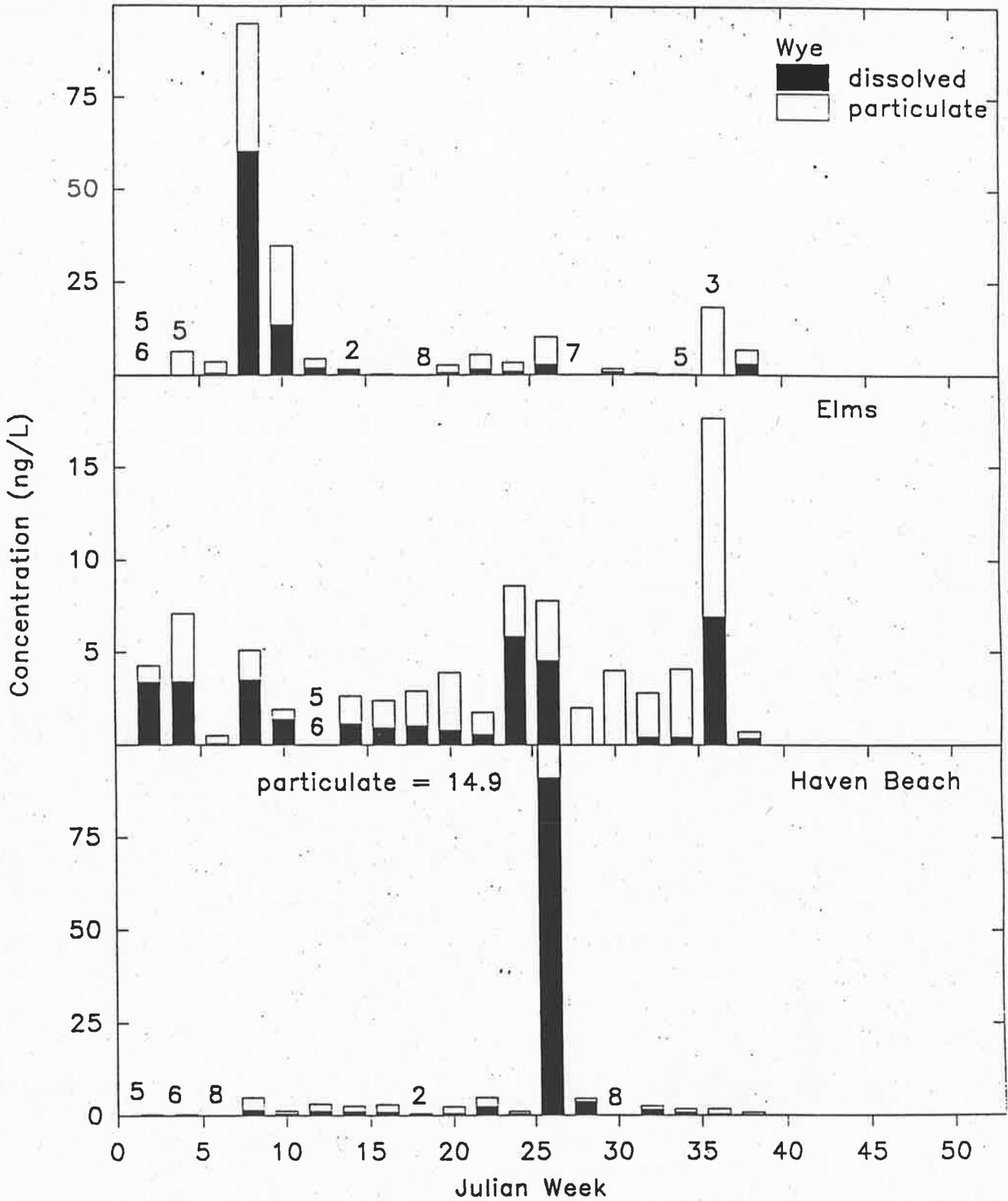


Fluoranthene, 1992

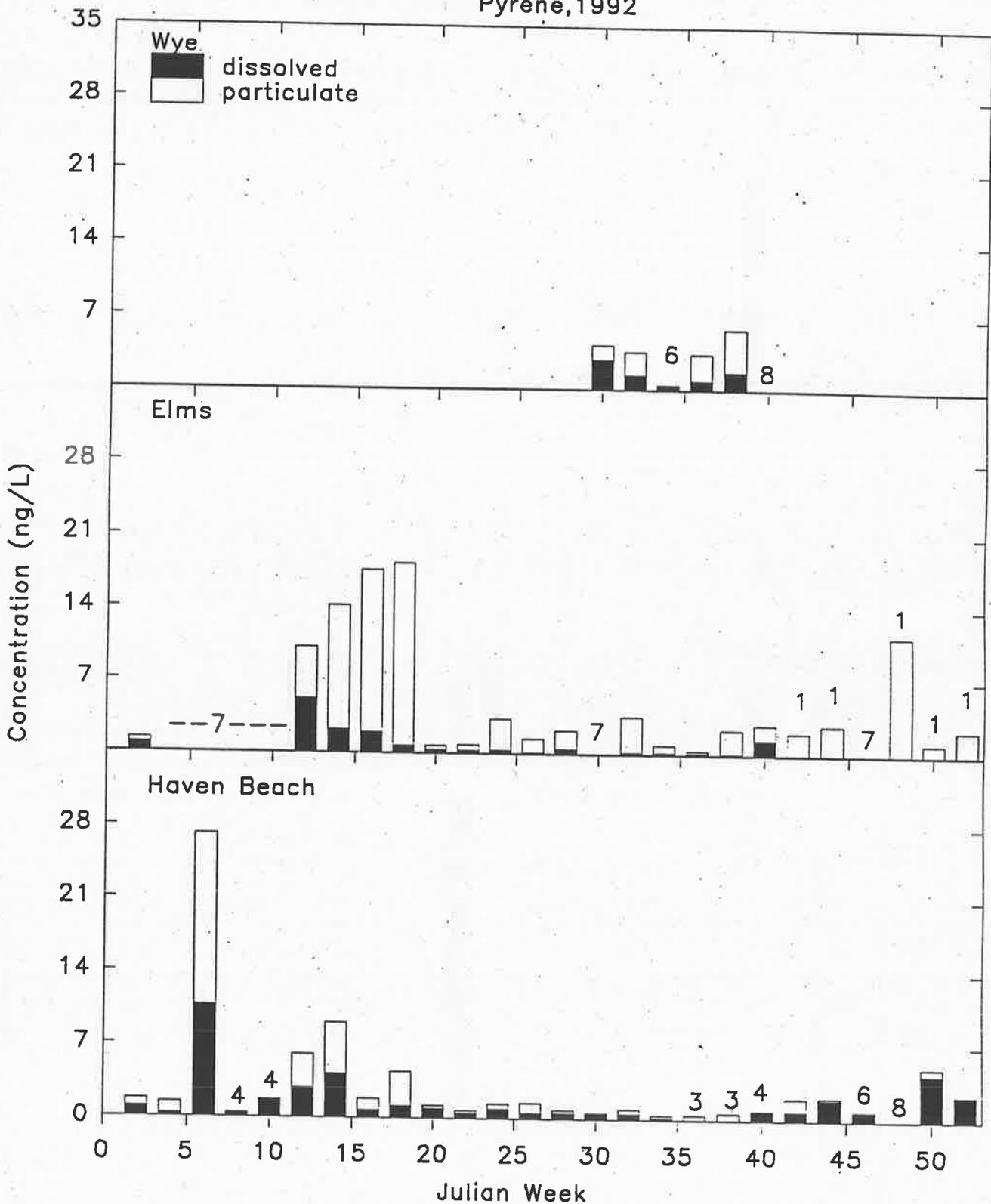


A1.28 Fluoranthene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved phase, 6=lost filter, 7=sampler down, 8=field blank).

Fluoranthene, 1993

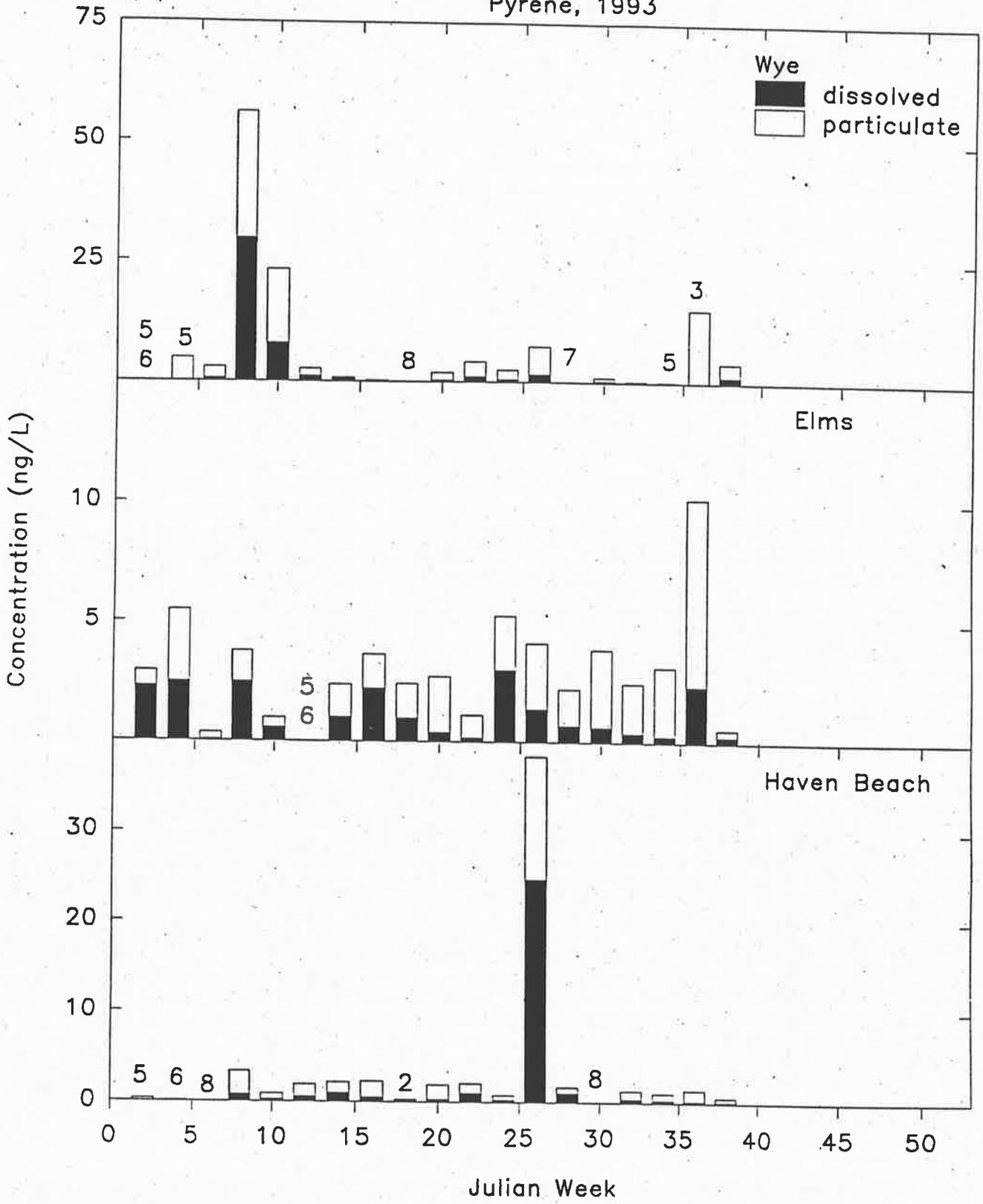


Pyrene, 1992

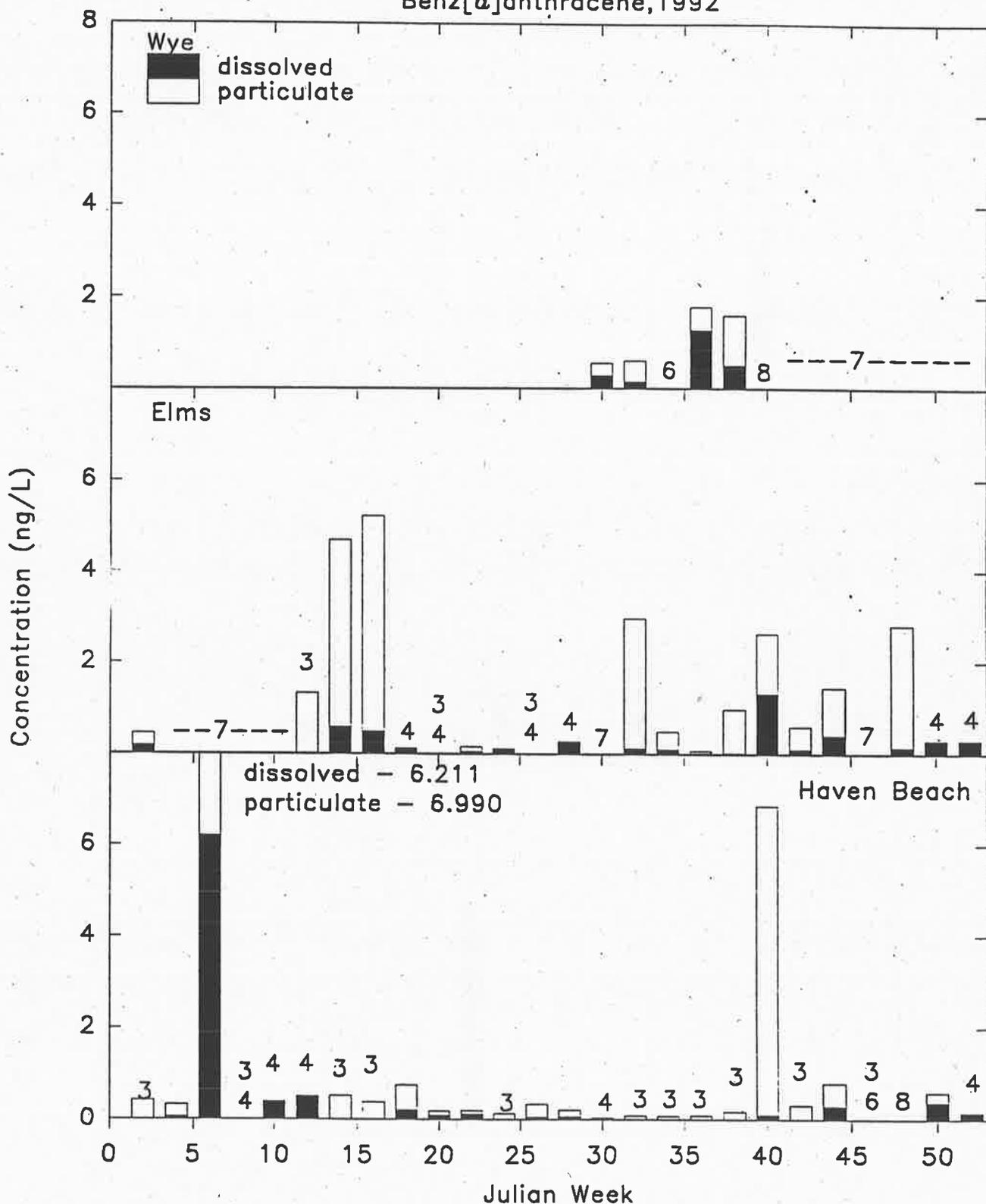


A1.29 Pyrene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8=field blank).

Pyrene, 1993

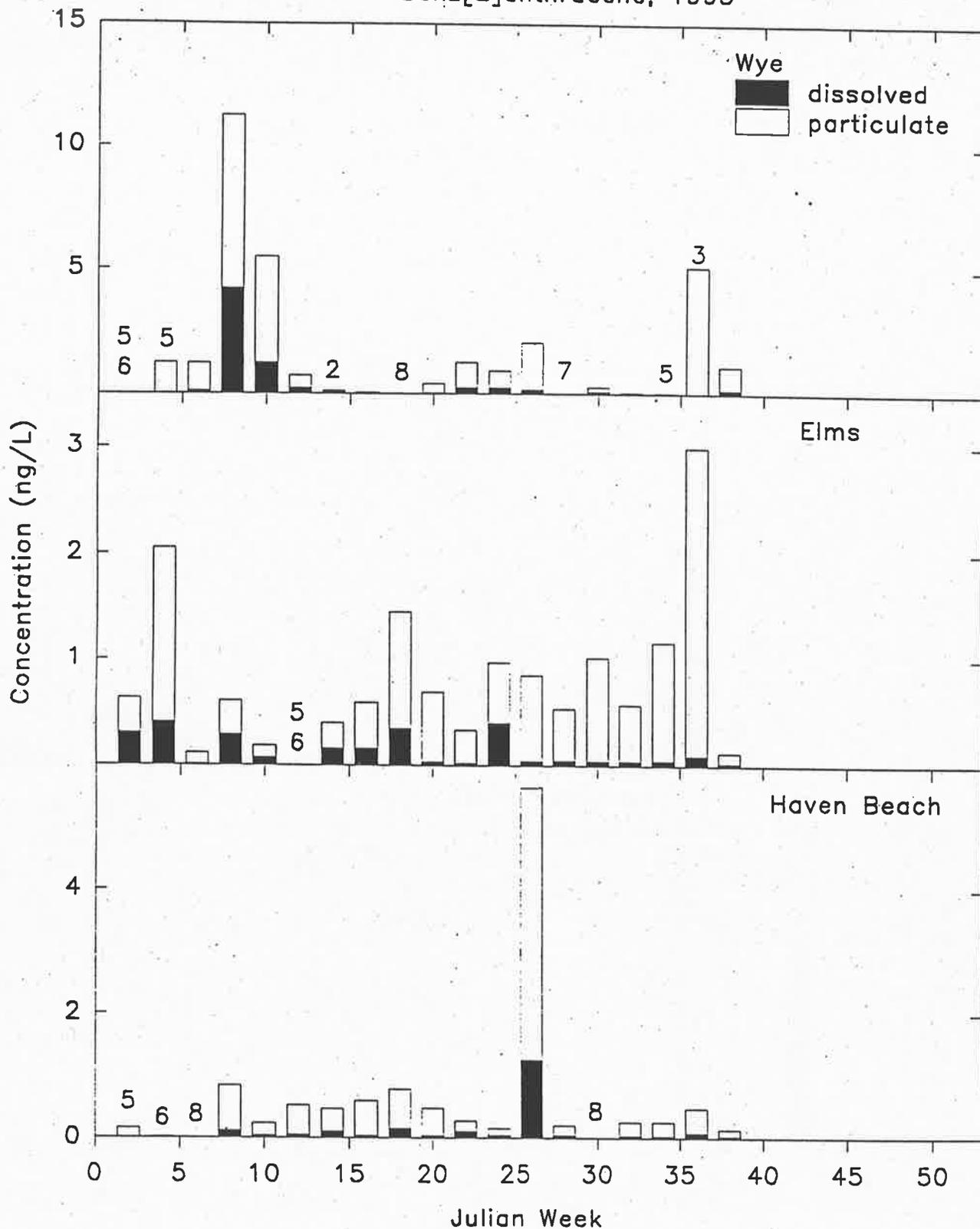


Benz[a]anthracene, 1992

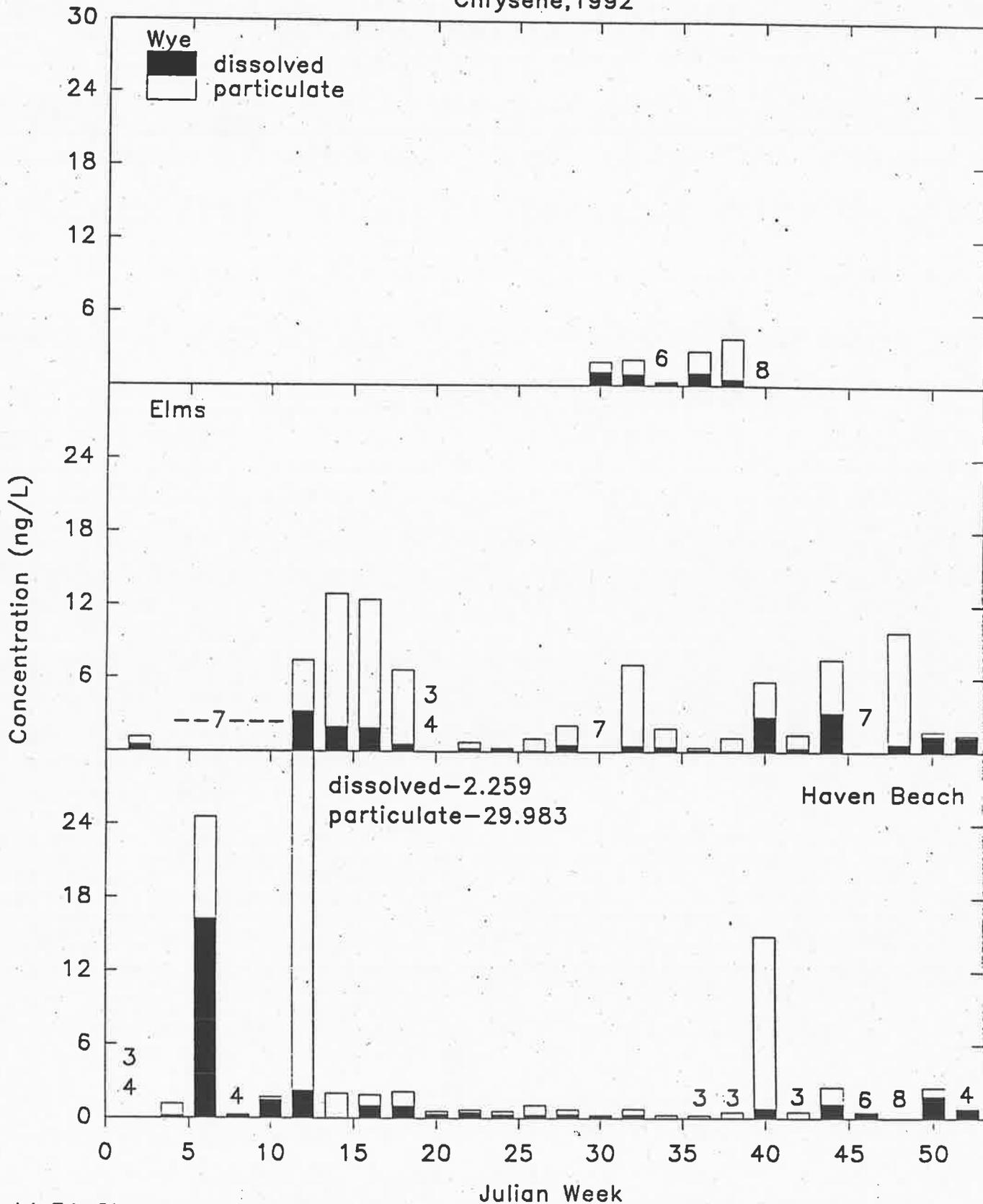


A1.30 Benz[a]anthracene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Benz[a]anthracene, 1993

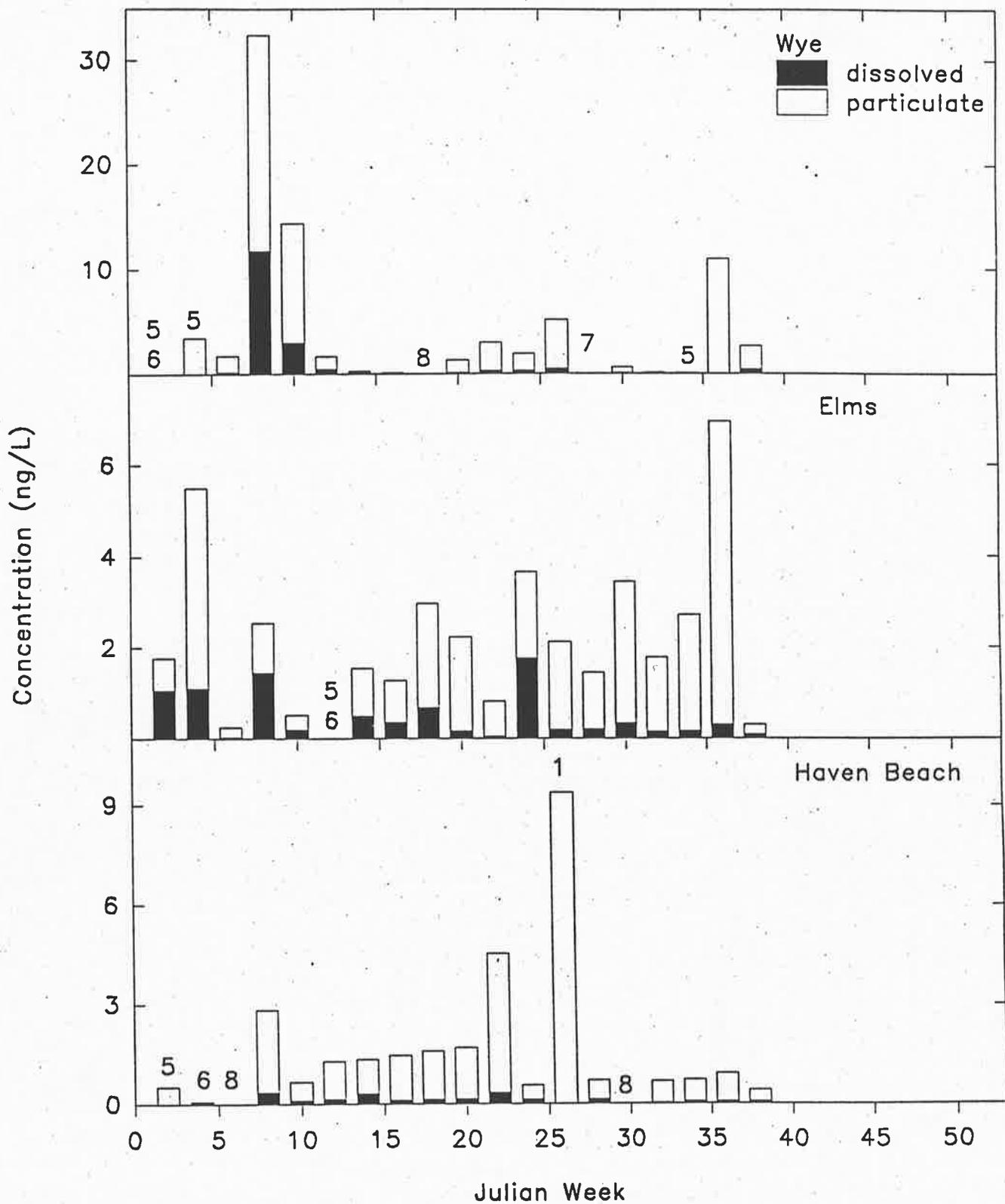


Chrysene, 1992

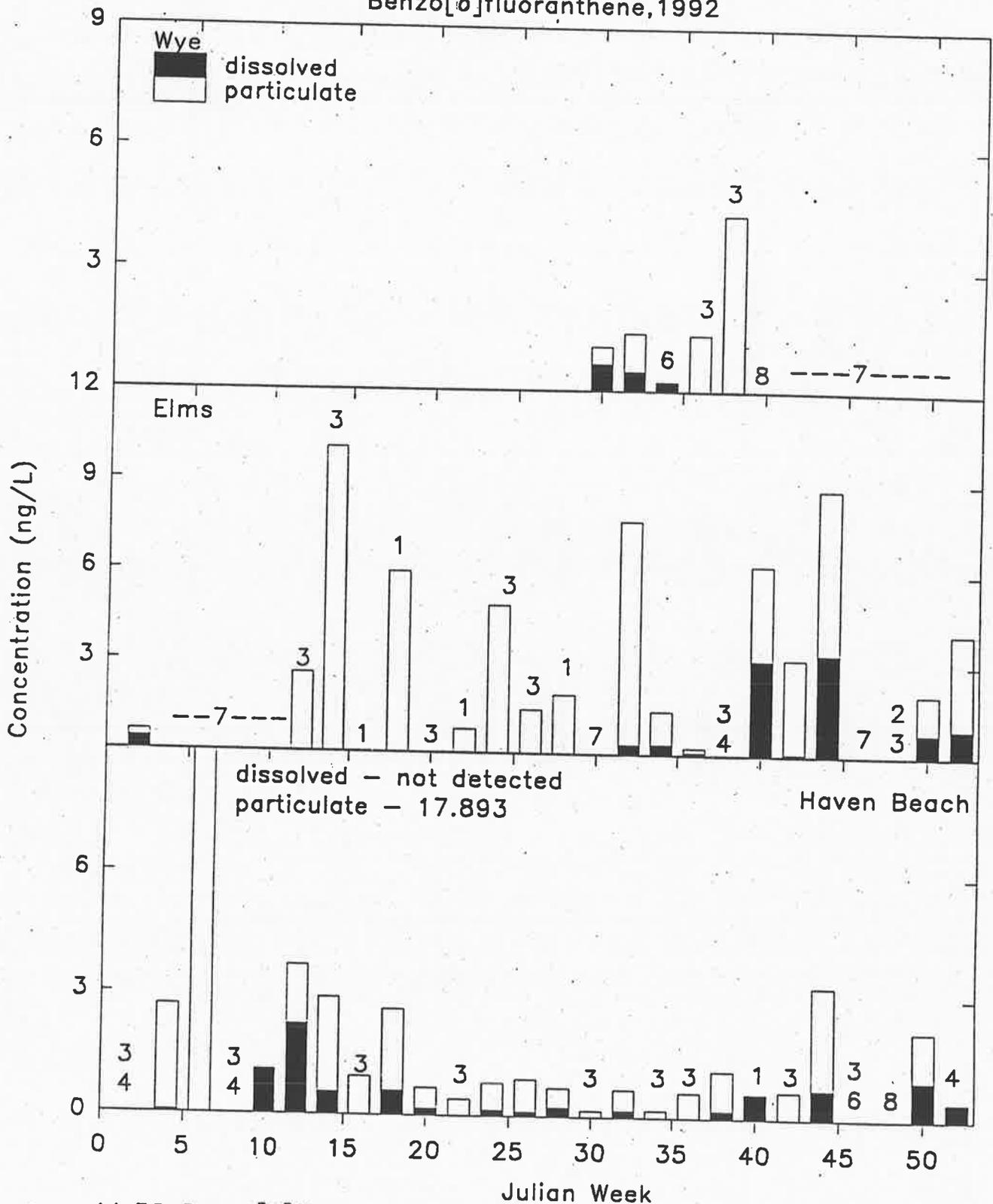


A1.31 Chrysene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Chrysene, 1993

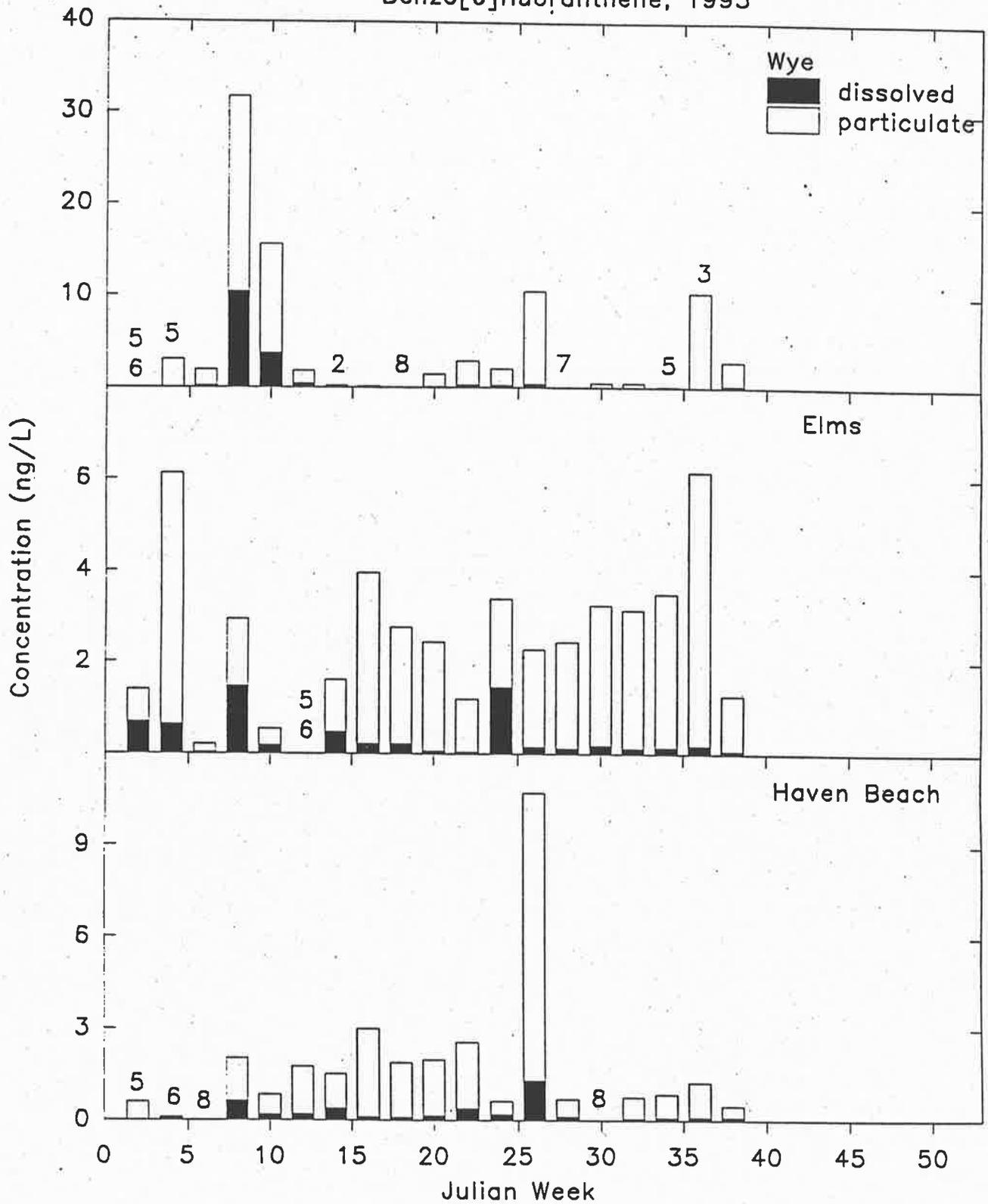


Benzo[b]fluoranthene, 1992

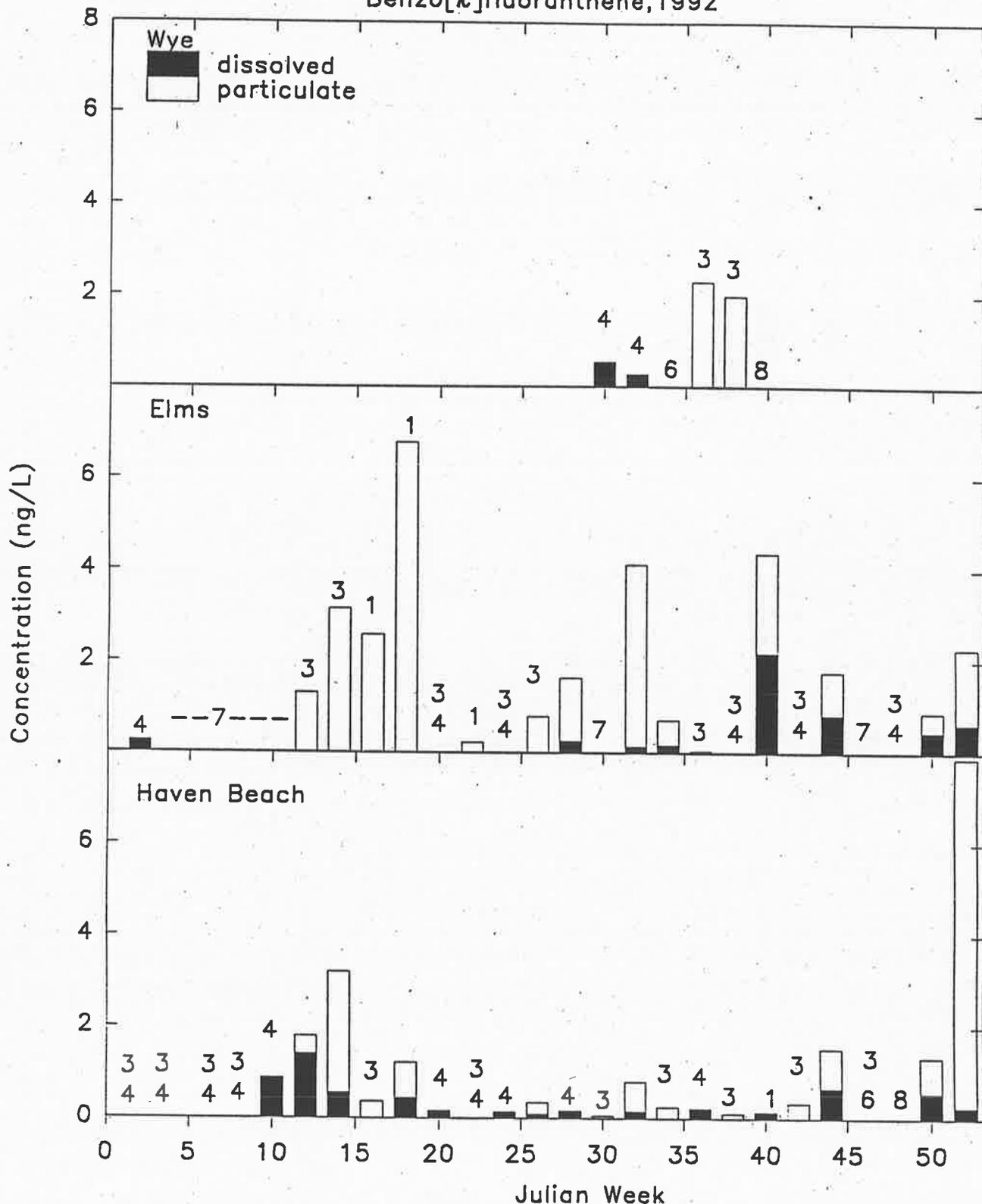


A1.32 Benzo[b]fluoranthene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Benzo[b]fluoranthene, 1993

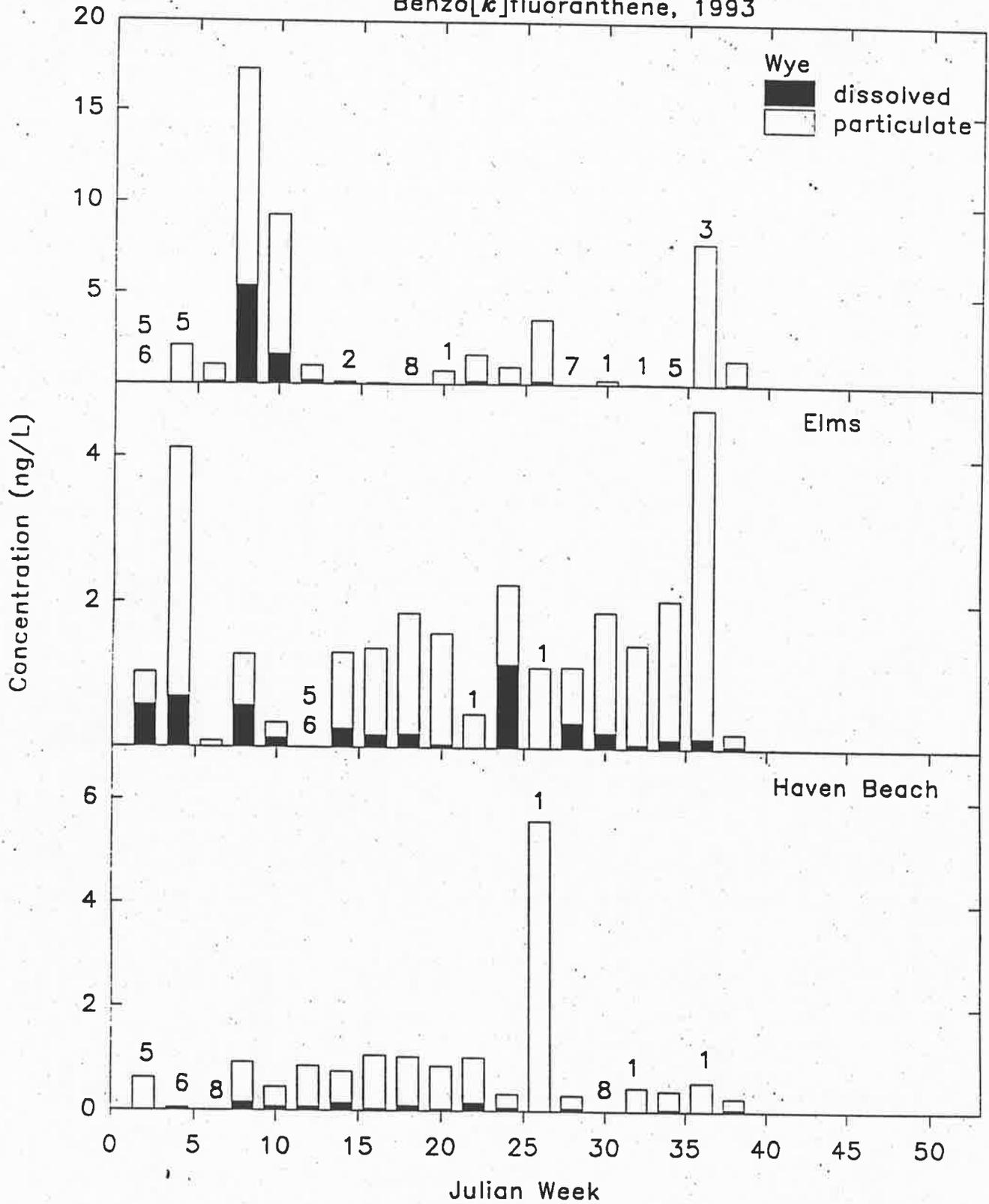


Benzo[k]fluoranthene, 1992

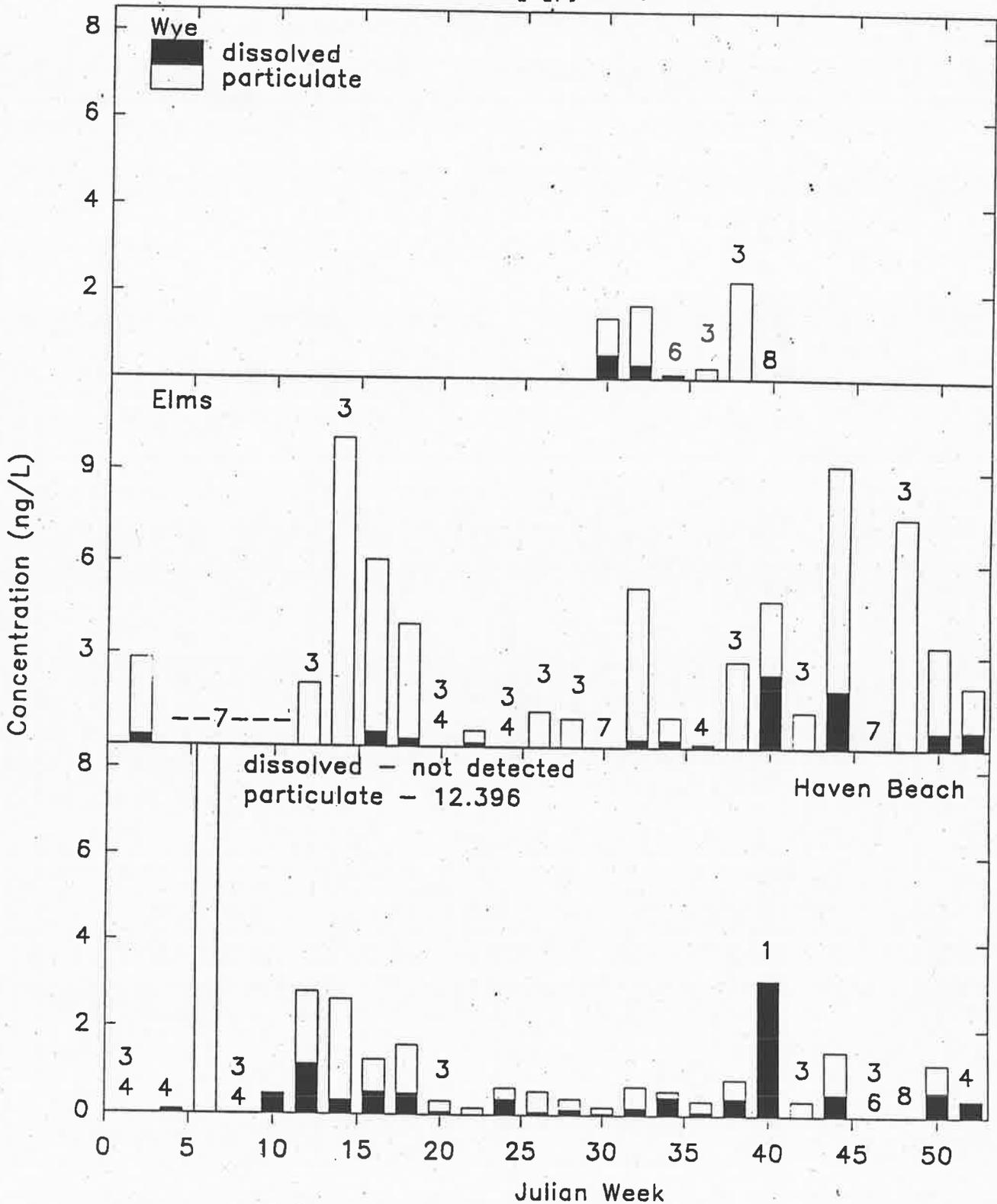


A1.33 Benzo[k]fluoranthene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved phase, 6=lost filter, 7=sampler down, 8= no ppt.).

Benzo[k]fluoranthene, 1993

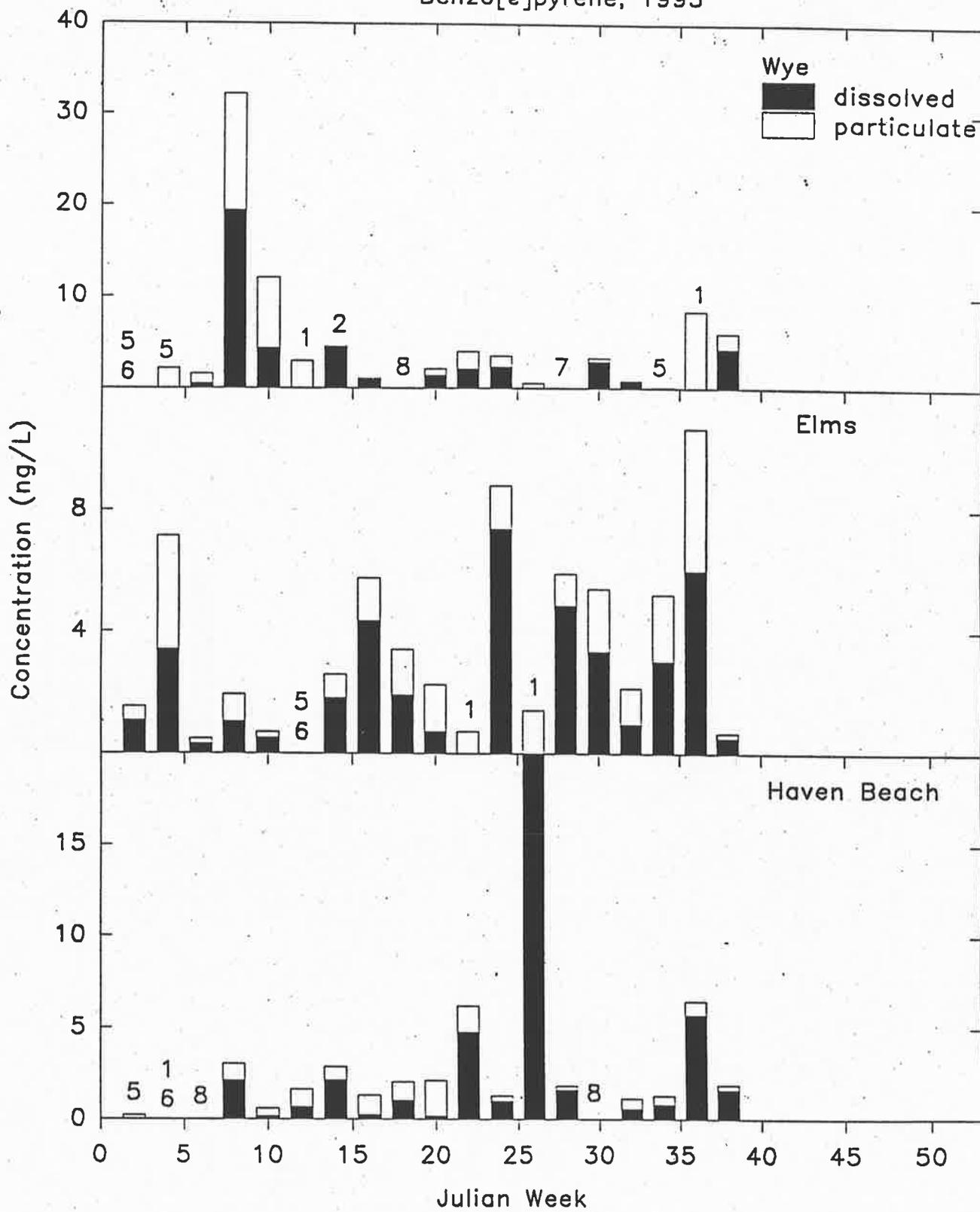


Benzo[e]pyrene, 1992

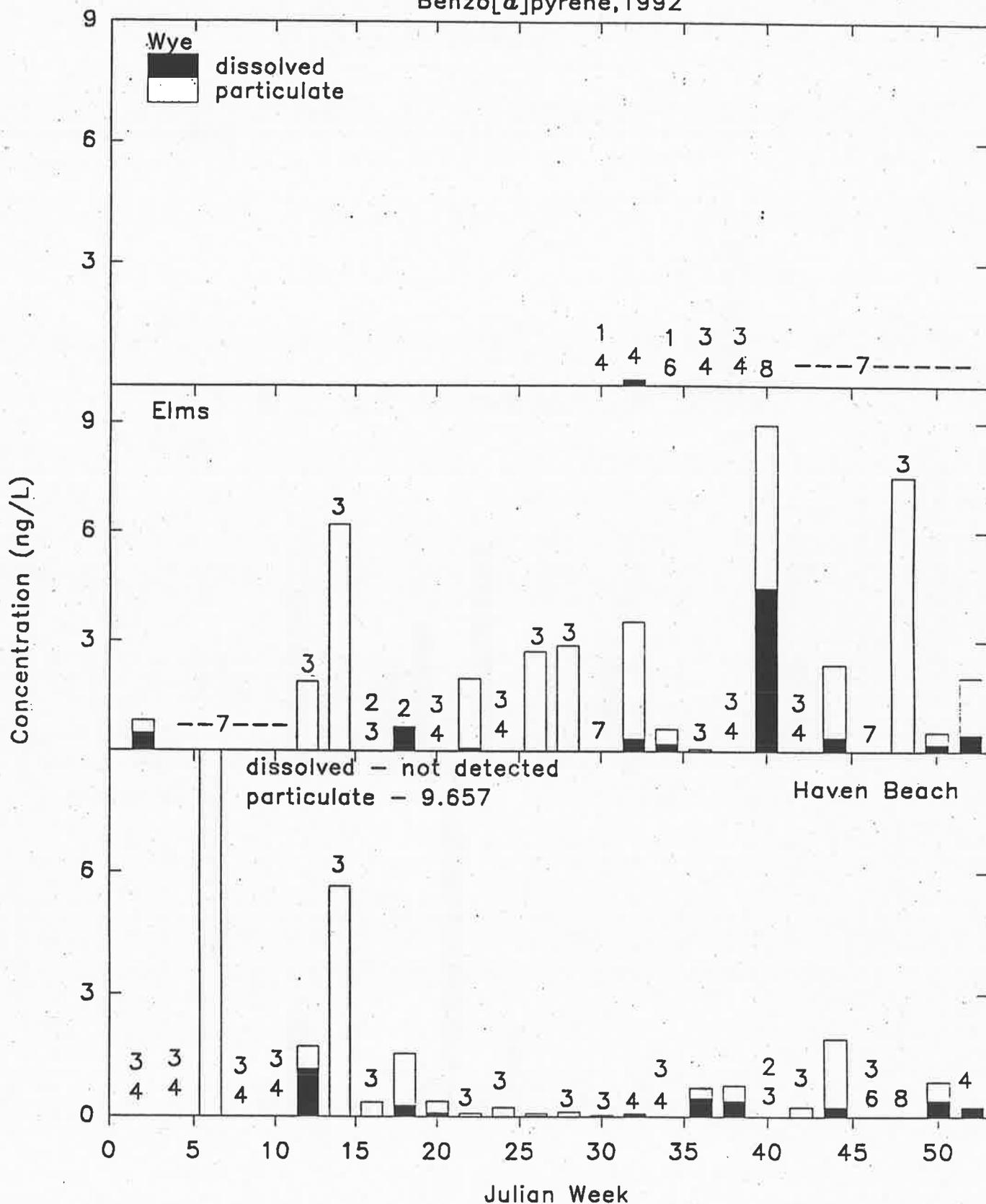


A1.34 Benzo[e]pyrene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8=field blank).

Benzo[e]pyrene, 1993

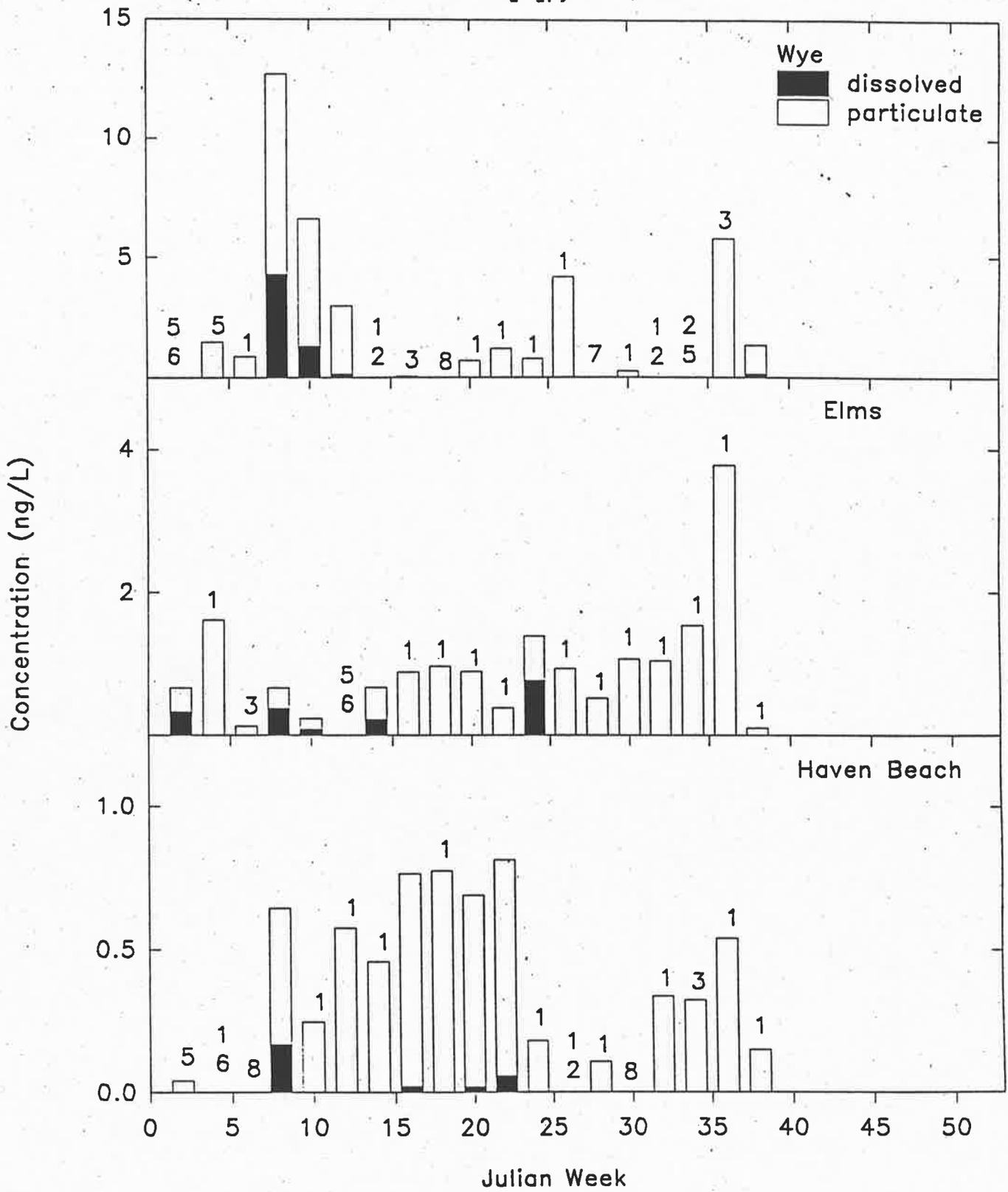


Benzo[a]pyrene, 1992

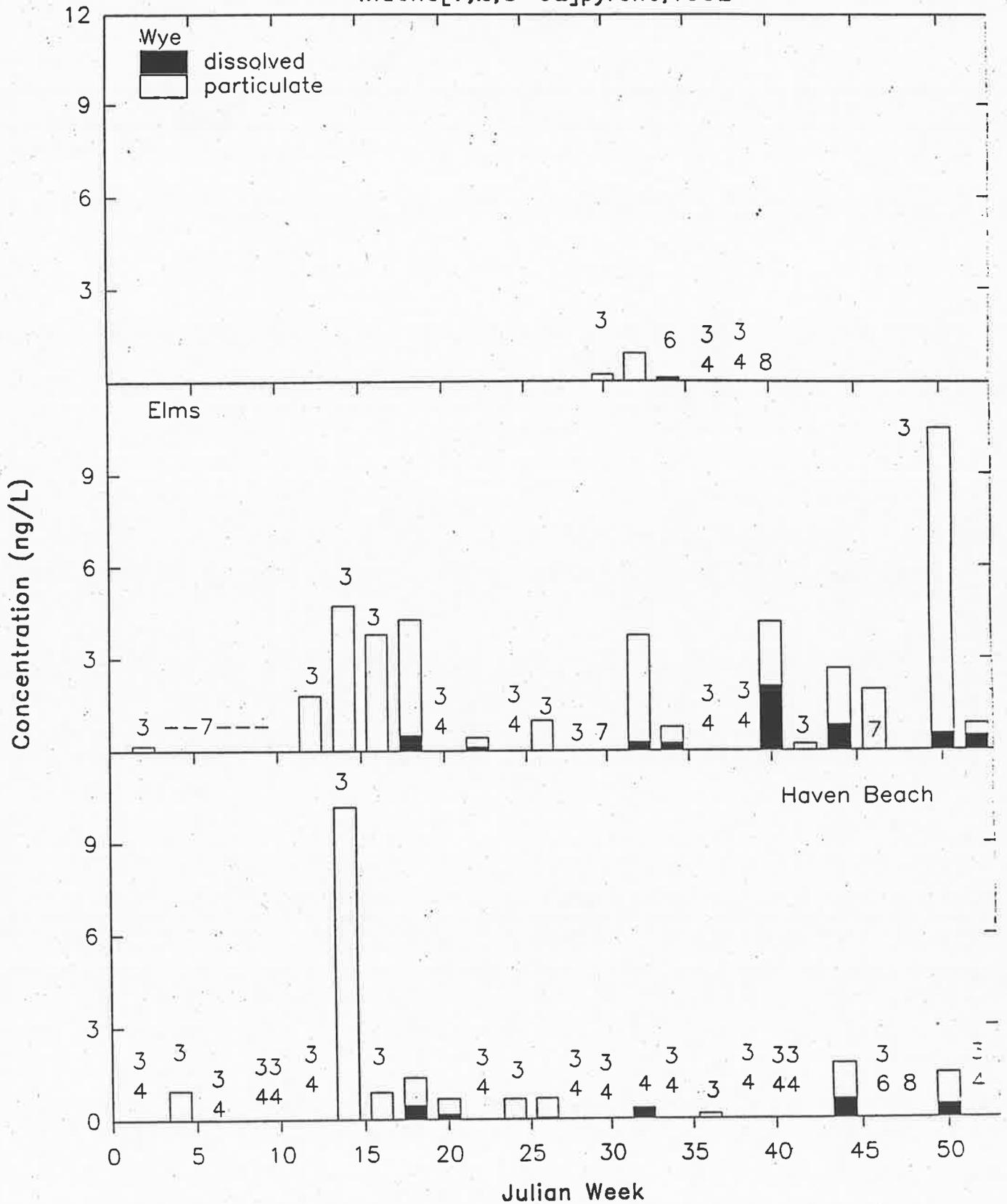


A1.35 Benzo[a]pyrene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Benzo[a]pyrene, 1993

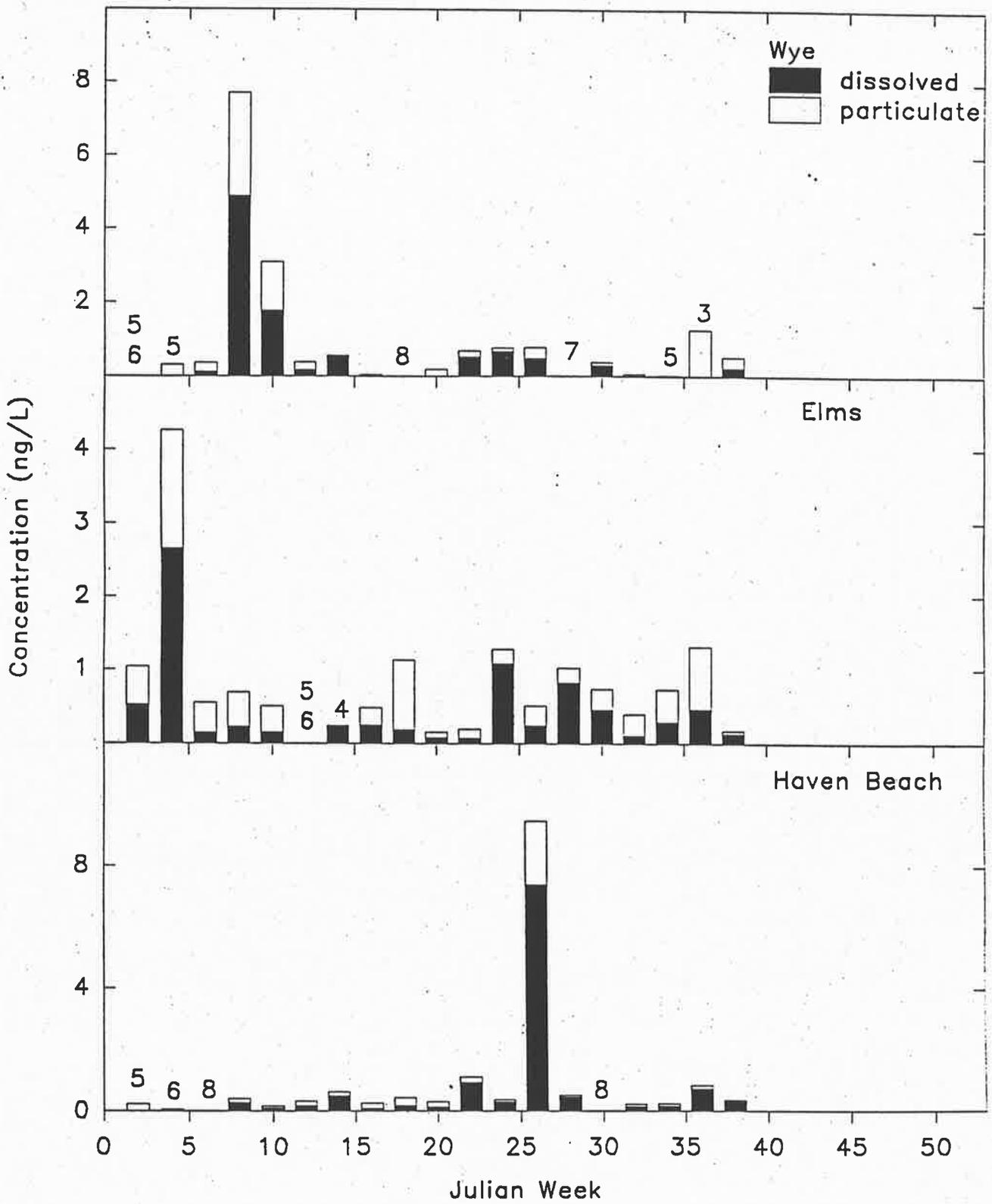


Indeno[1,2,3-cd]pyrene, 1992

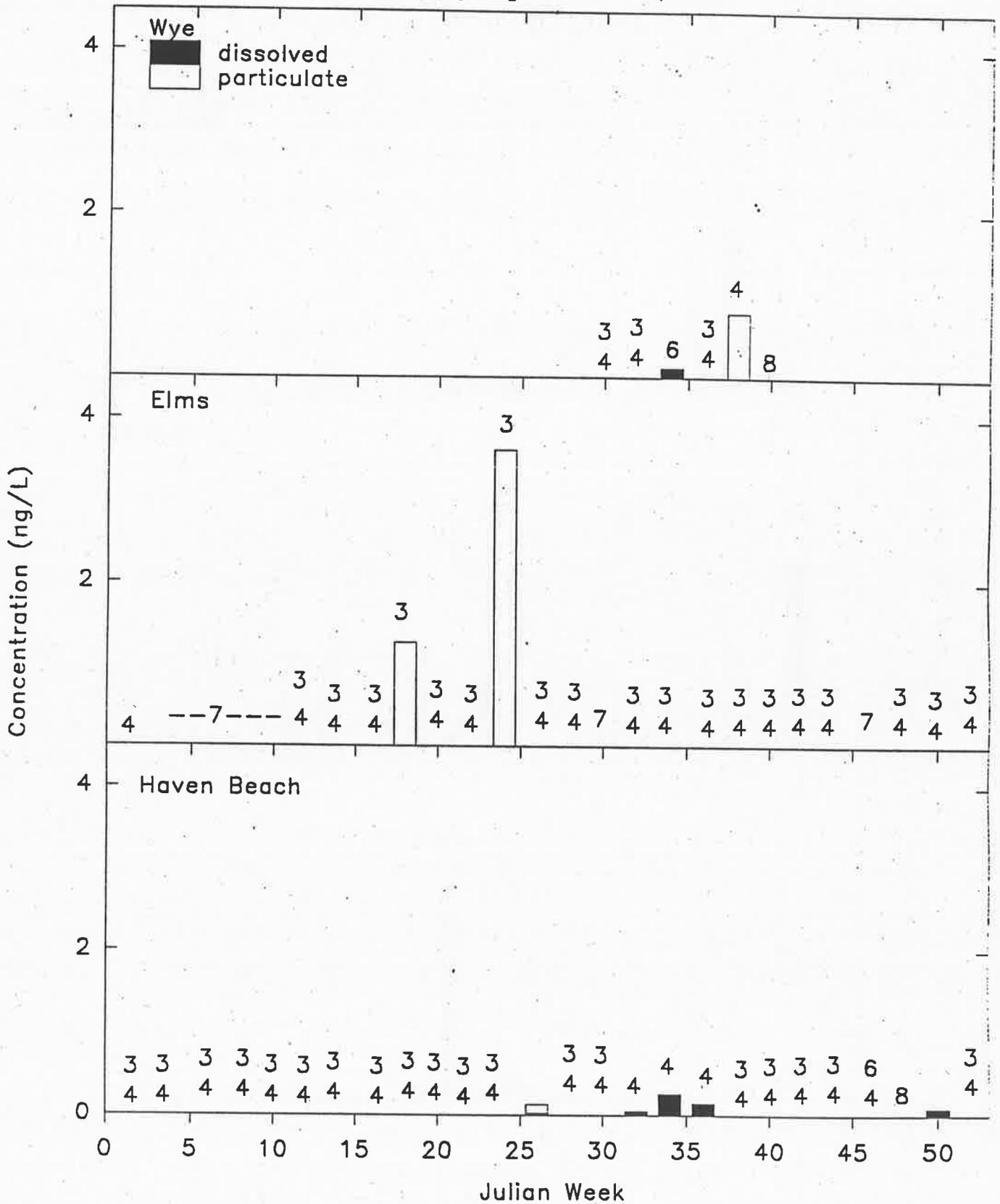


A1.36 Indeno[1,2,3-cd]pyrene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8=field blank).

Indeno[1,2,3-cd]pyrene, 1993

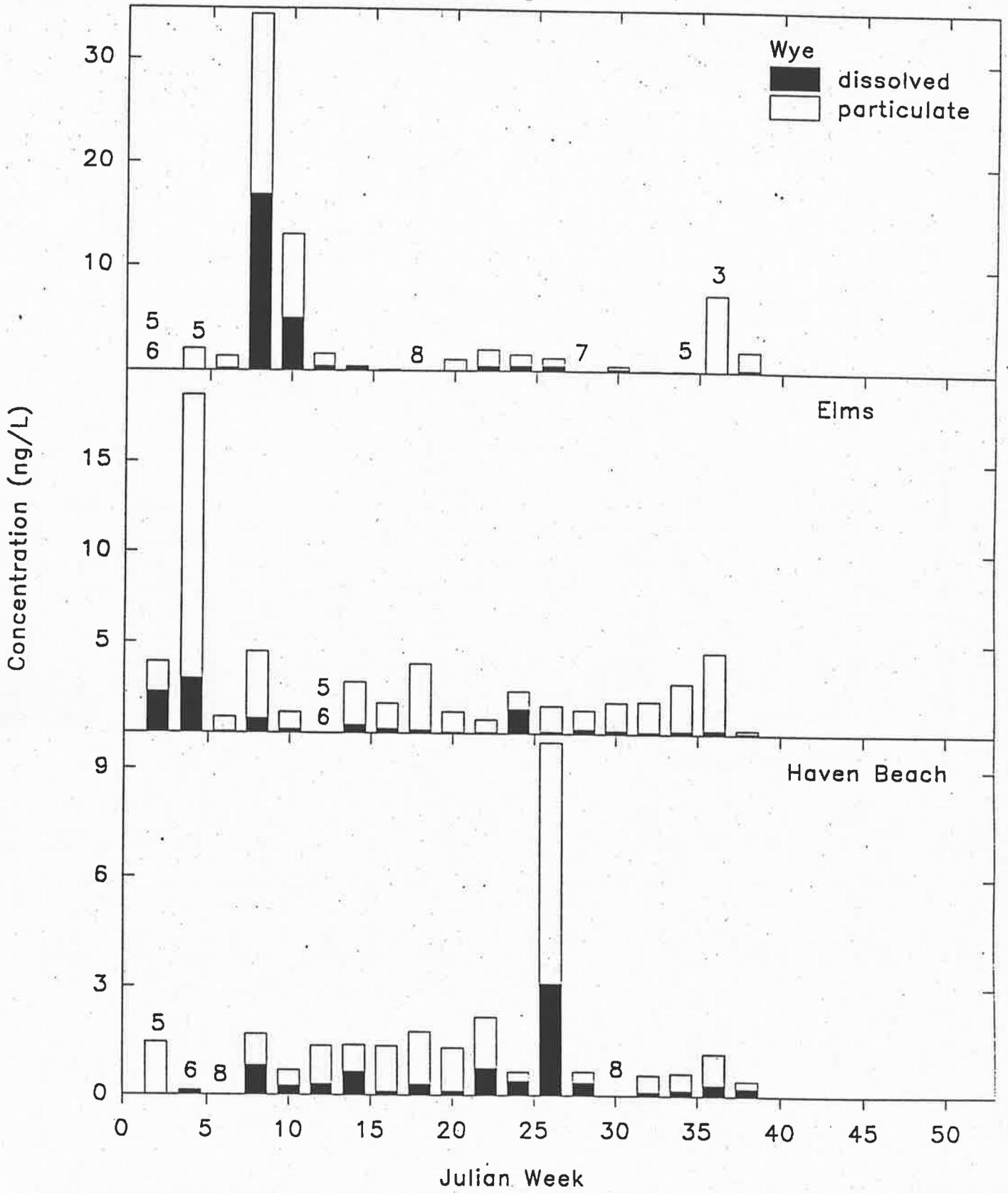


Dibenz[*a,h*]anthracene, 1992

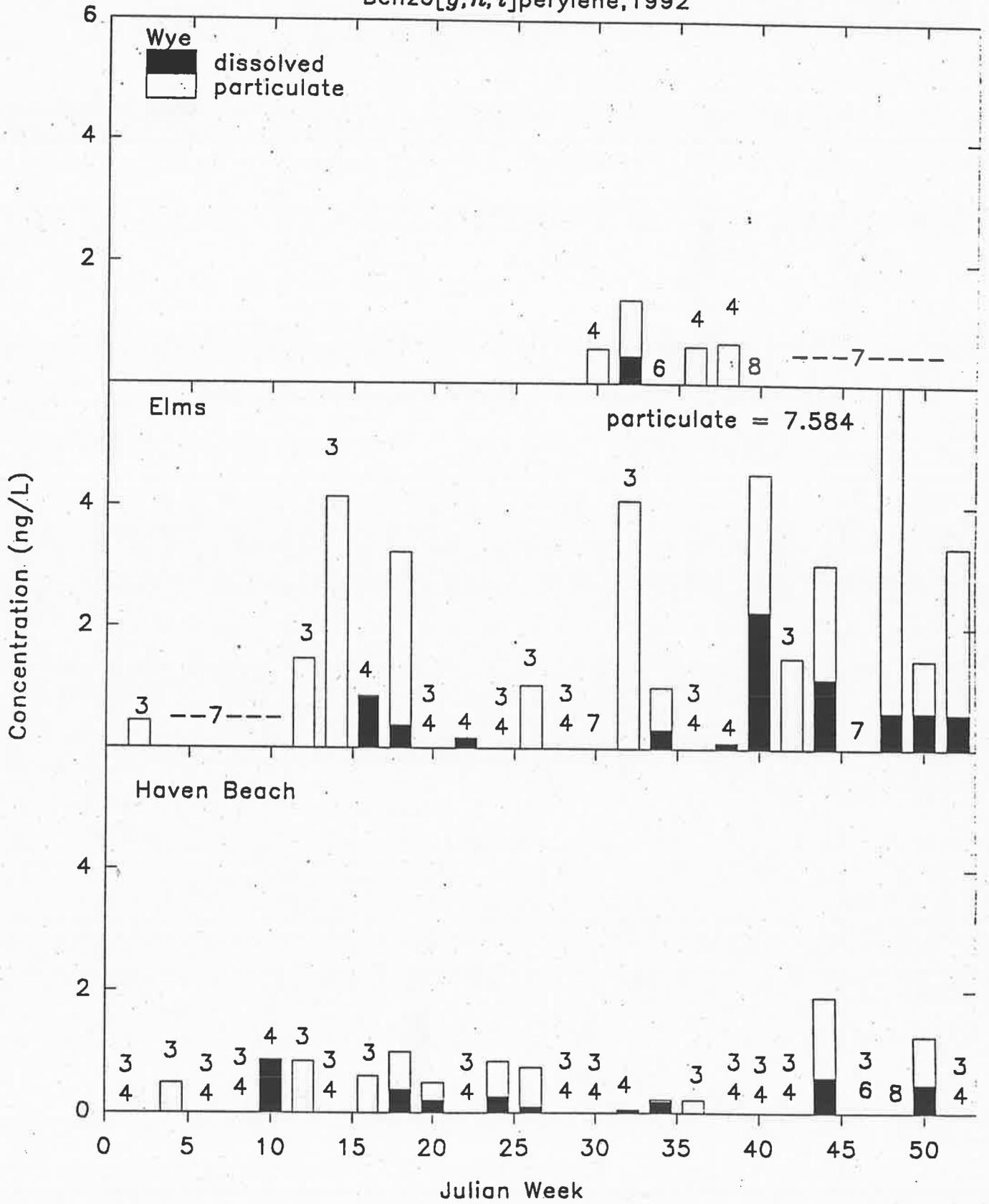


A1.37 Dibenzo[*a,h*]anthracene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8=field blank).

Dibenz[*a,h*]anthracene, 1993

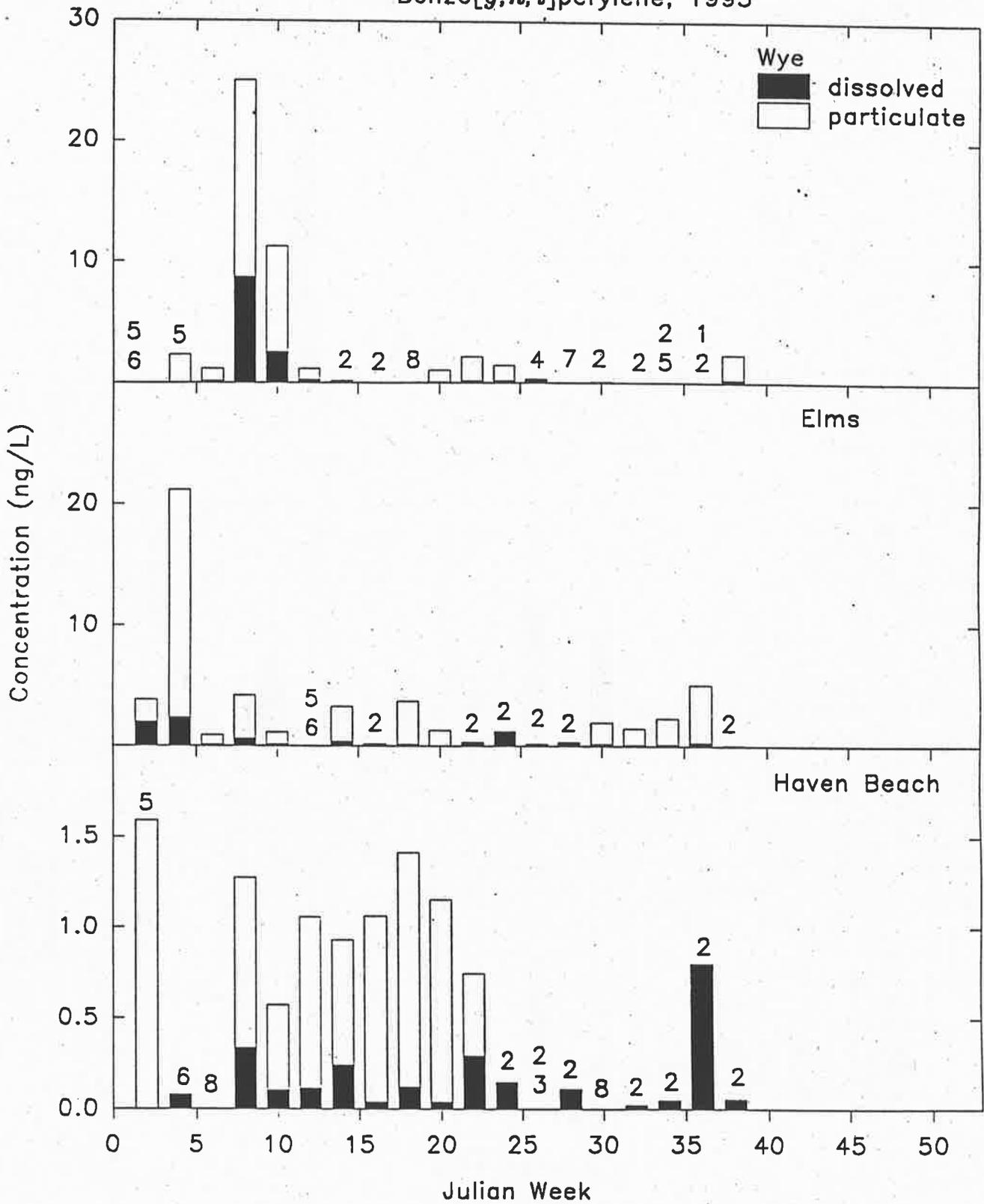


Benzo[*g,h,i*]perylene, 1992

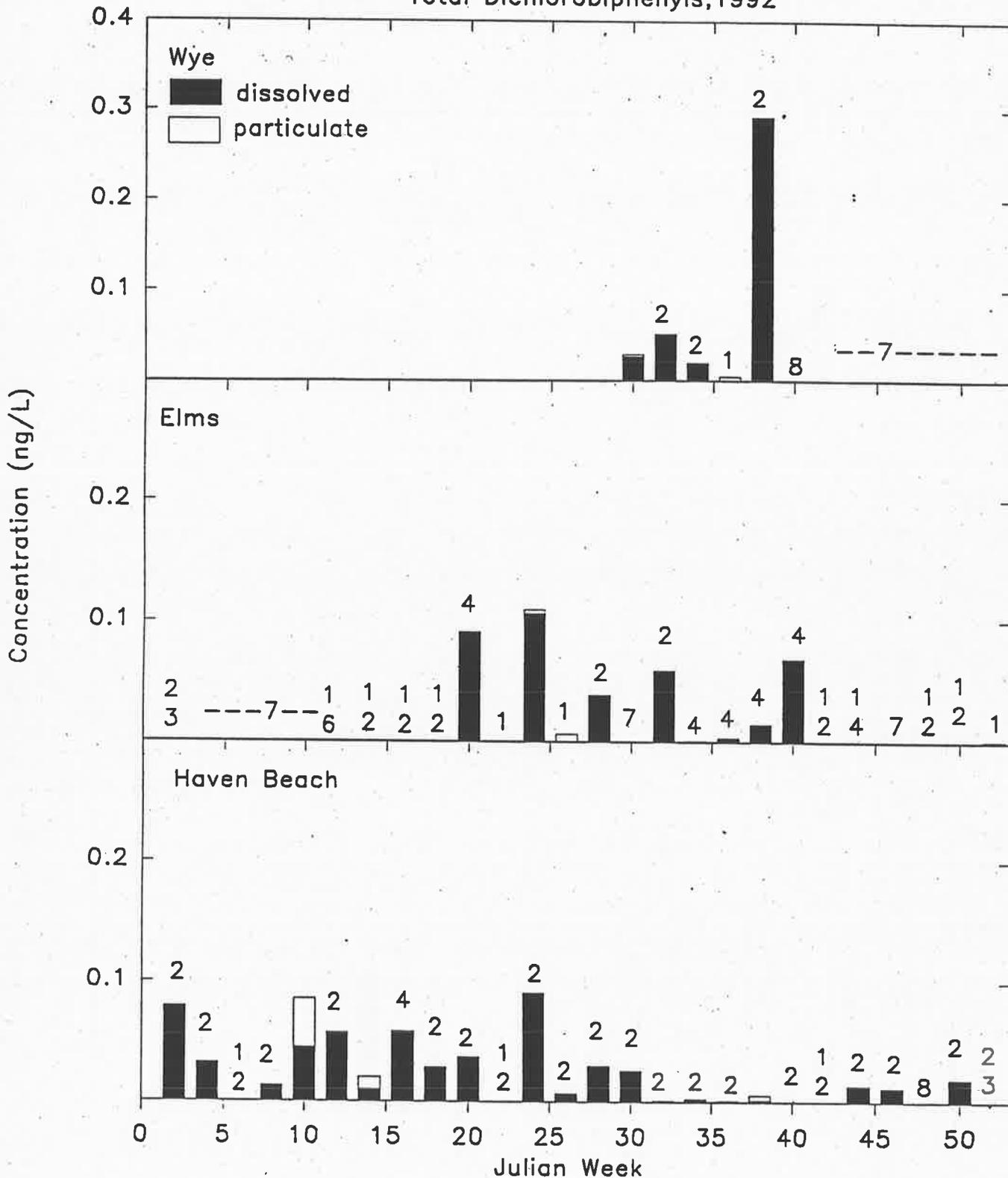


A1.38 Benzo[*g,h,i*]perylene concentrations in precipitation integrated biweekly at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=not detected in dissolved phase, 4=not detected on filter, 5=lost dissolved sample, 6=lost filter, 7=sampler down, 8= no ppt.).

Benzo[*g,h,i*]perylene, 1993

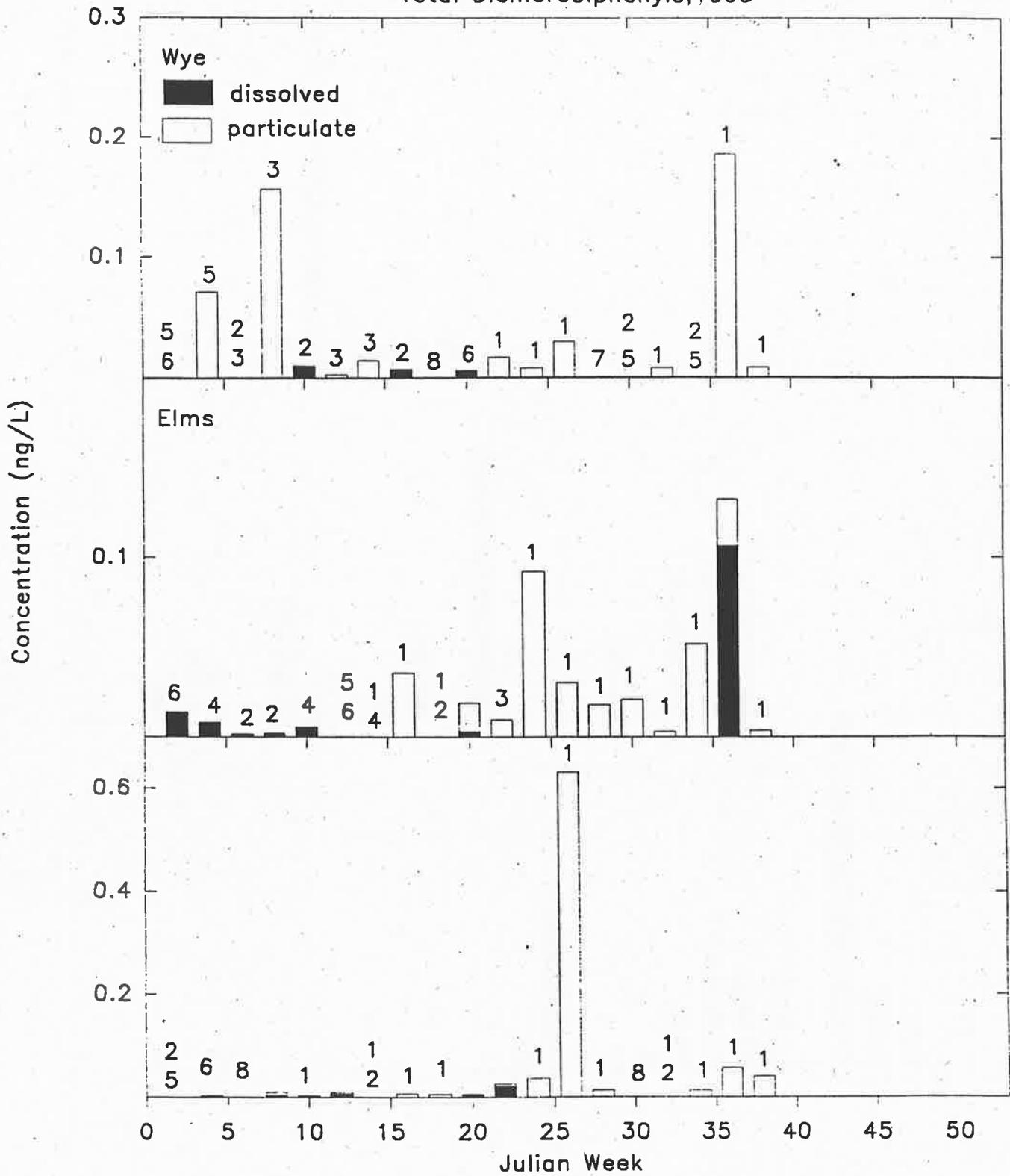


Total Dichlorobiphenyls, 1992

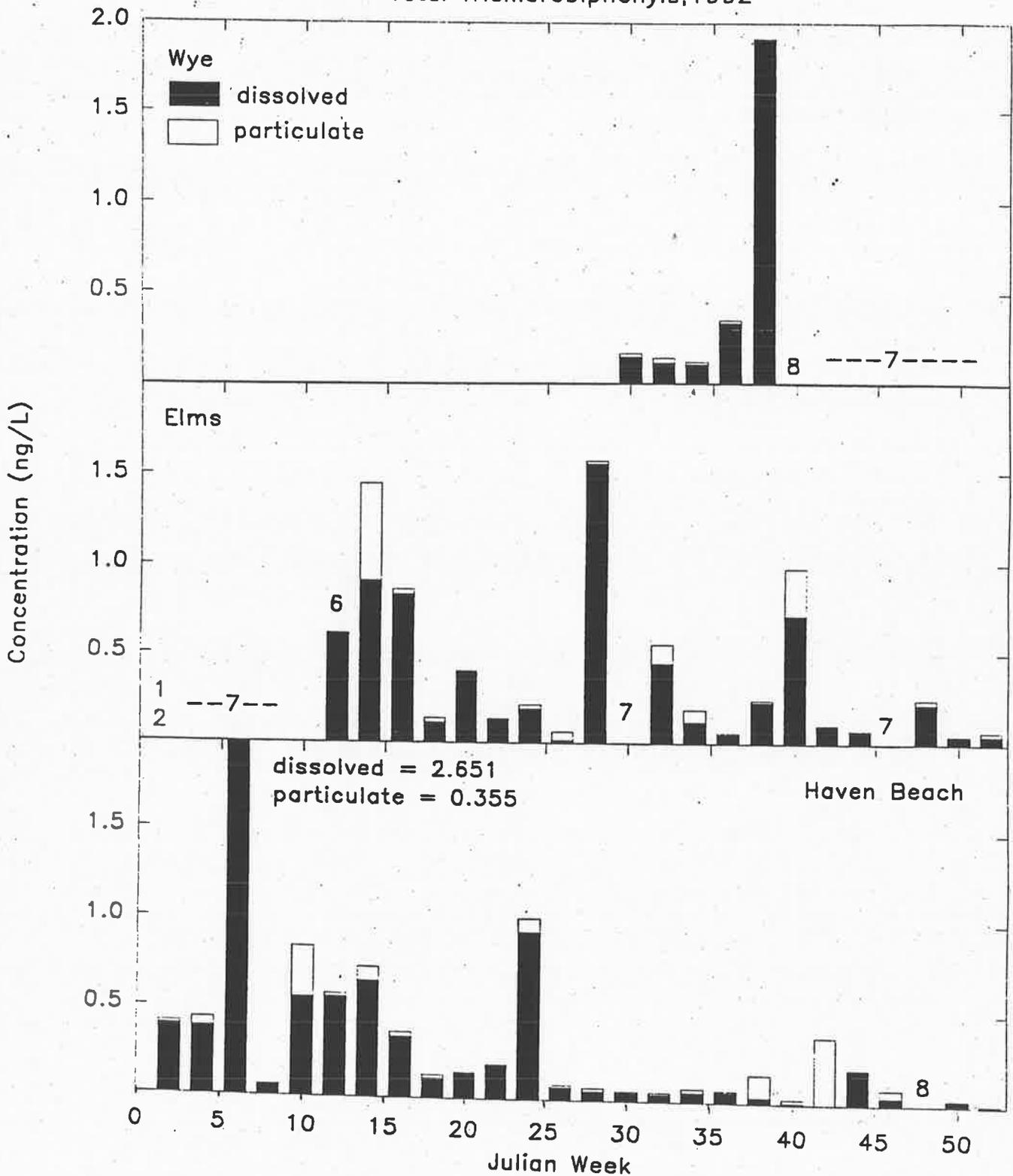


A1.39 Total dichlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Dichlorobiphenyls, 1993

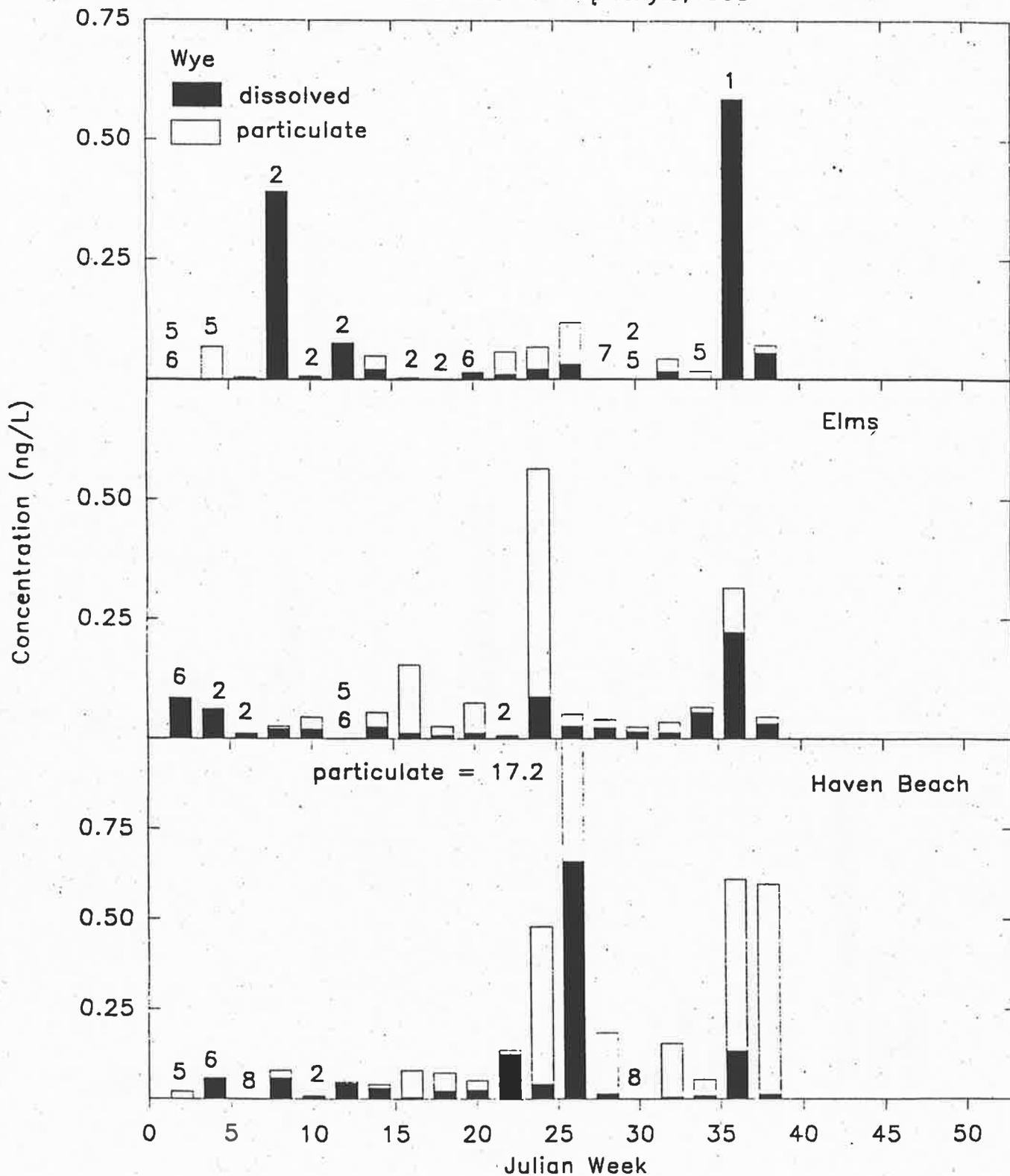


Total Trichlorobiphenyls, 1992

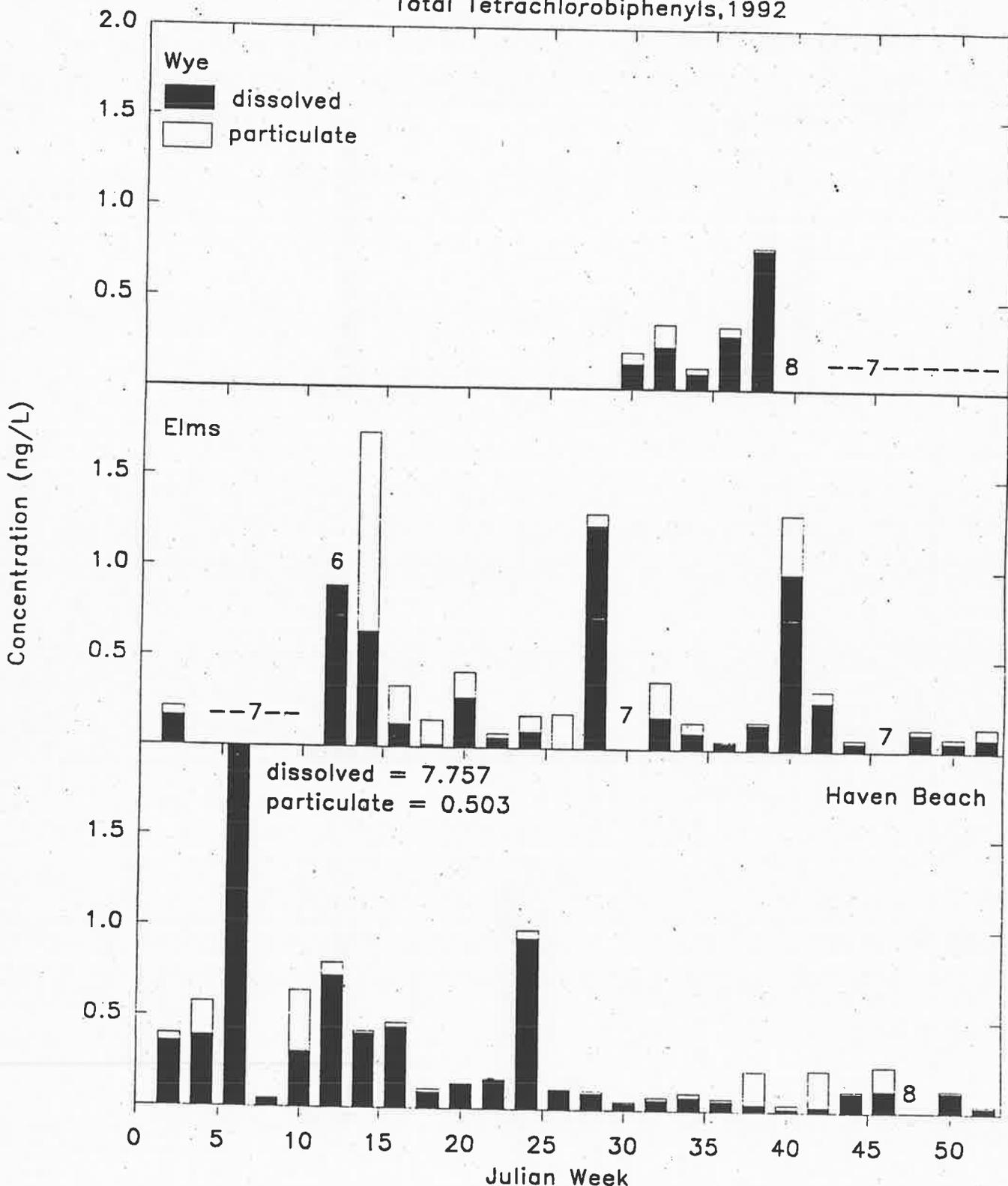


A1.40 Total trichlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Trichlorobiphenyls, 1993

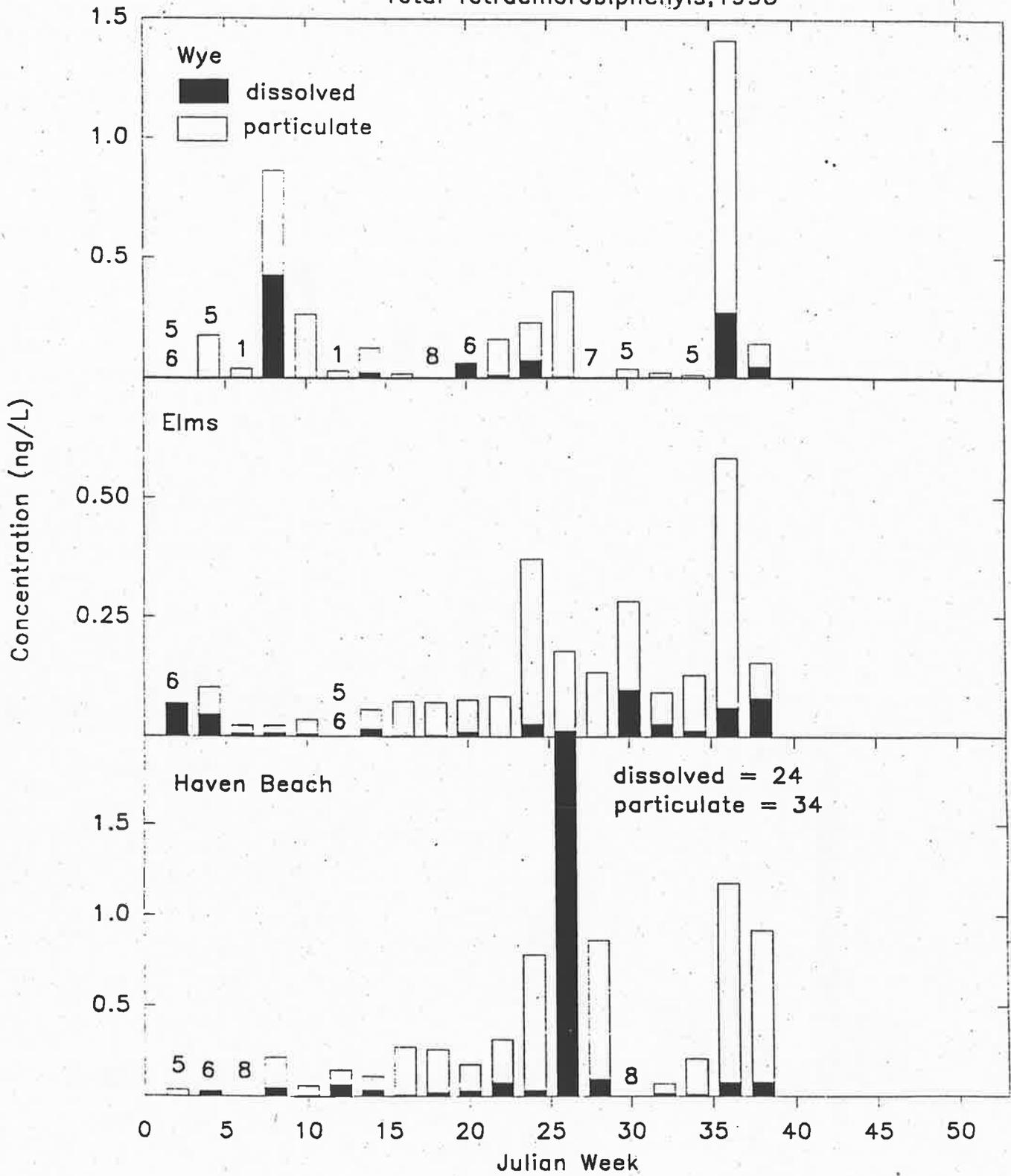


Total Tetrachlorobiphenyls, 1992

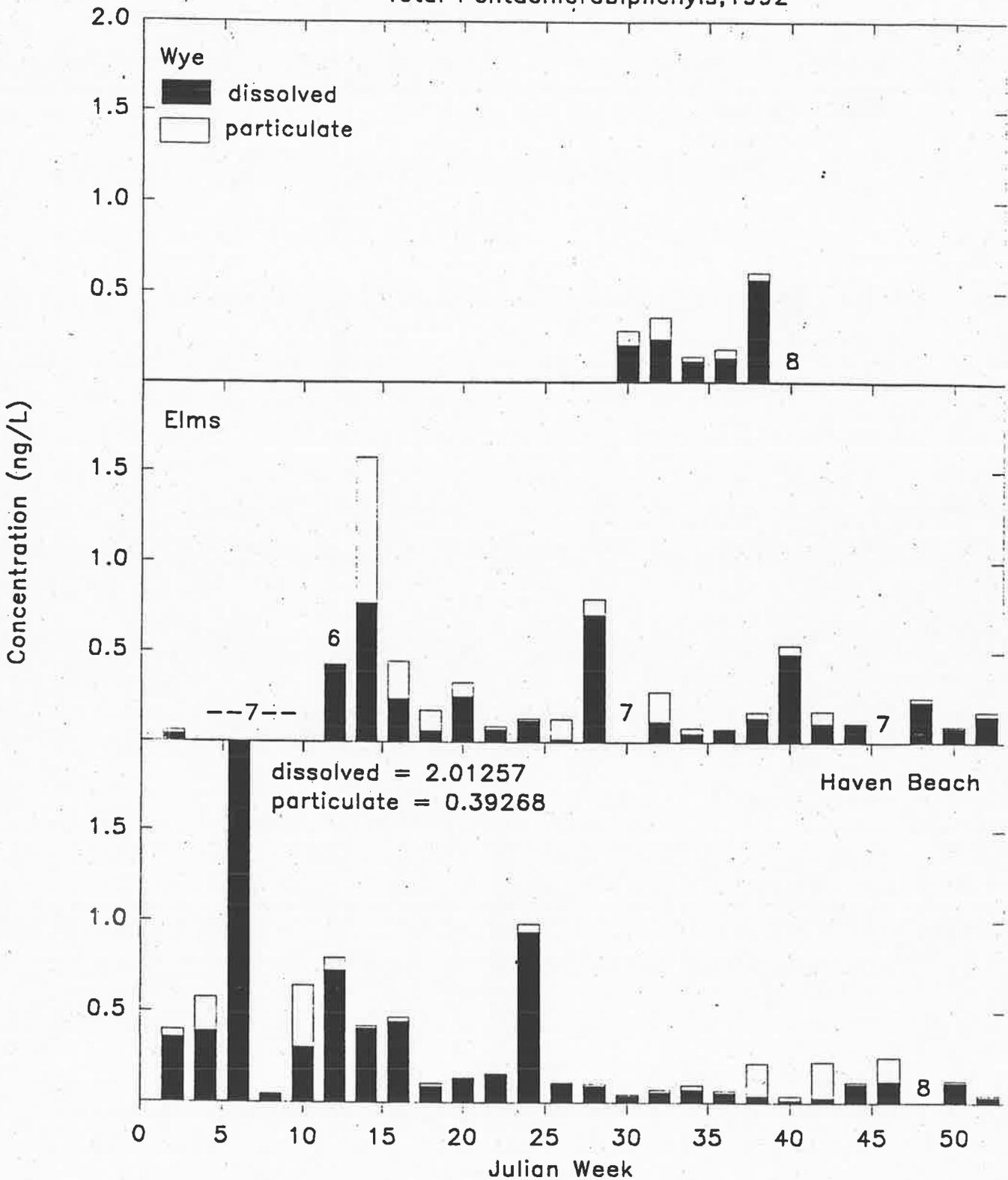


A1.41 Total tetrachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Tetrachlorobiphenyls, 1993

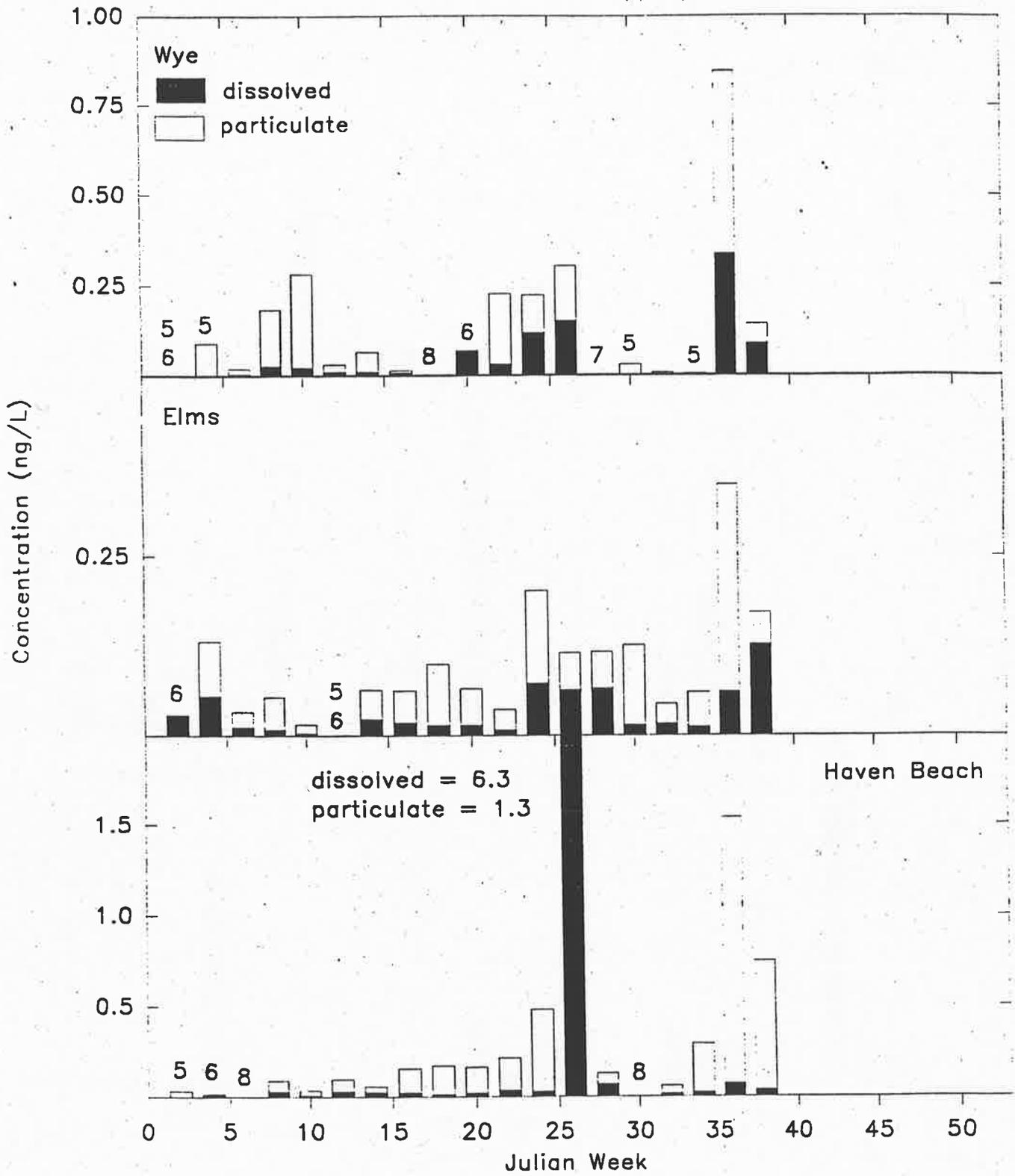


Total Pentachlorobiphenyls, 1992

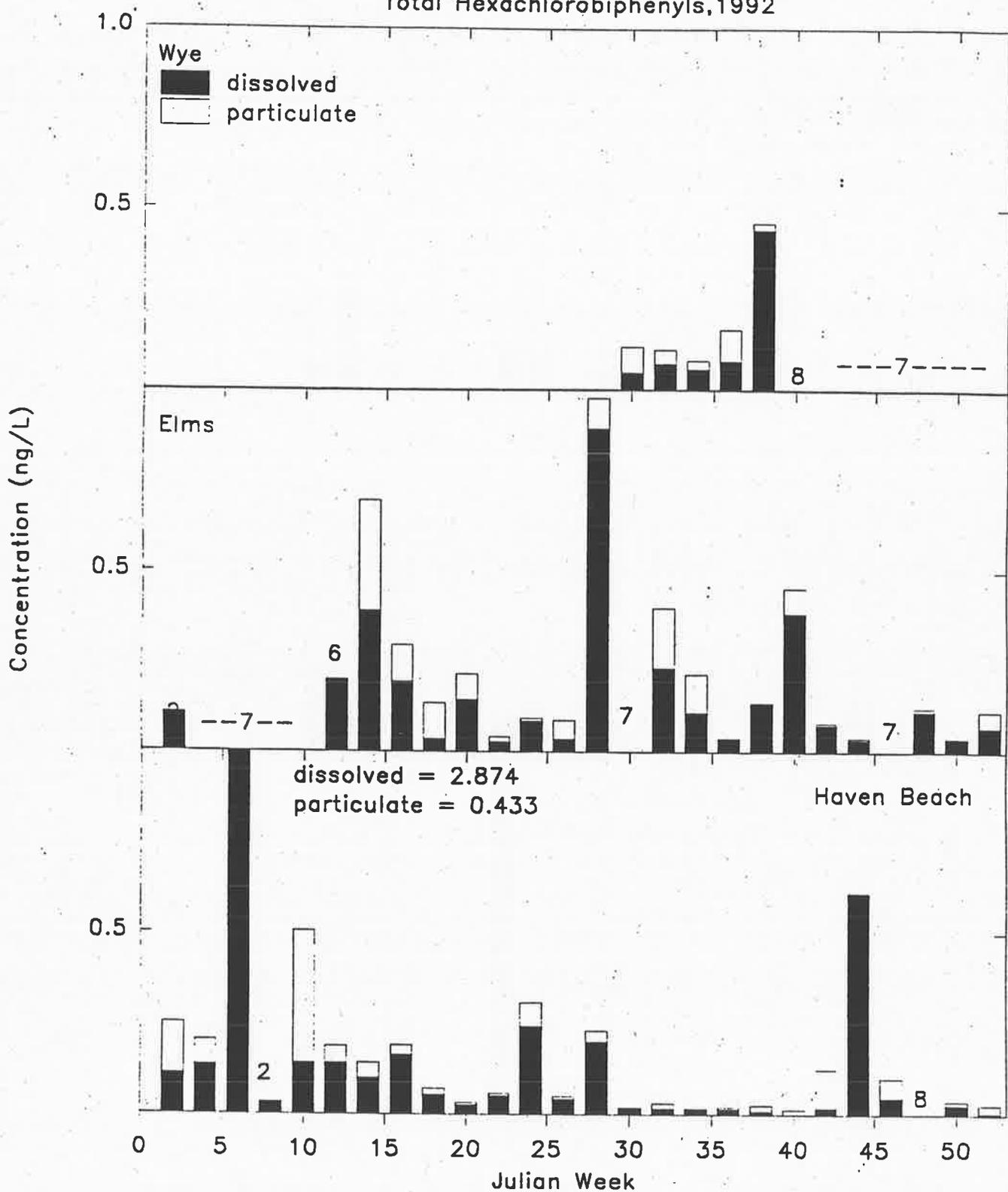


A1.42 Total pentachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.)

Total Pentachlorobiphenyls, 1993

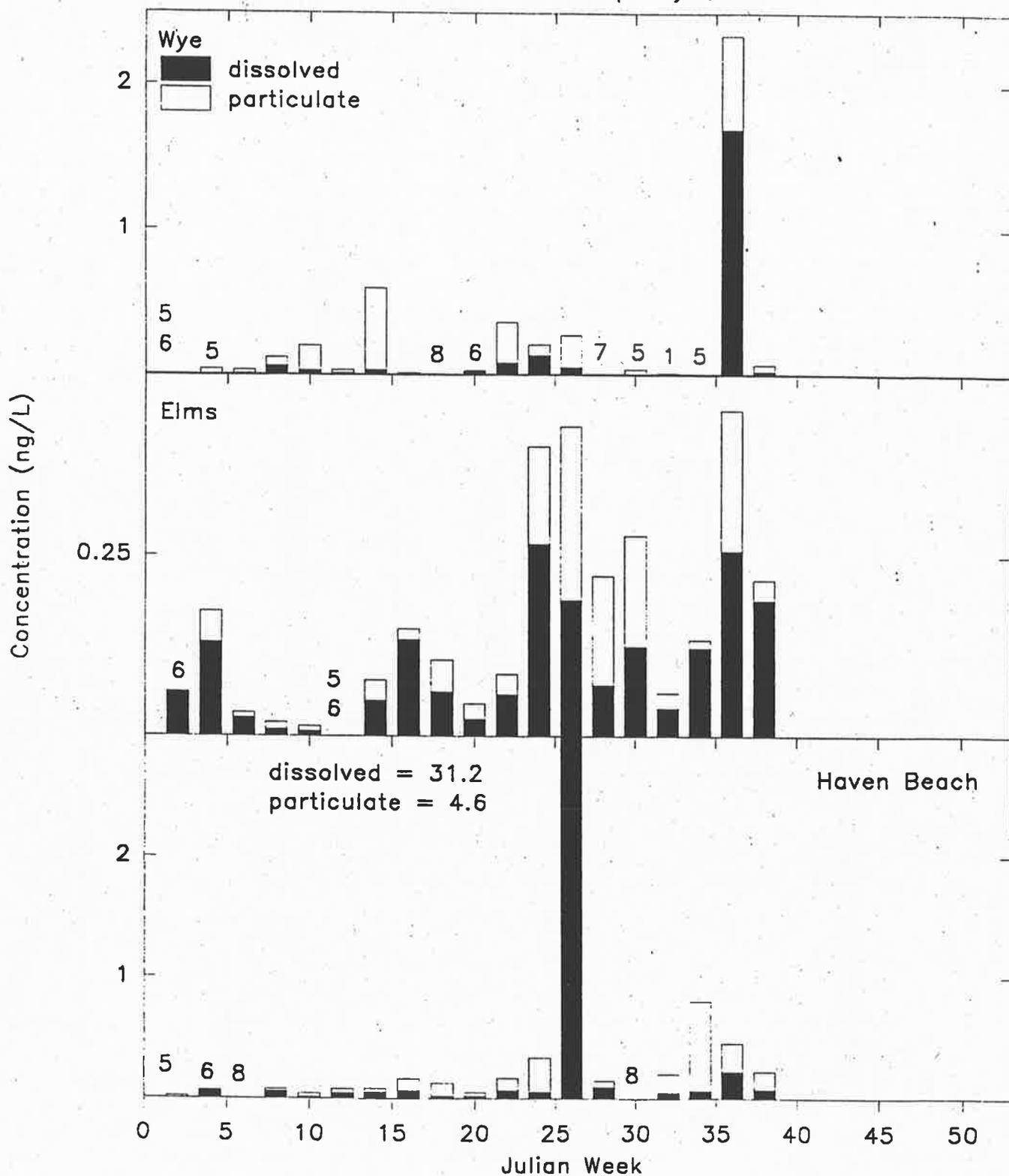


Total Hexachlorobiphenyls, 1992

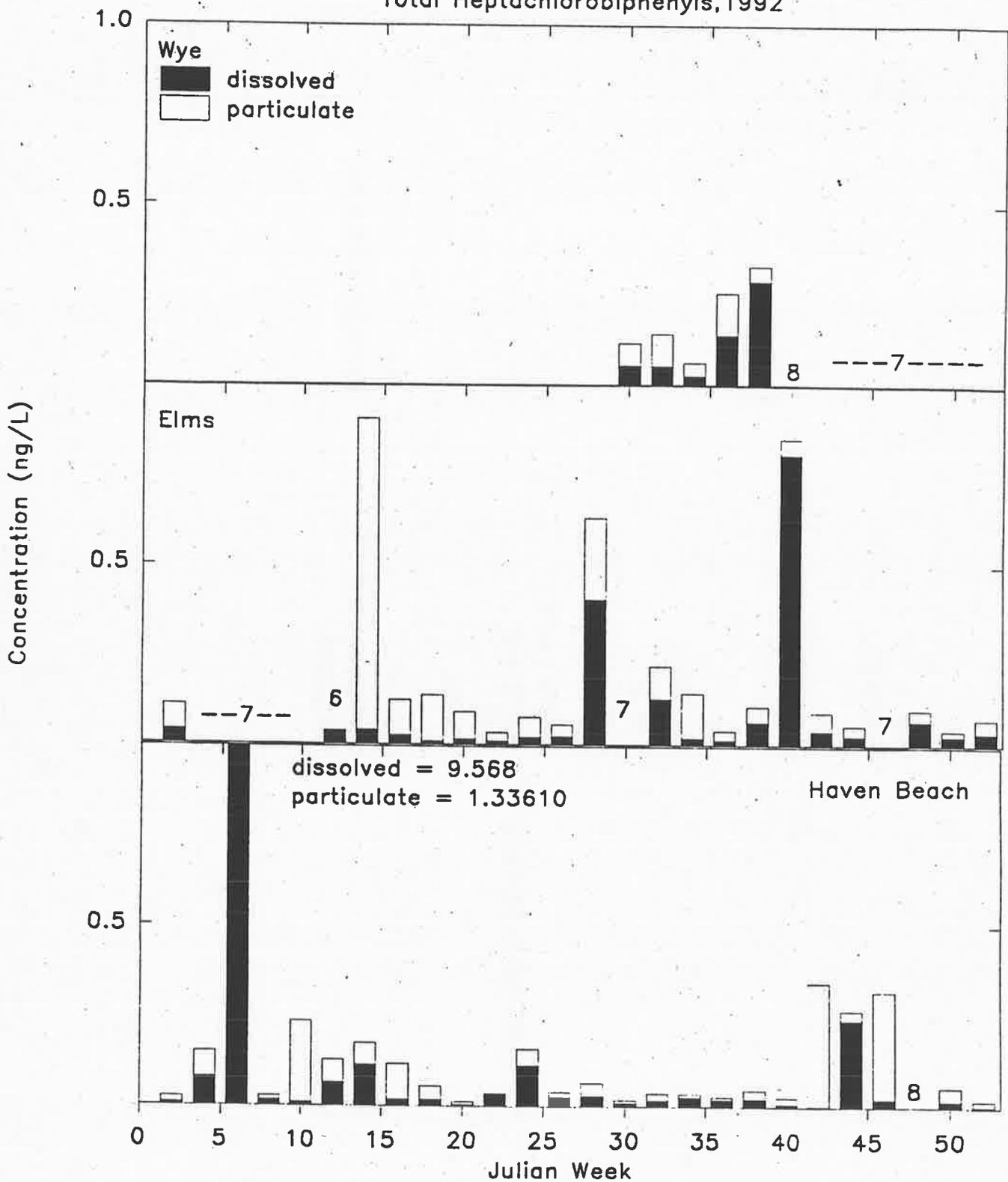


A1.43 Total hexachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Hexachlorobiphenyls, 1993

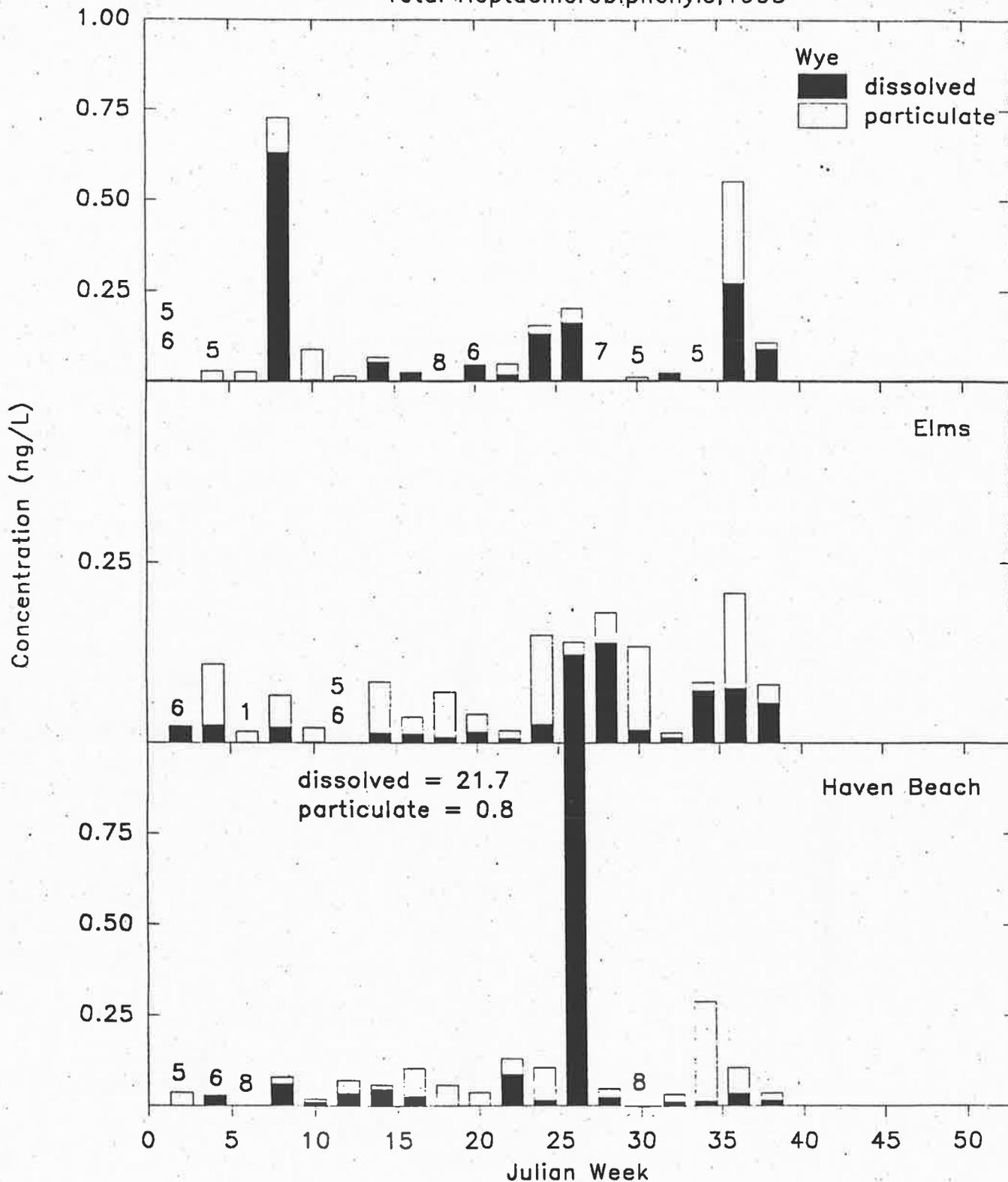


Total Heptachlorobiphenyls, 1992

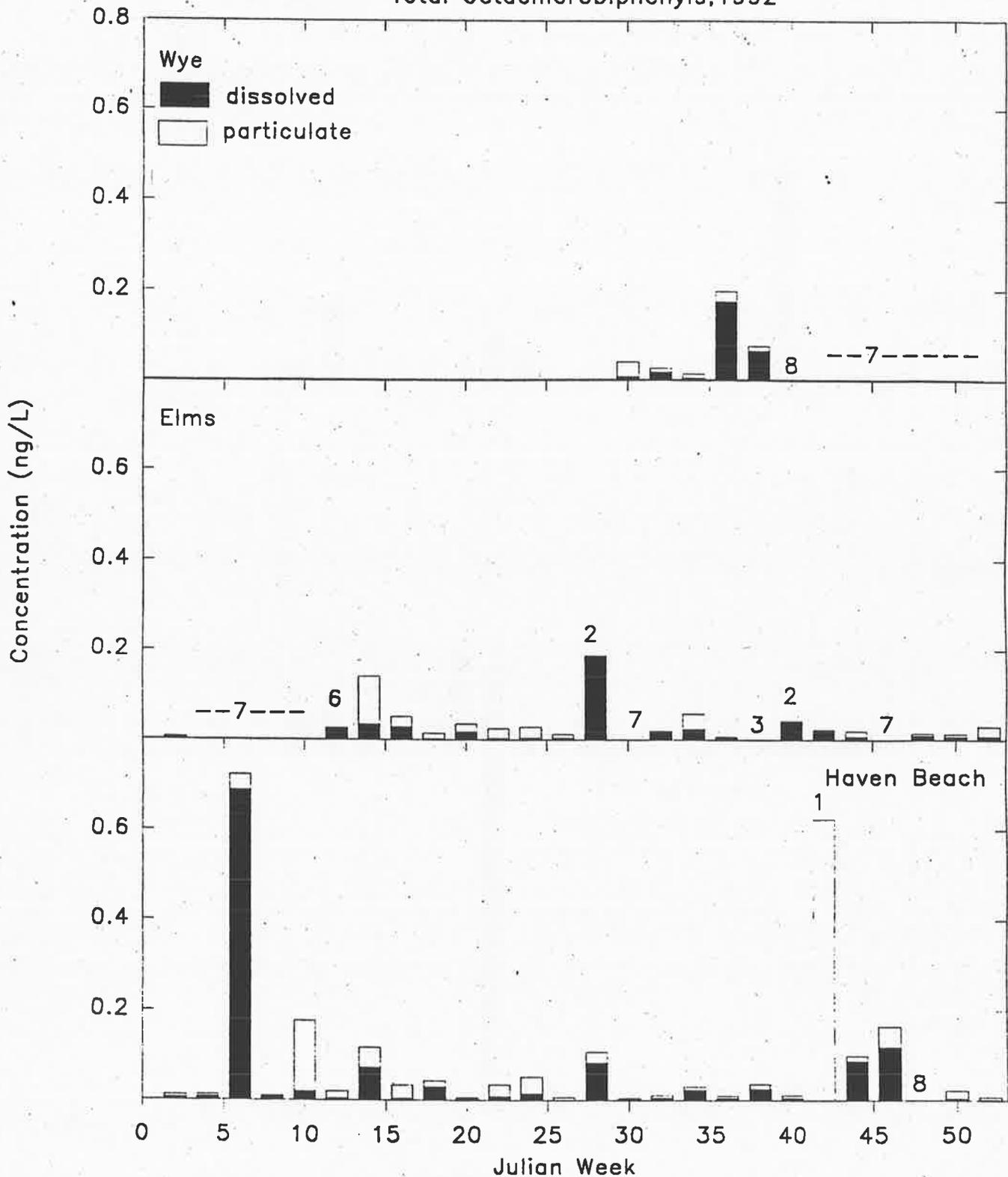


A1.44 Total heptachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Heptachlorobiphenyls, 1993

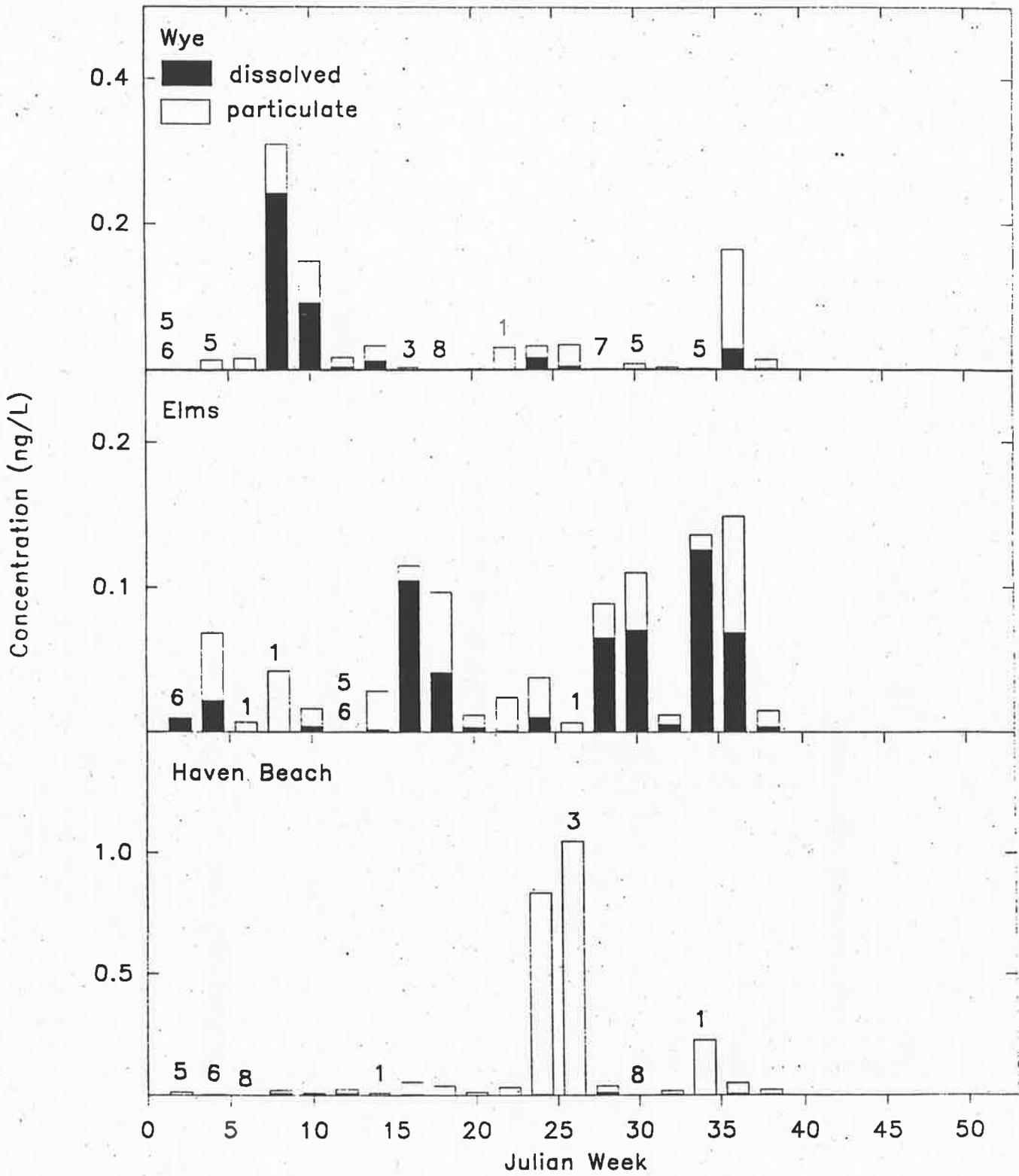


Total Octachlorobiphenyls, 1992

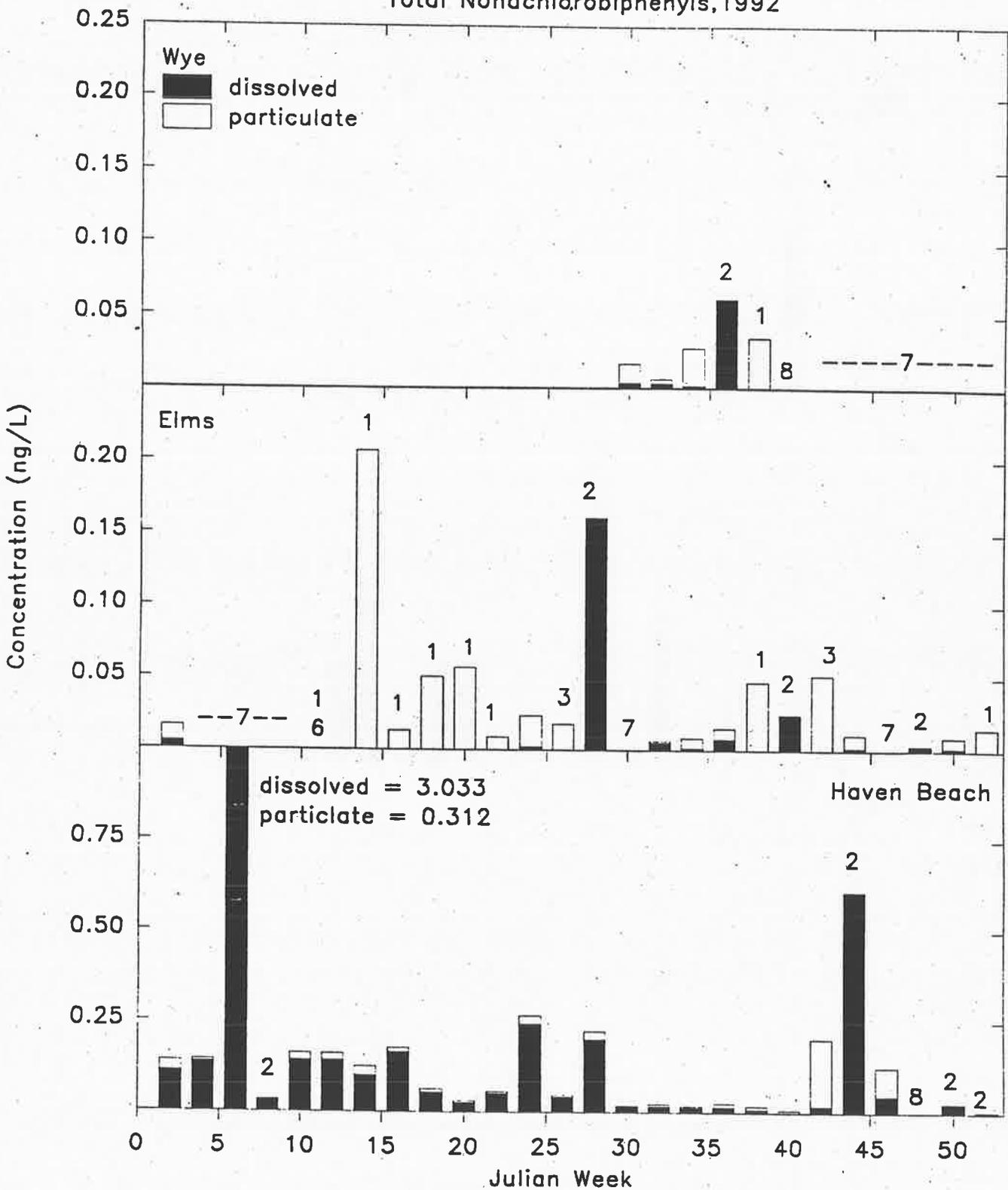


A1.45 Total octachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Octachlorobiphenyls, 1993

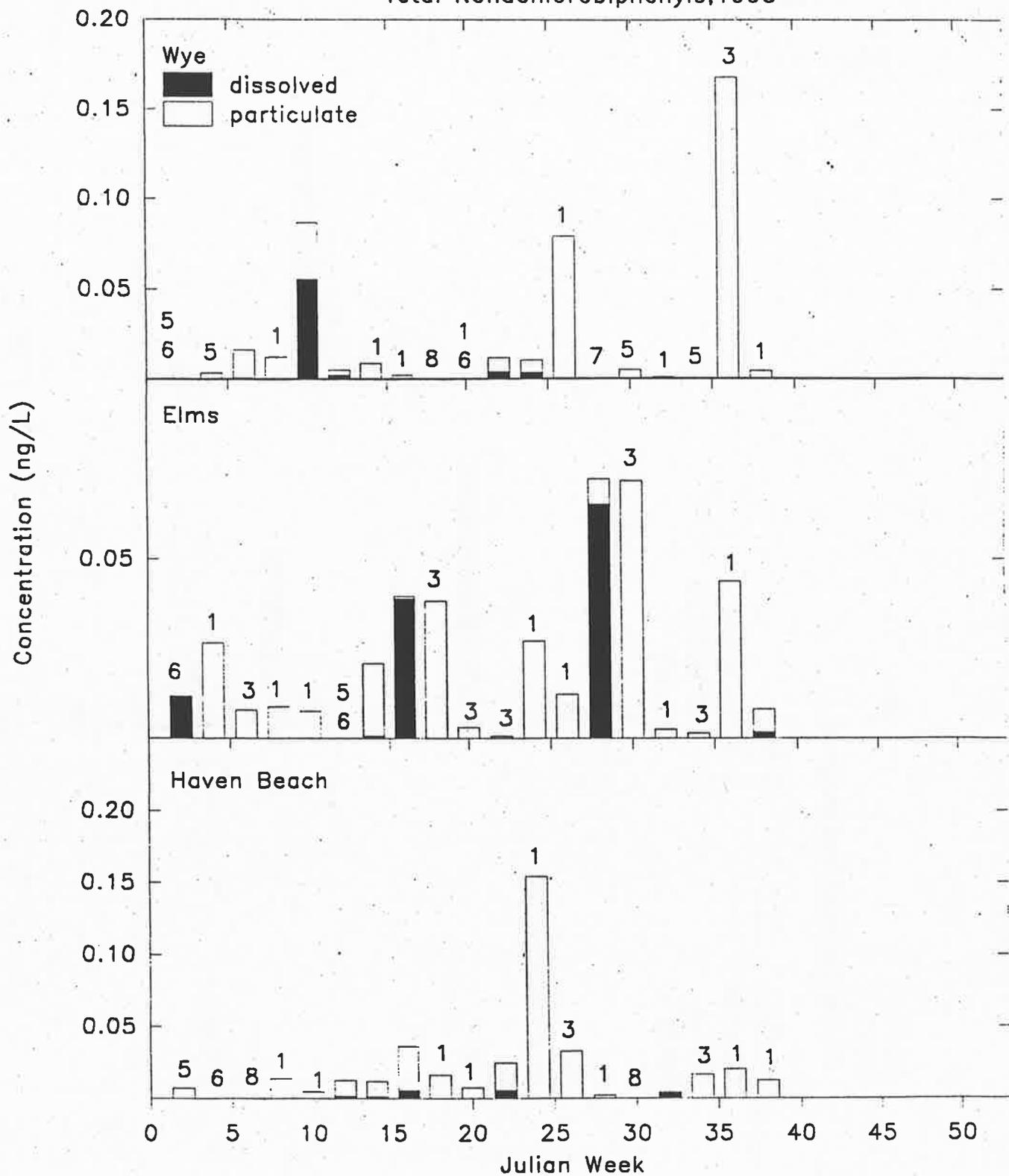


Total Nonachlorobiphenyls, 1992

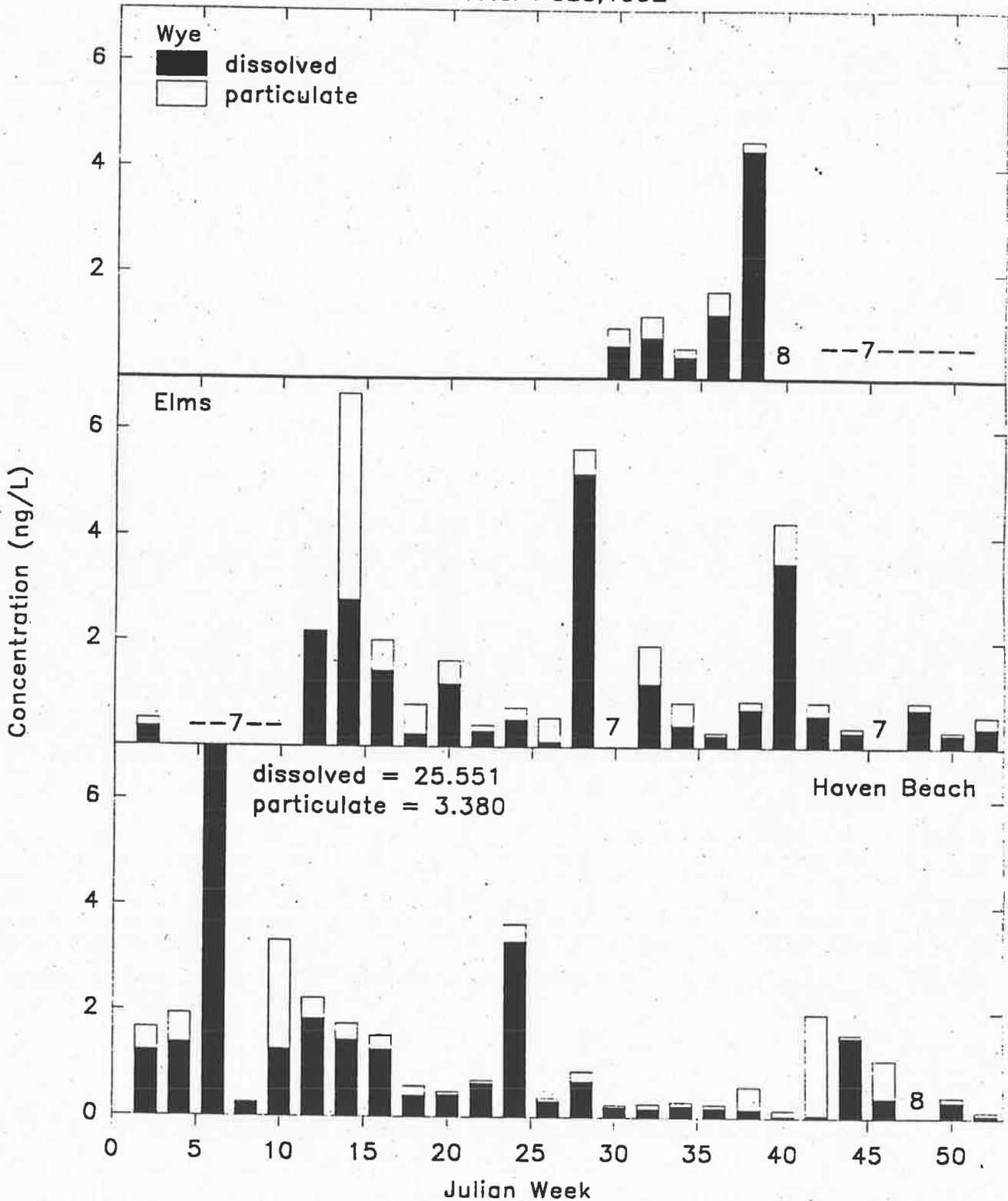


A1.46 Total nonachlorobiphenyls collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate phase, 3=not detected in dissolved phase, 4= not detected in particulate phase, 5=lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8= no ppt.).

Total Nonachlorobiphenyls, 1993

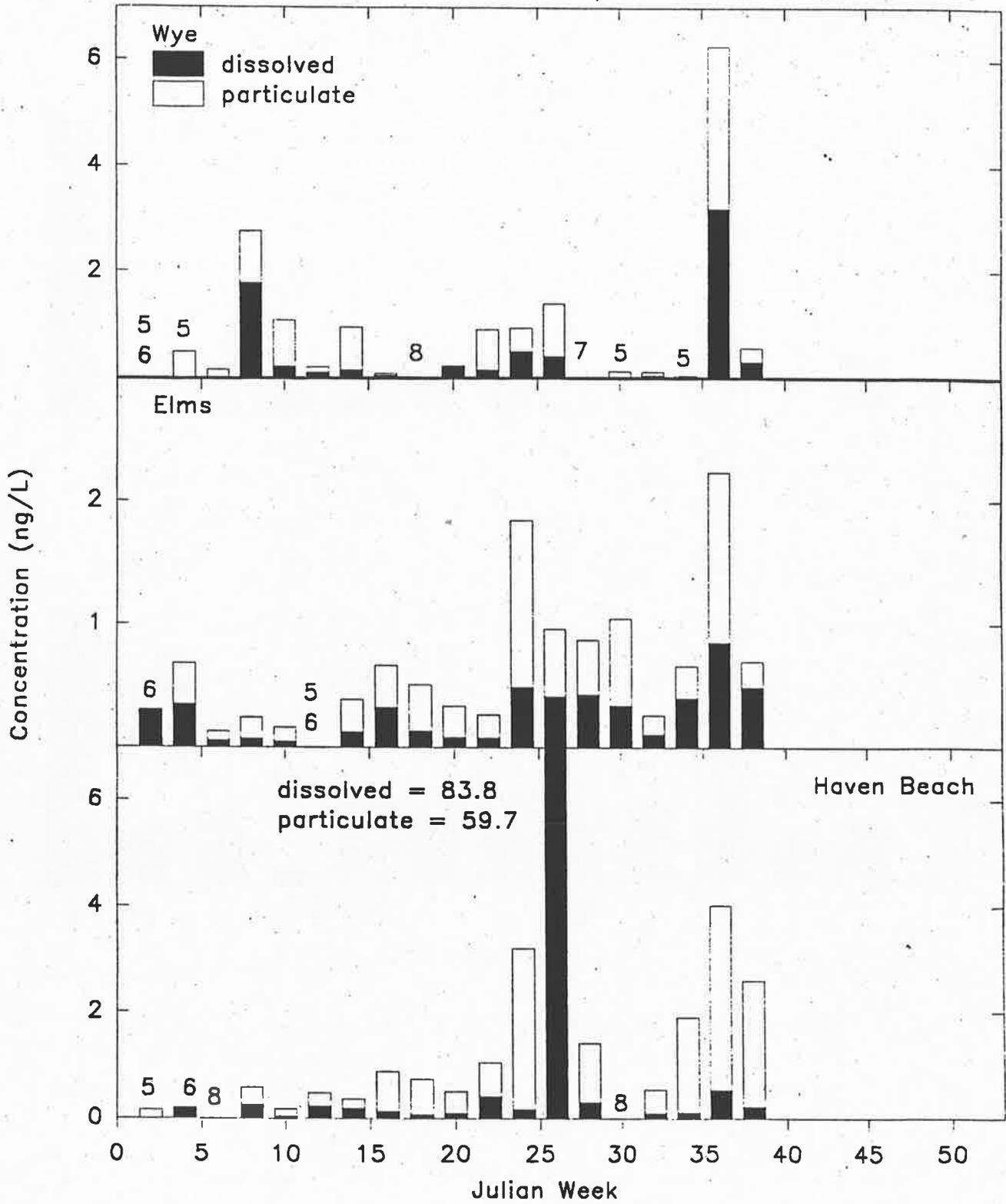


Total PCBs, 1992

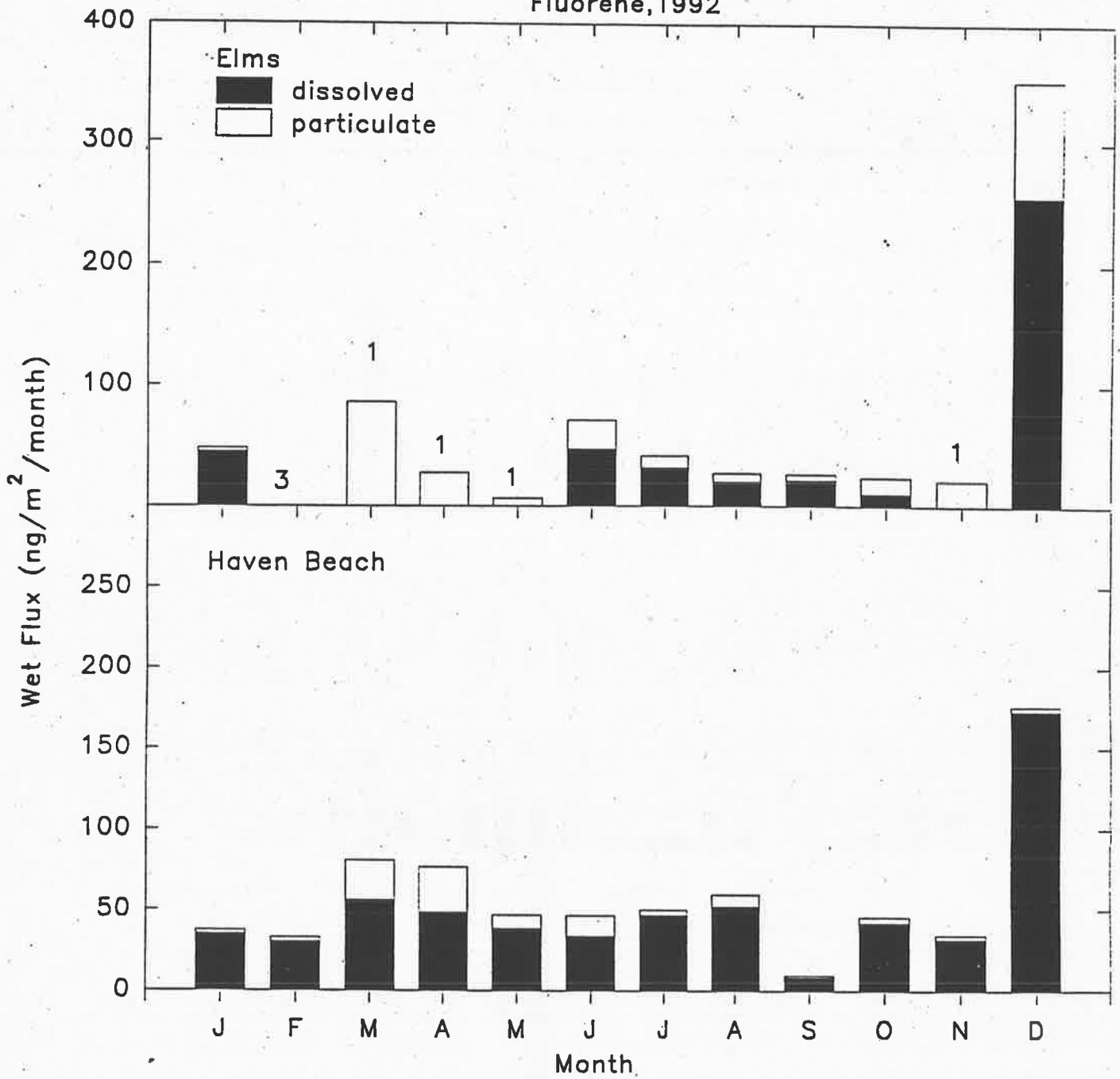


A1.47 Total PCBs collected in rain biweekly at the Wye, Elms, and Haven Beach sites. (1= not quantifiable in dissolved phase, 2= not quantifiable in particulate, 3 = not detected in dissolved phase, 4 = not detected on particulate sample, 5 = lost dissolved sample, 6=lost particulate sample, 7=sampler down, 8 = no ppt.).

Total PCBs, 1993

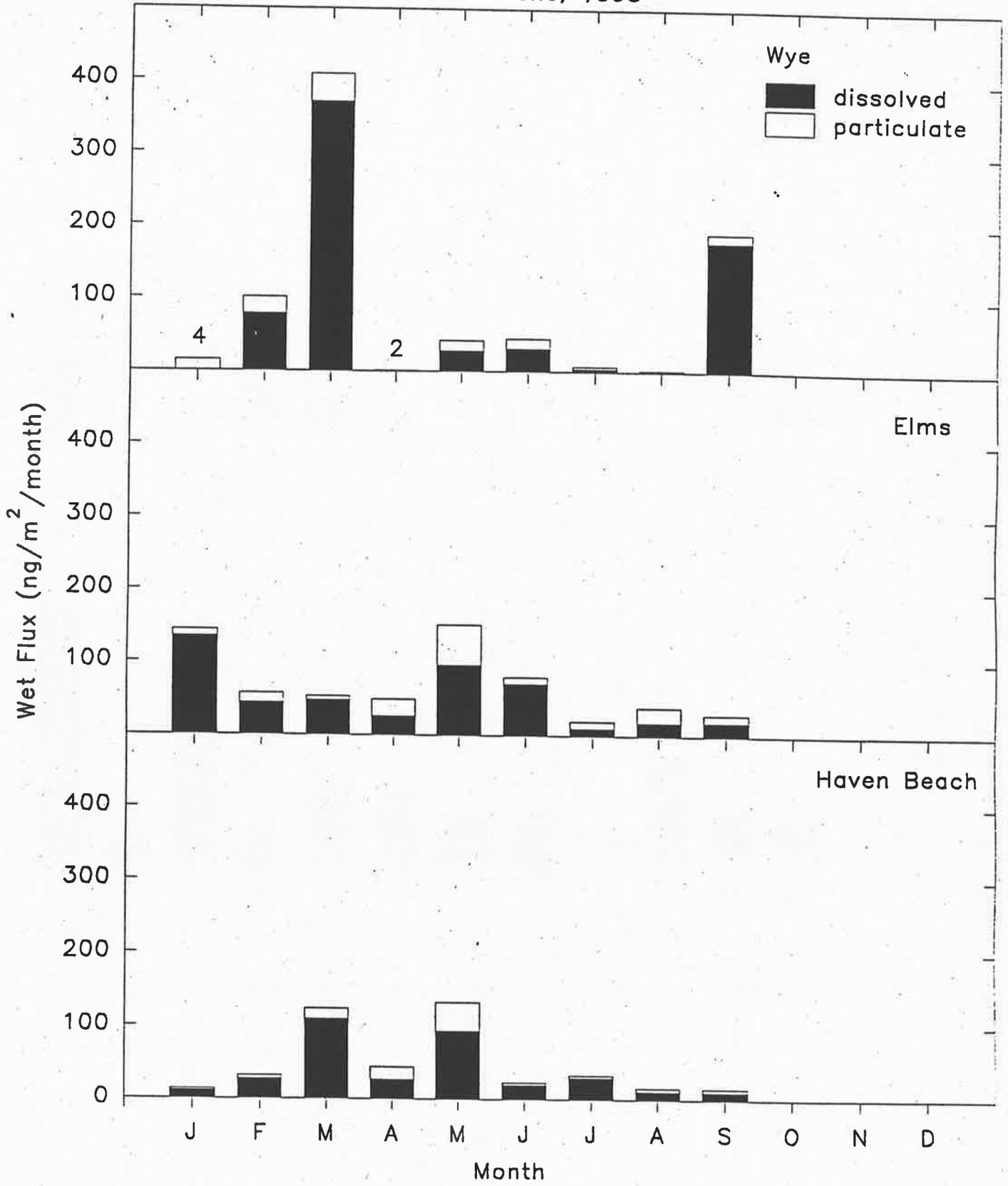


Fluorene, 1992

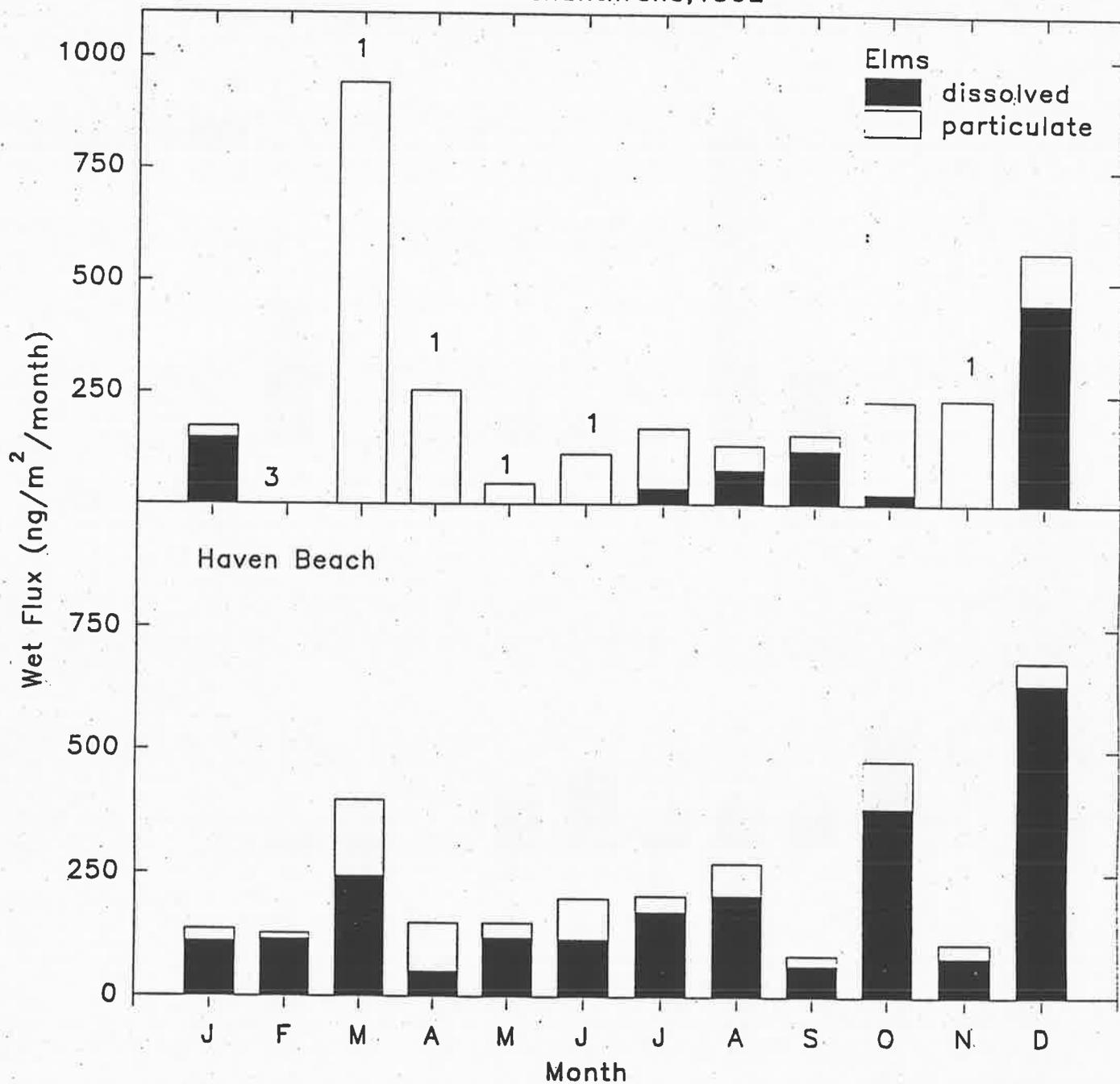


A1.48 Fluorene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Fluorene, 1993

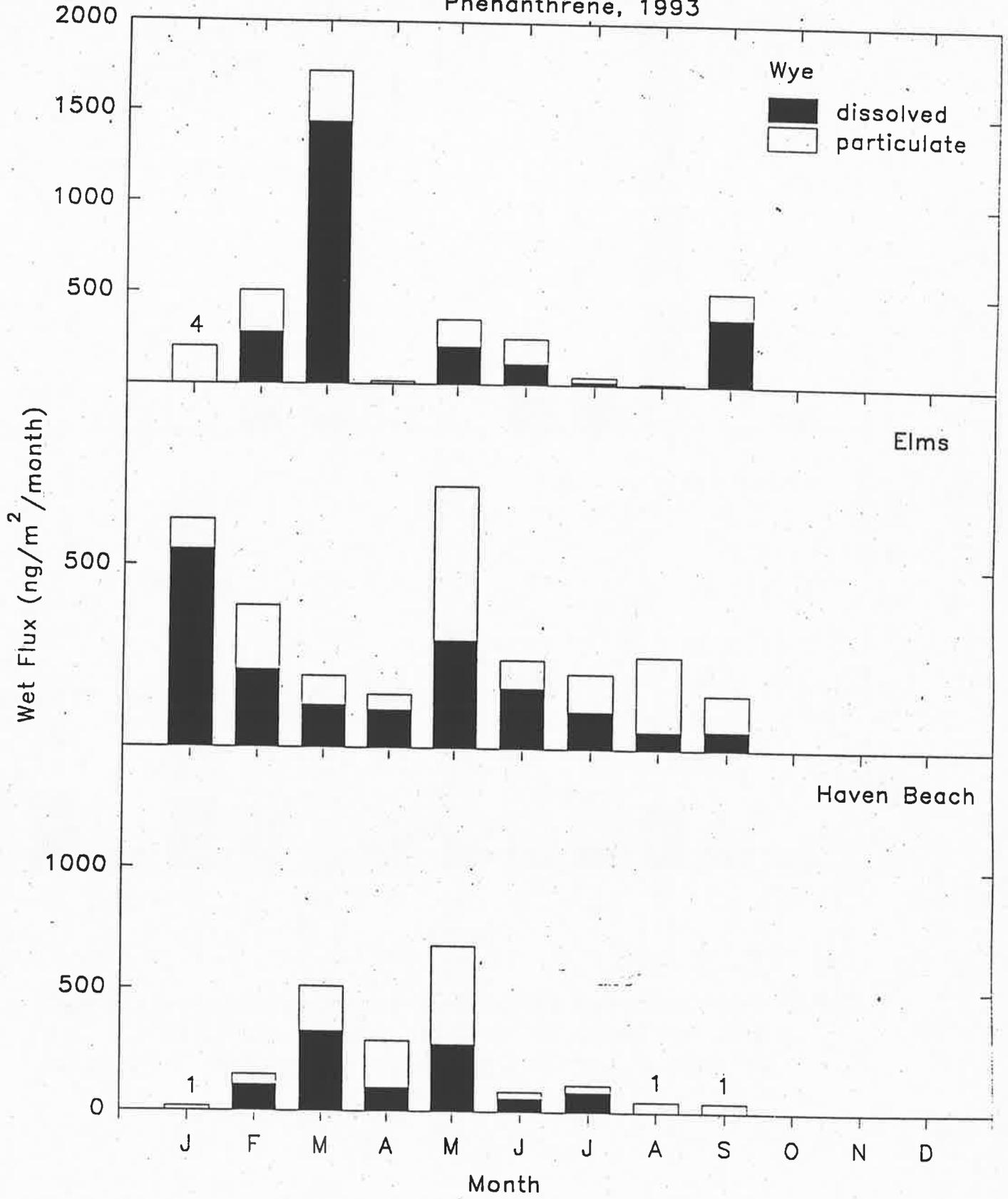


Phenanthrene, 1992

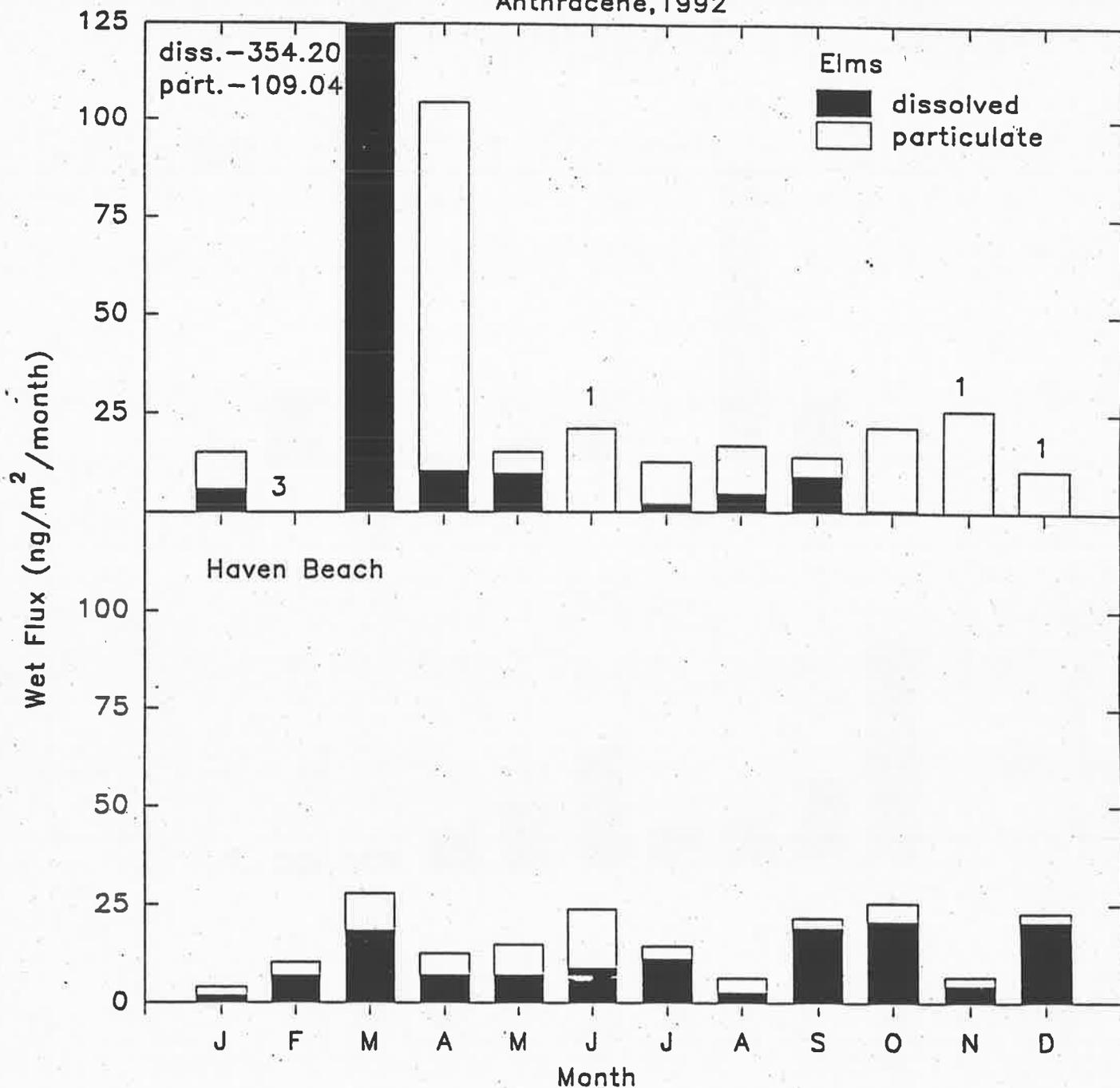


A1.49 Phenanthrene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Phenanthrene, 1993

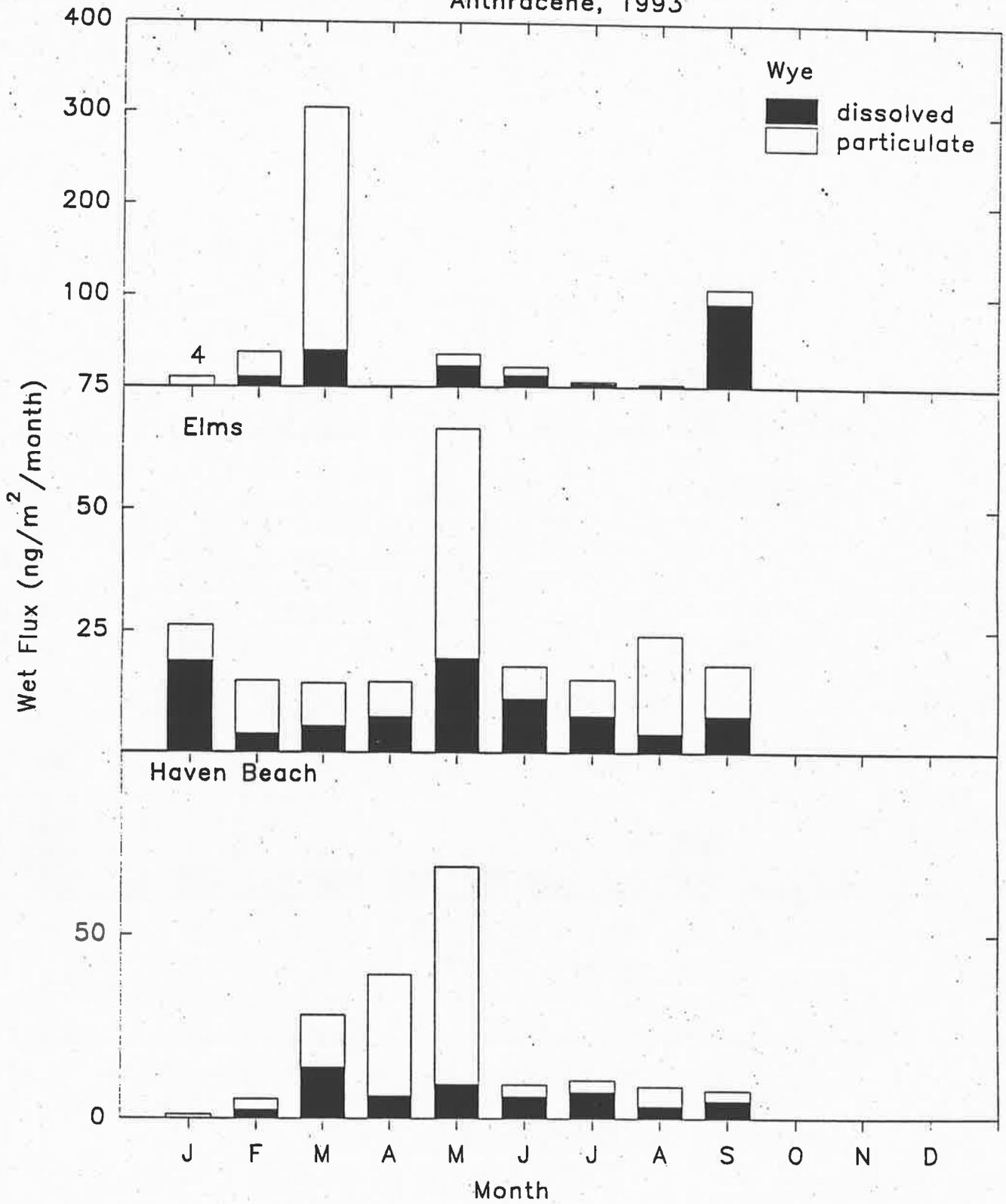


Anthracene, 1992

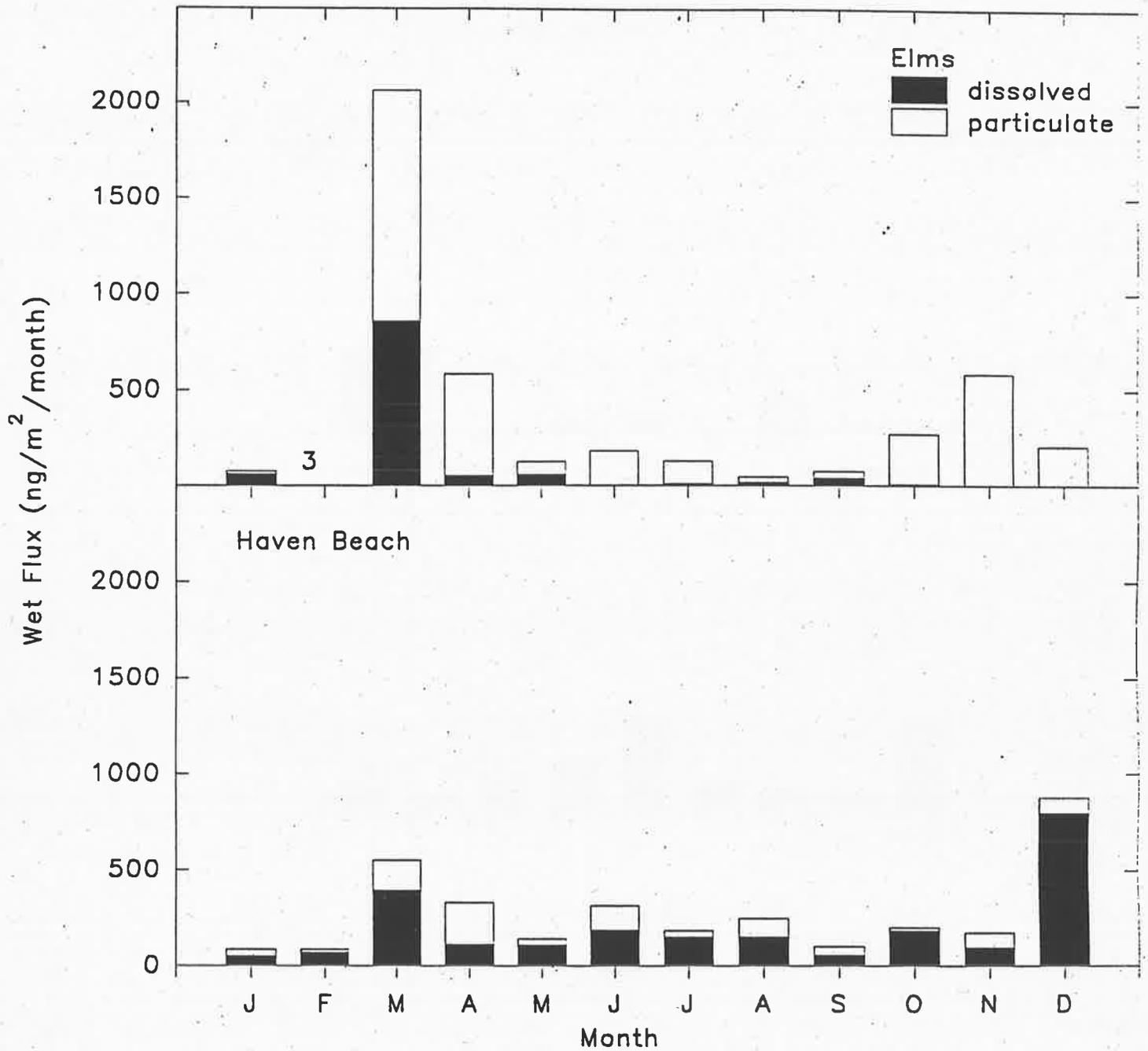


A1.50 Anthracene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Anthracene, 1993

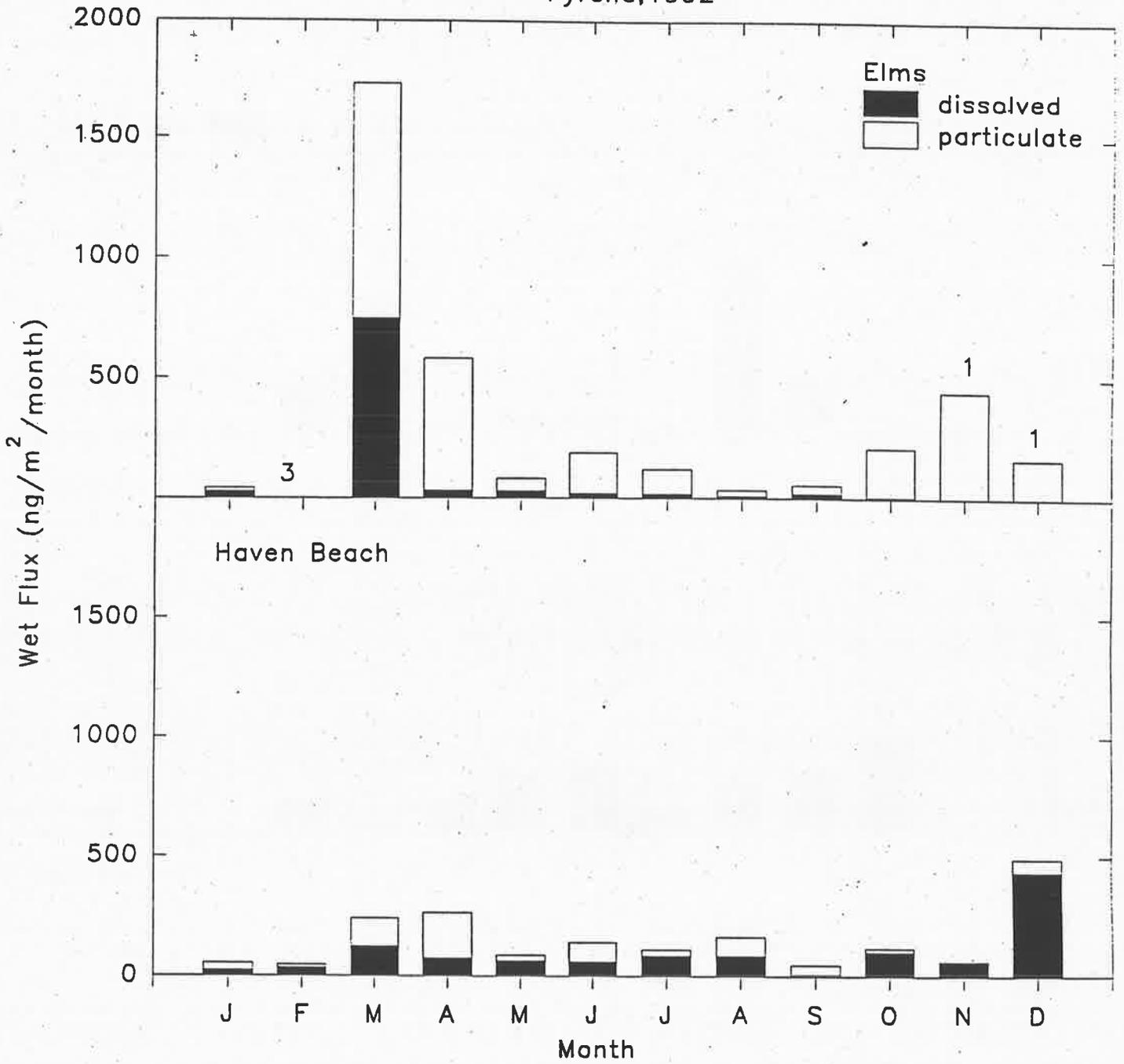


Fluoranthene, 1992



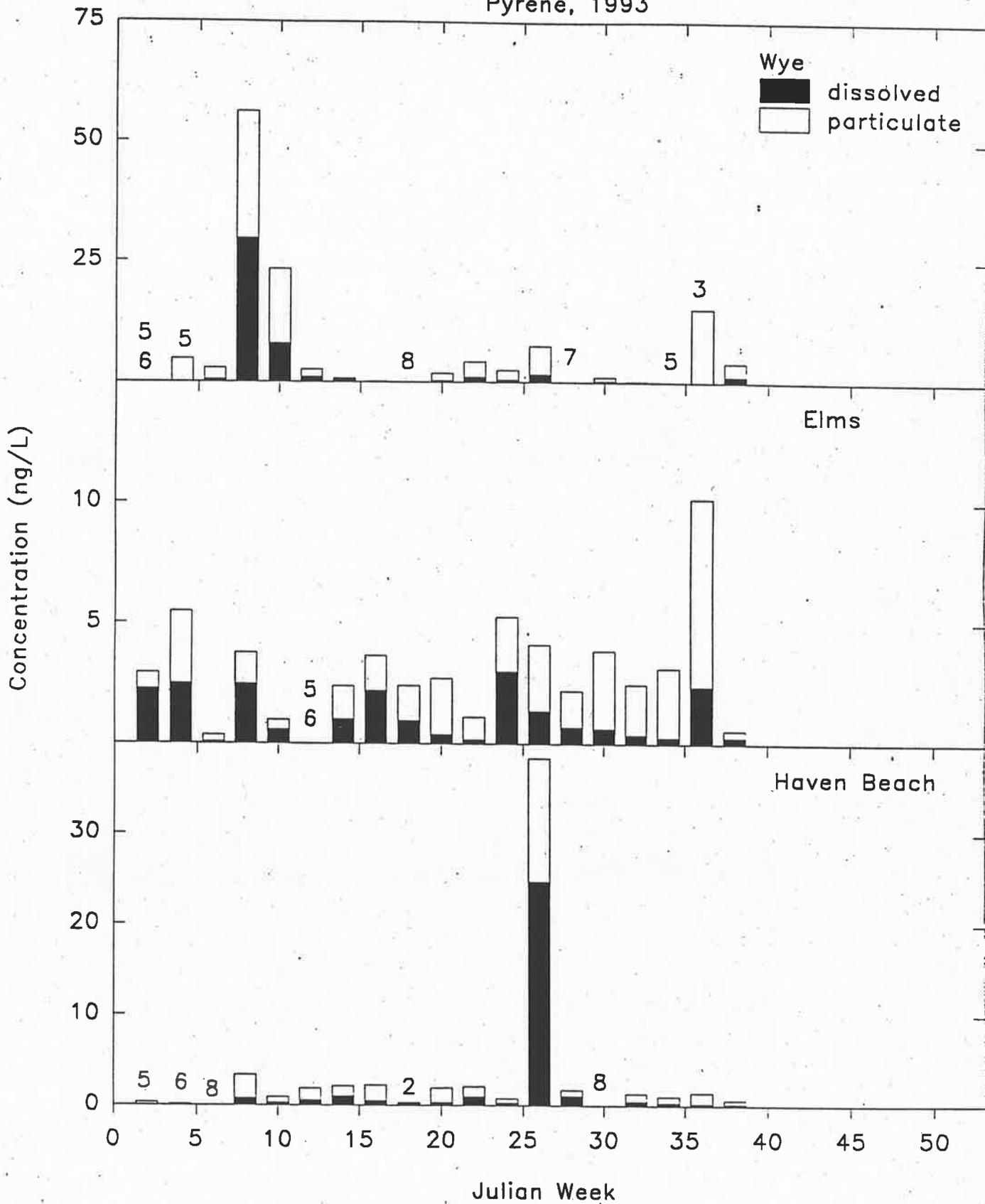
A1.51 Fluoranthene wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=sampler down, 4 = lostm dissolved sample).

Pyrene, 1992

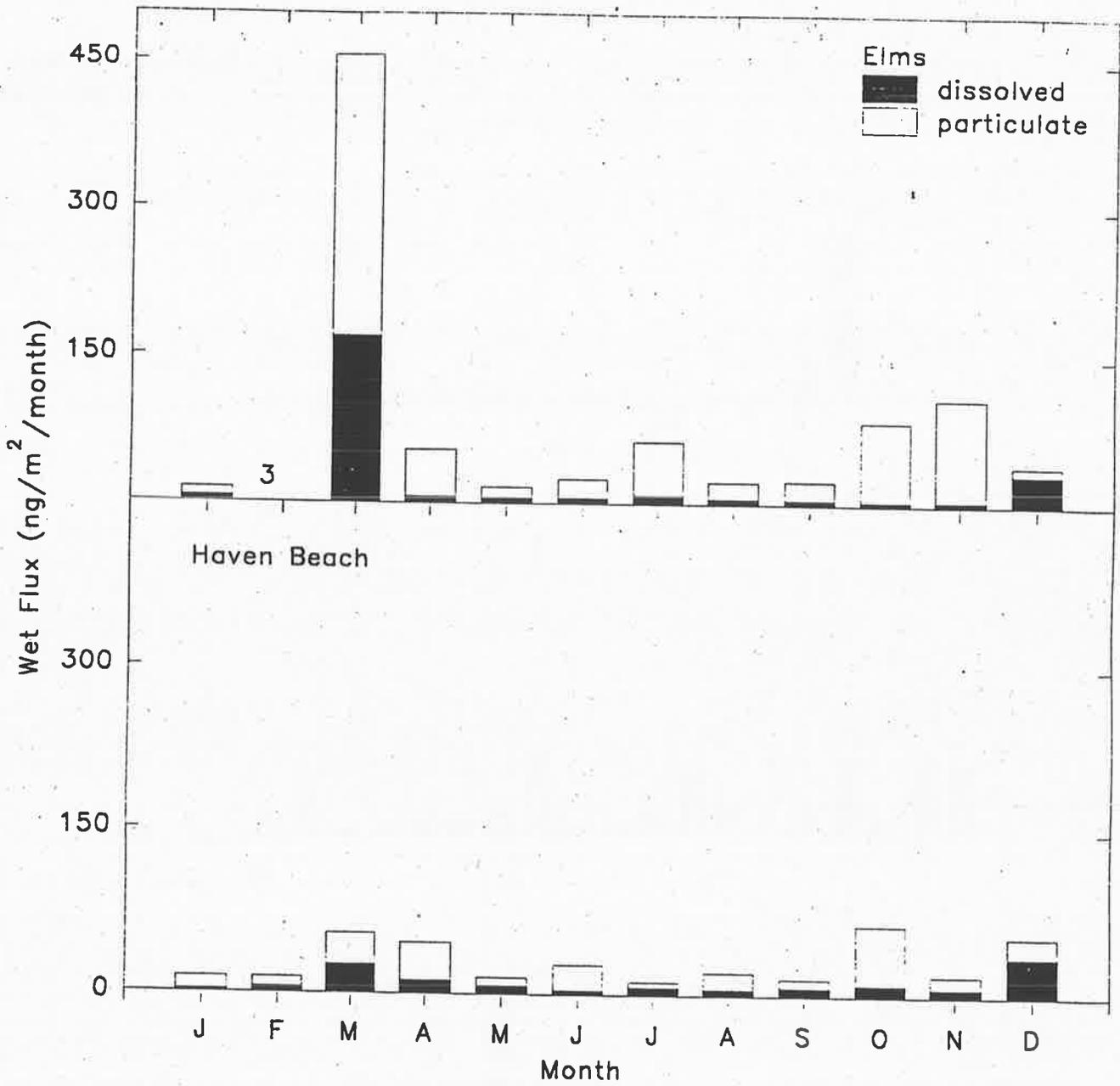


A1.52 Pyrene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Pyrene, 1993

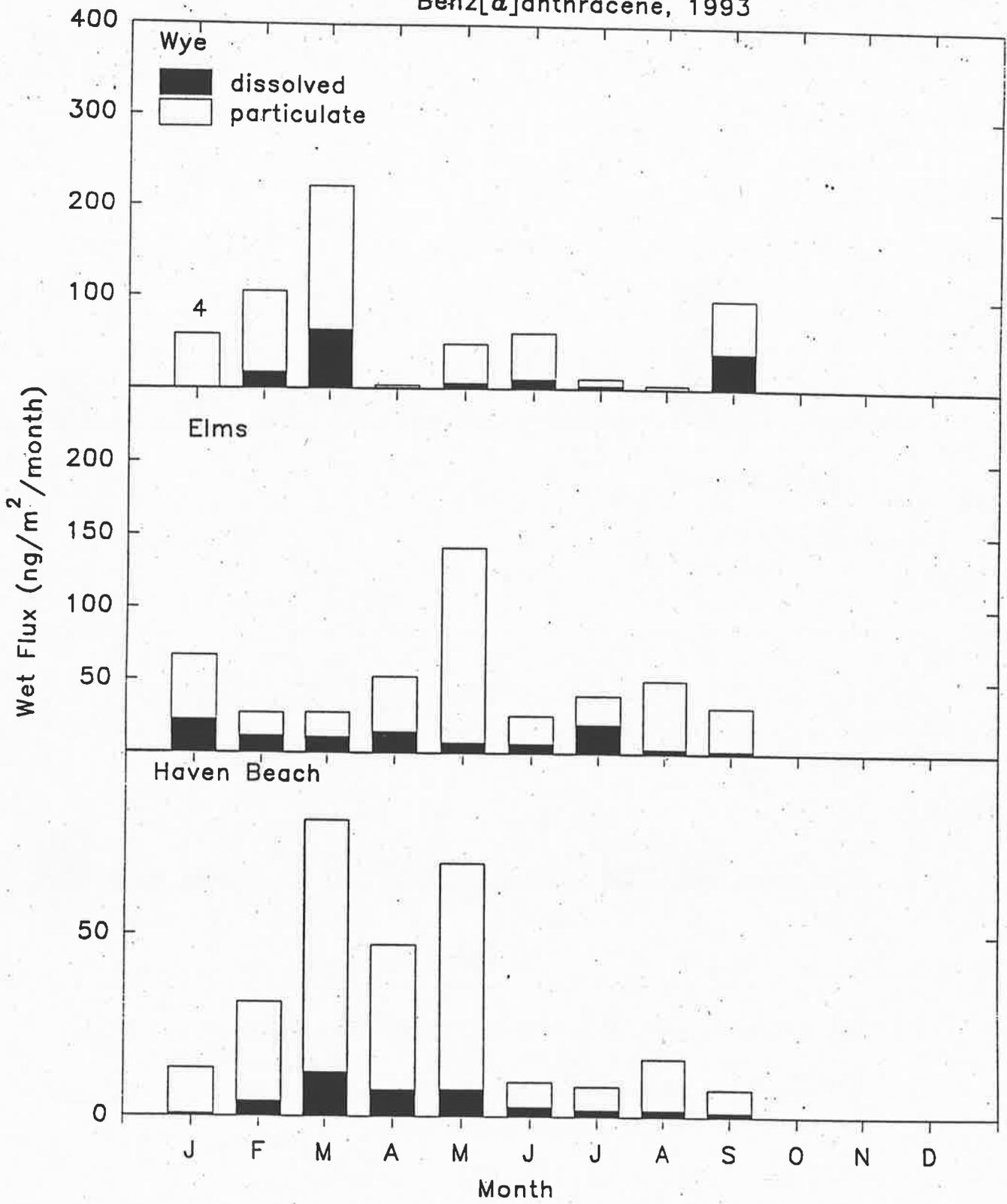


Benz[*a*]anthracene, 1992

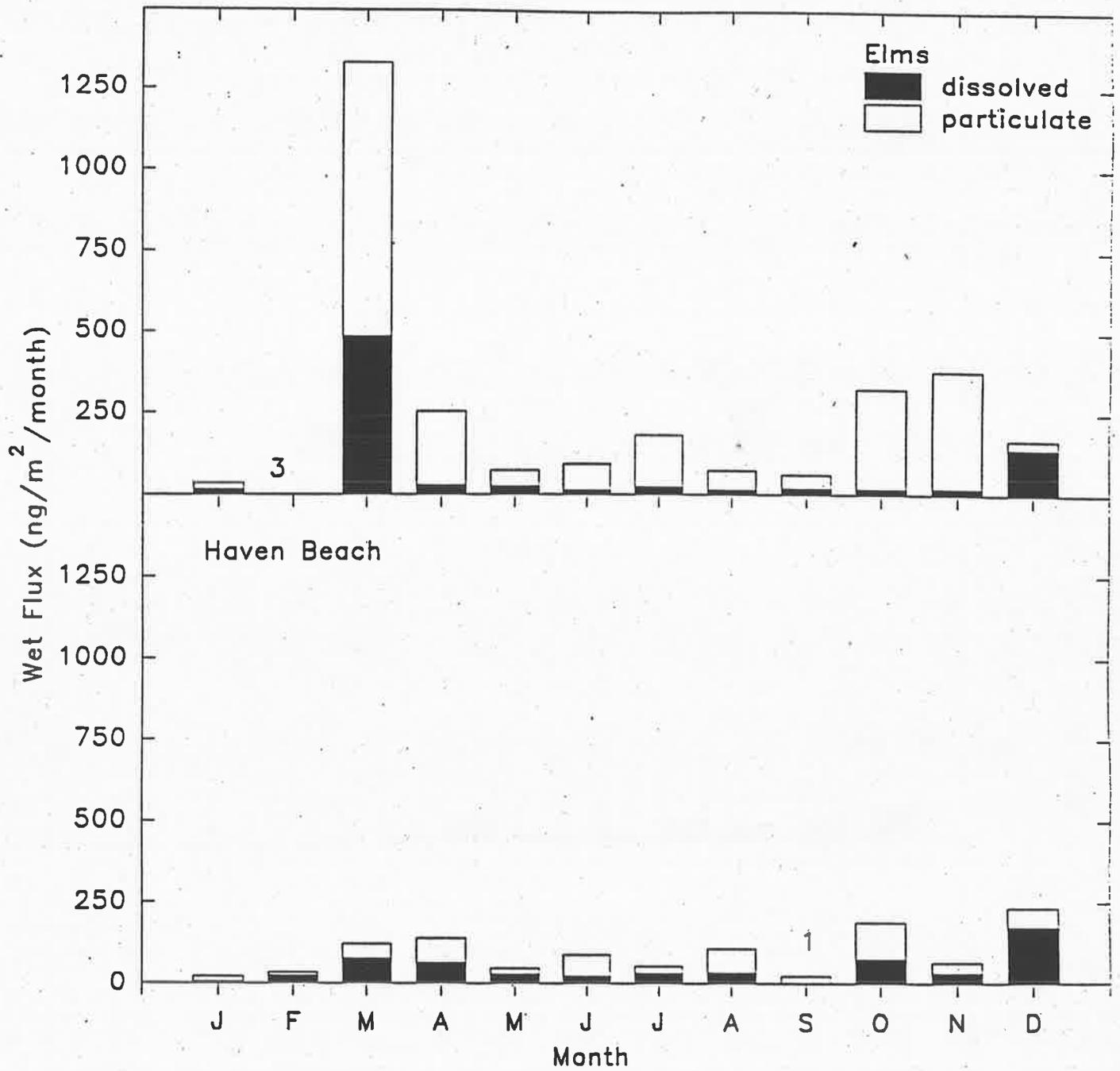


A1.53 Benz[*a*]anthracene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Benz[a]anthracene, 1993

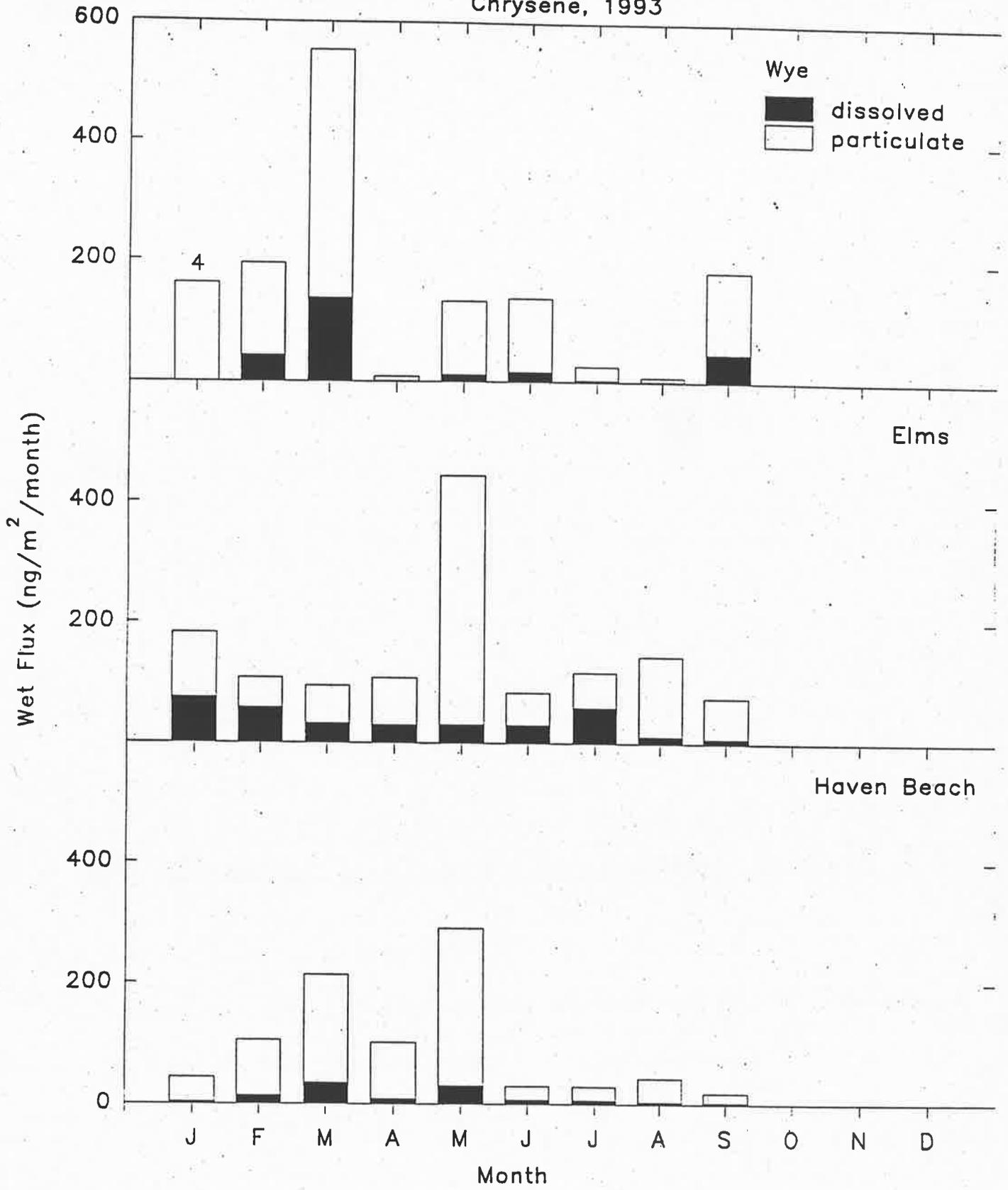


Chrysene, 1992

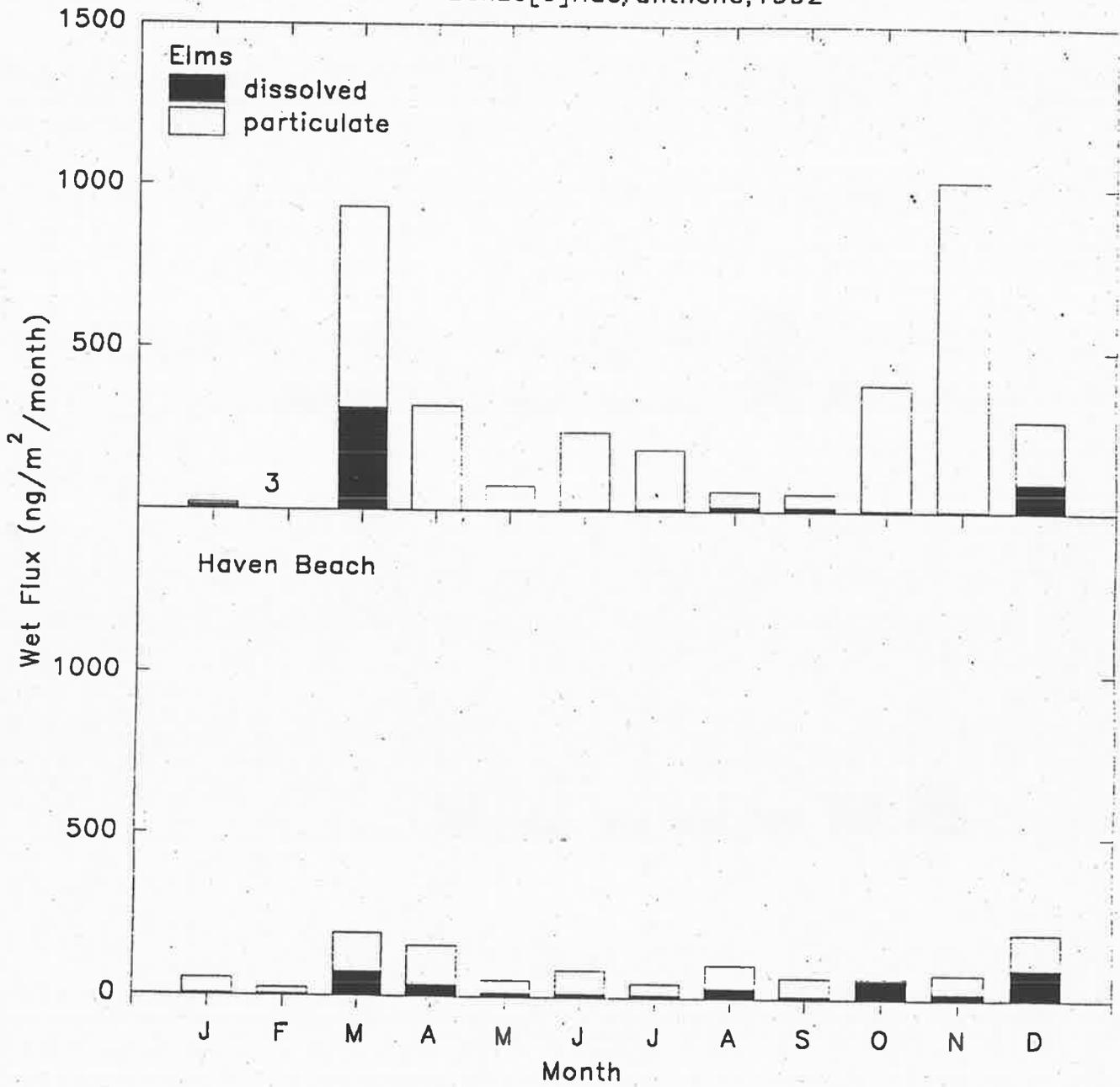


A1.54 Chrysene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Chrysene, 1993

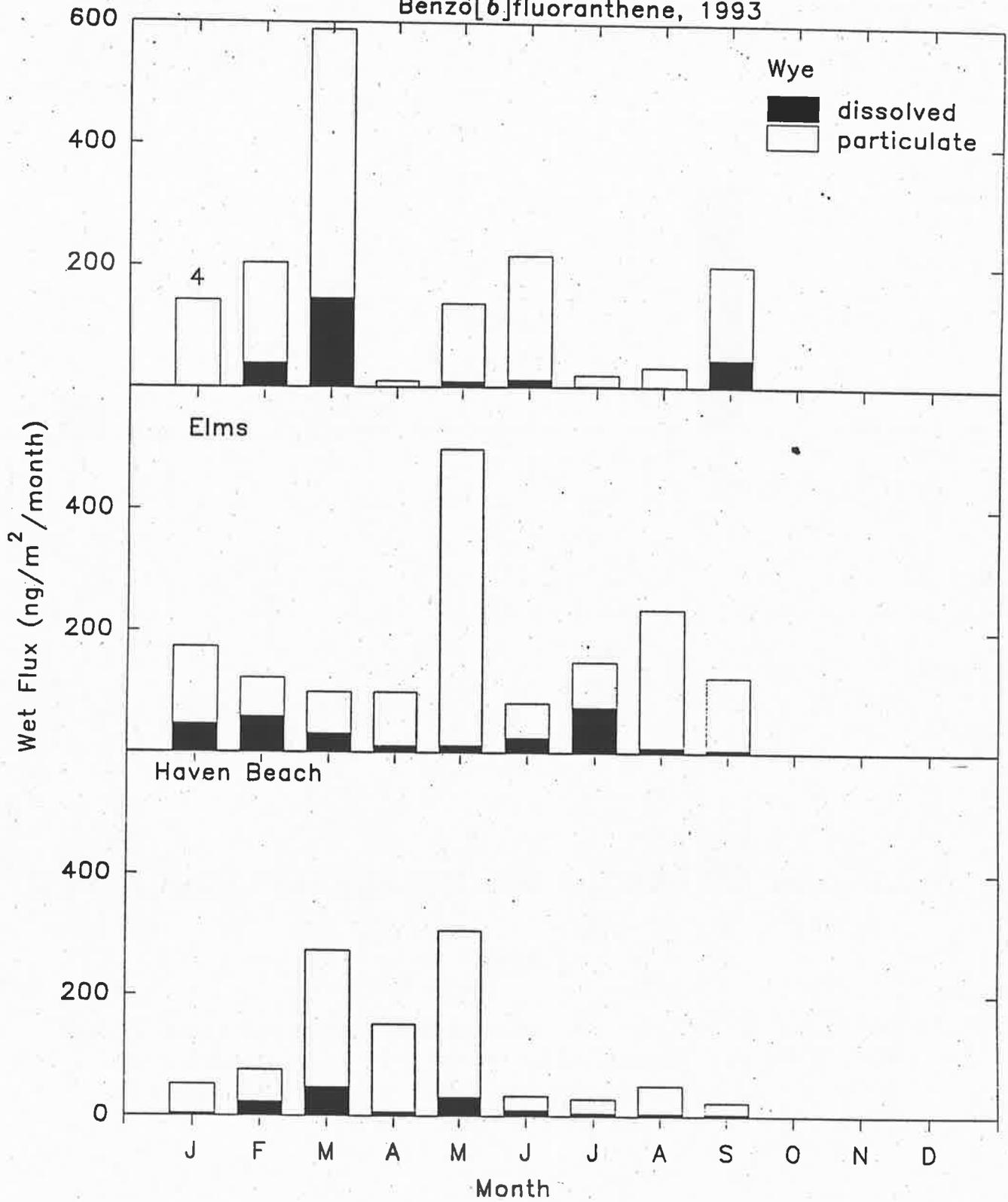


Benzo[b]fluoranthene, 1992

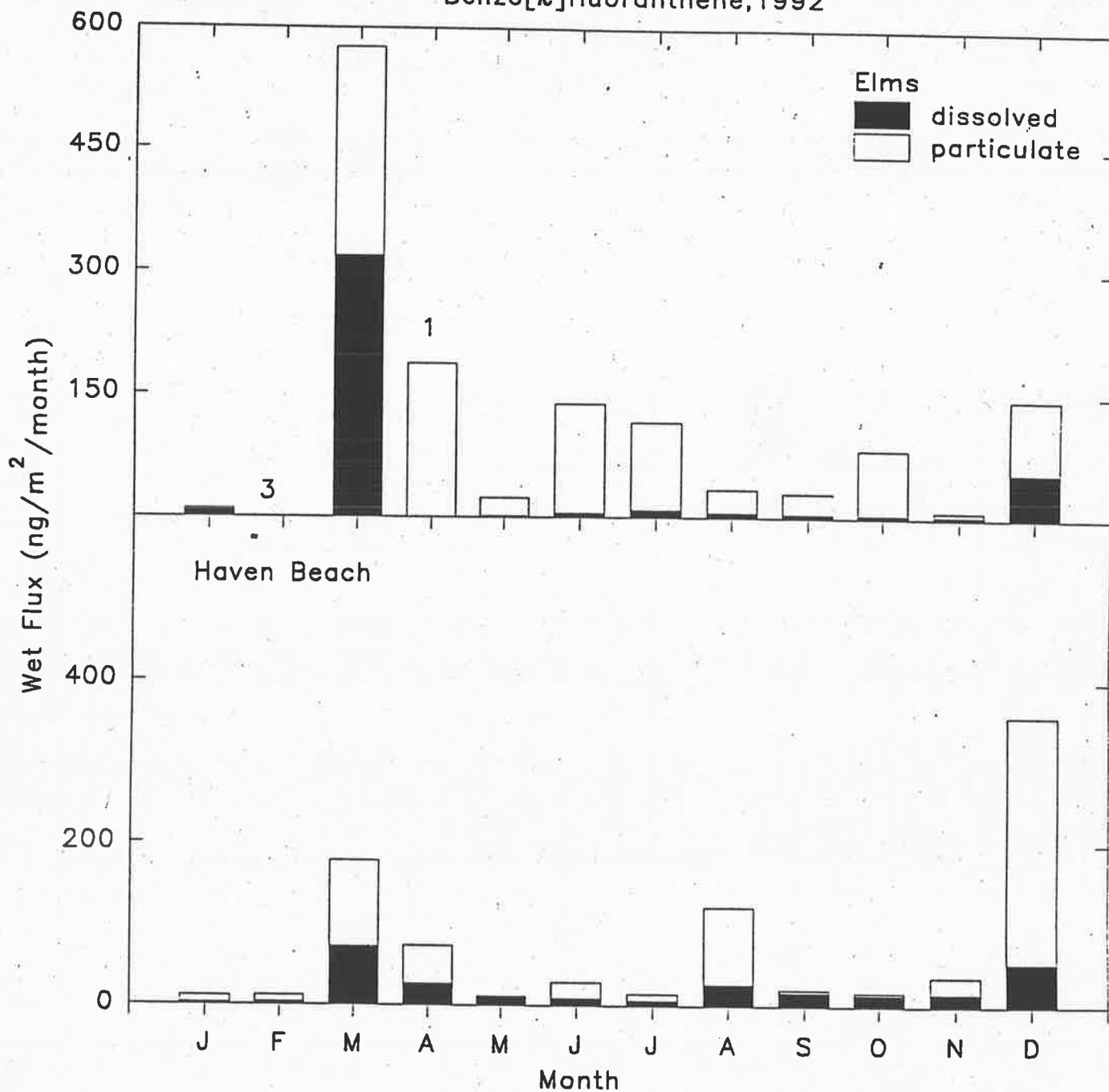


A1.55 Benzo[b]fluoranthene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Benzo[b]fluoranthene, 1993

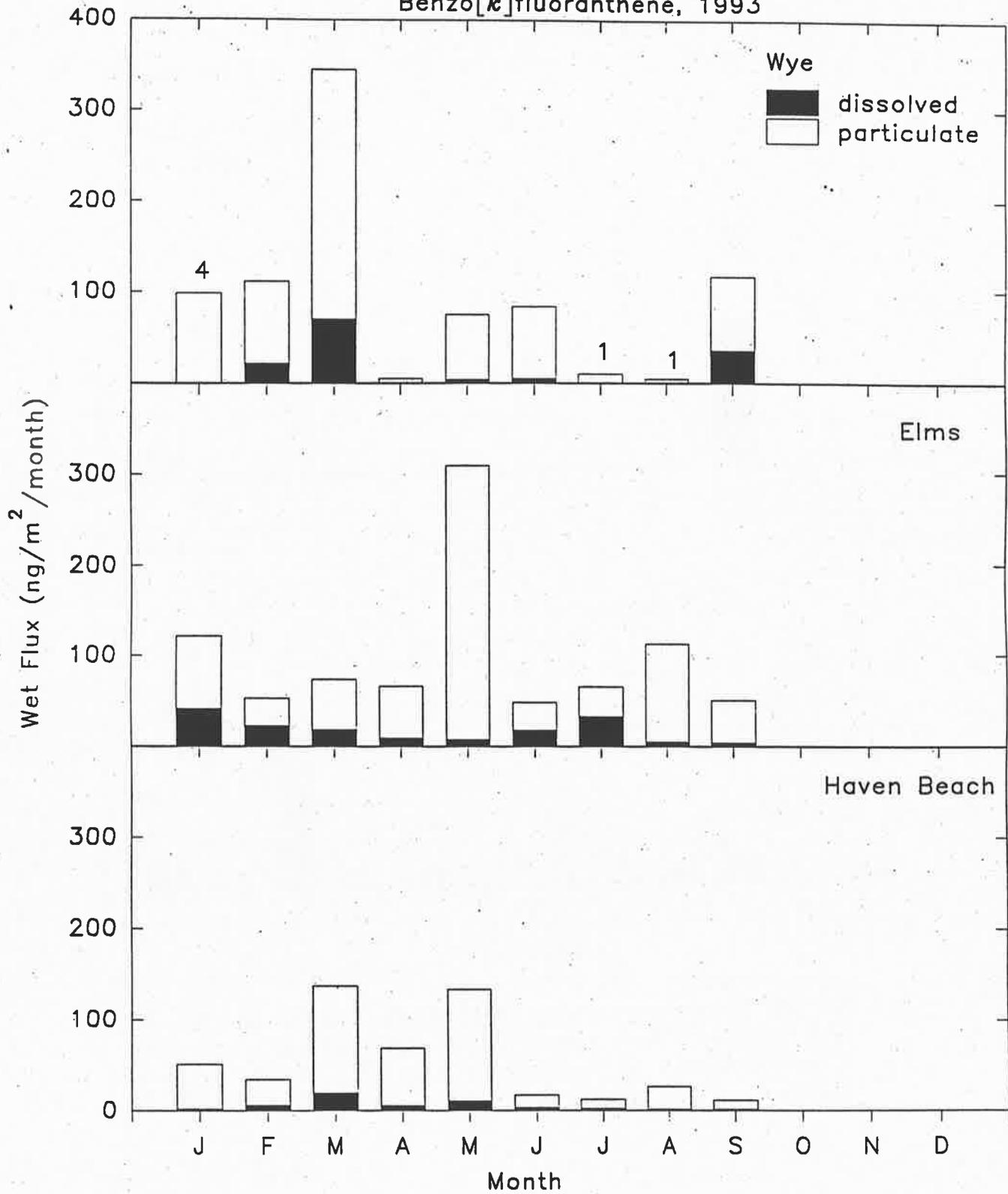


Benzo[k]fluoranthene, 1992

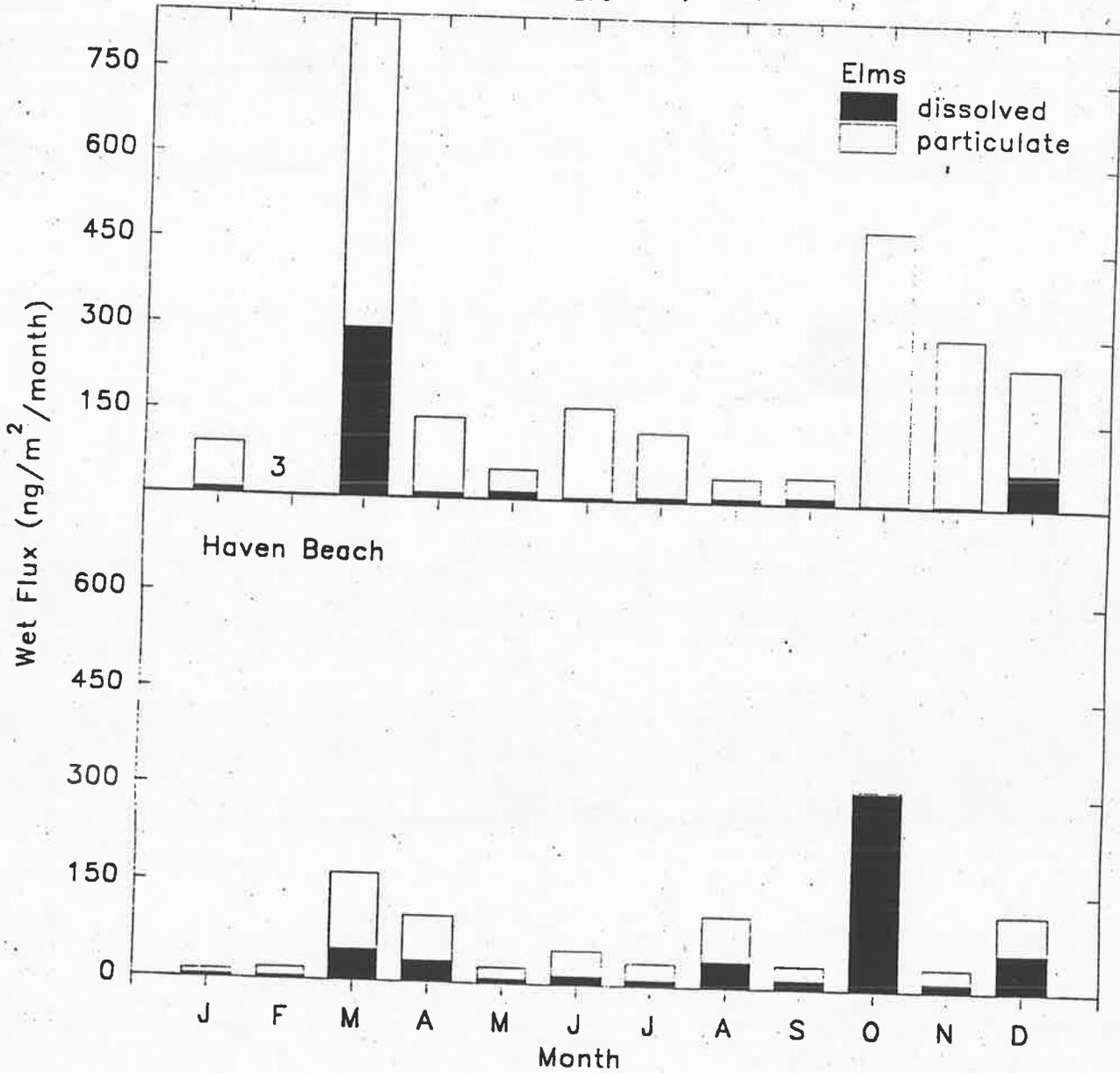


A1.56 Benzo[k]fluoranthene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Benzo[*k*]fluoranthene, 1993

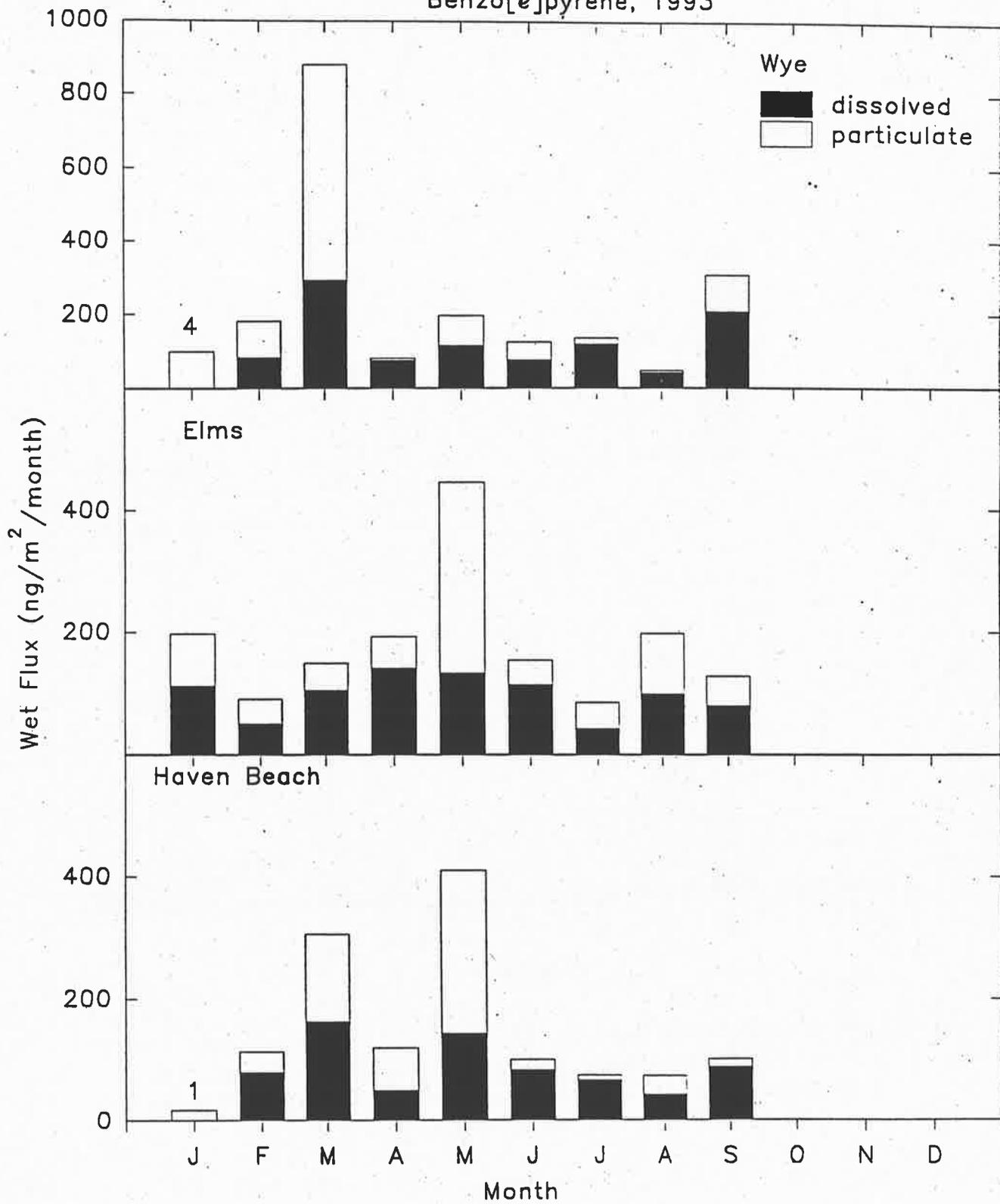


Benzo[e]pyrene, 1992

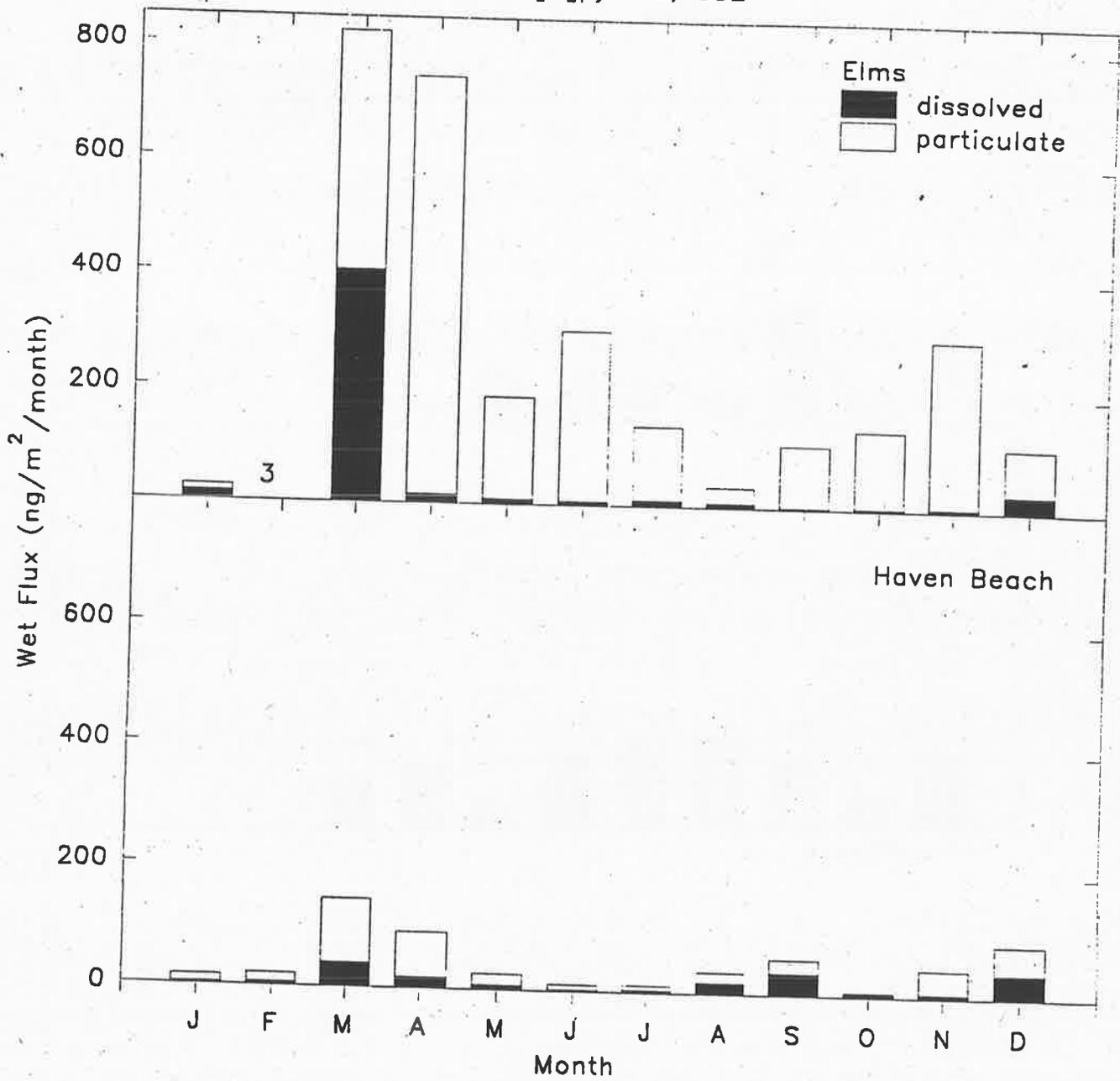


A1.57 Benzo[e]pyrene wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample.).

Benzo[e]pyrene, 1993

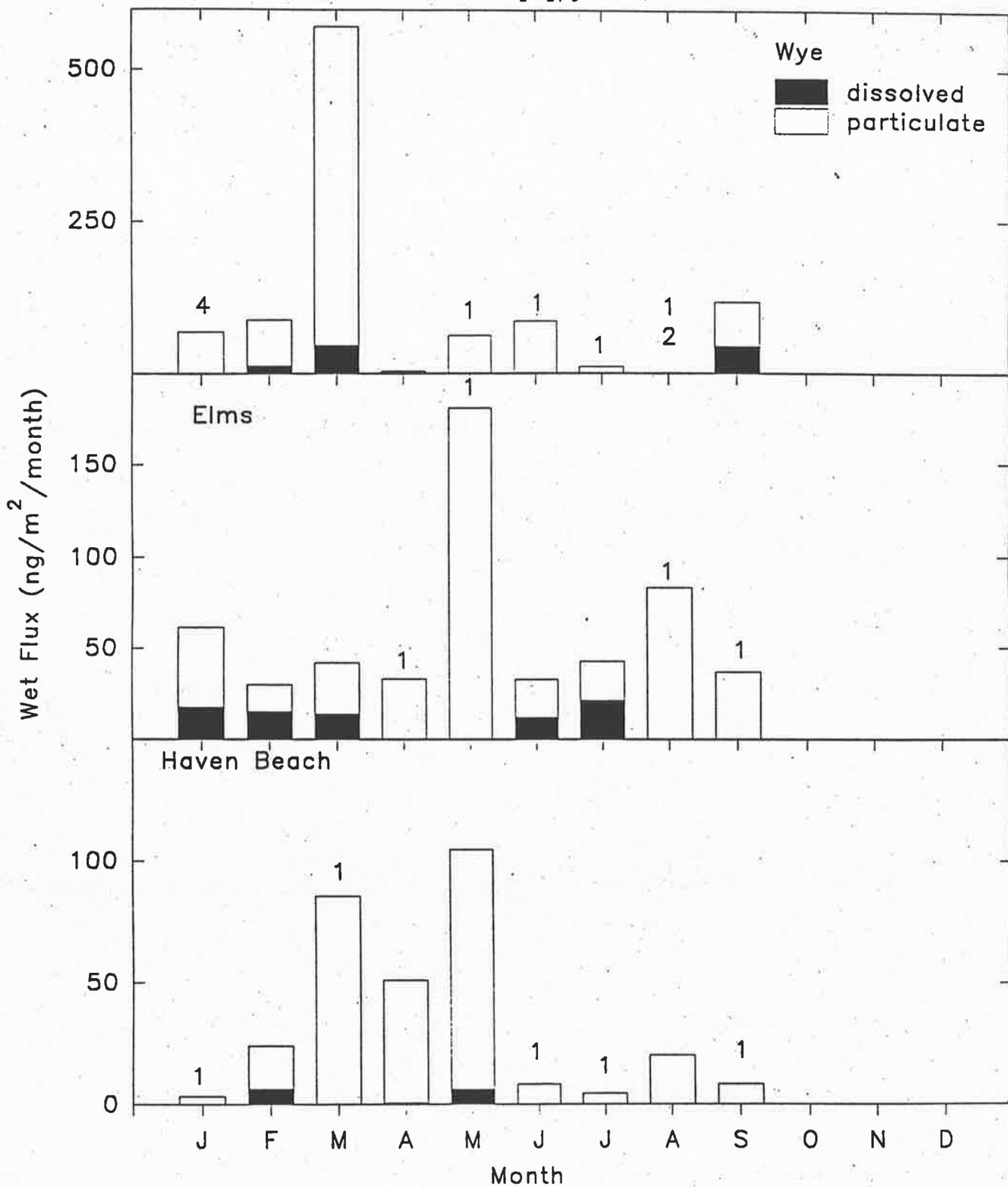


Benzo[a]pyrene, 1992

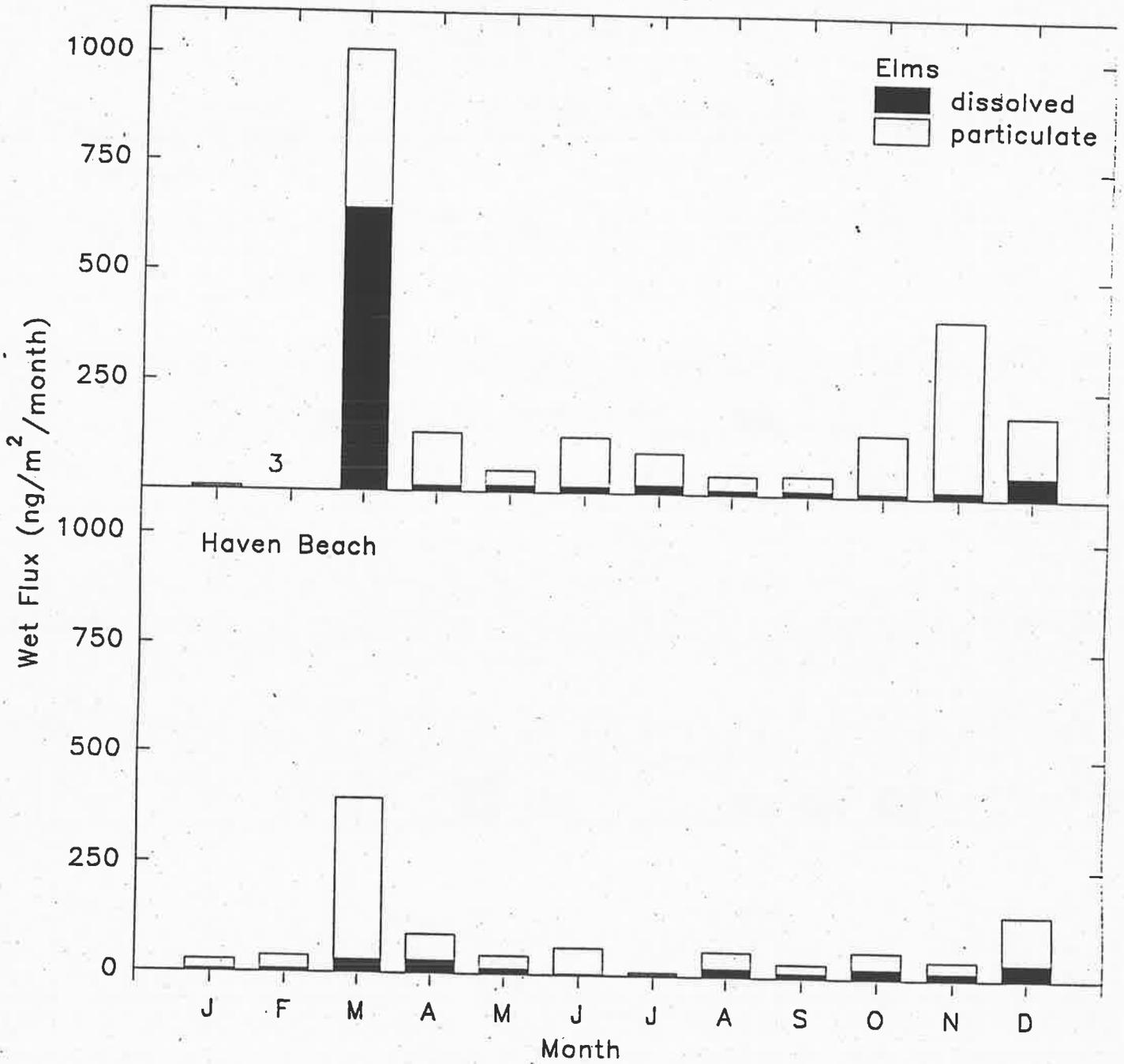


A1.58 Benzo[a]pyrene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Benzo[a]pyrene, 1993

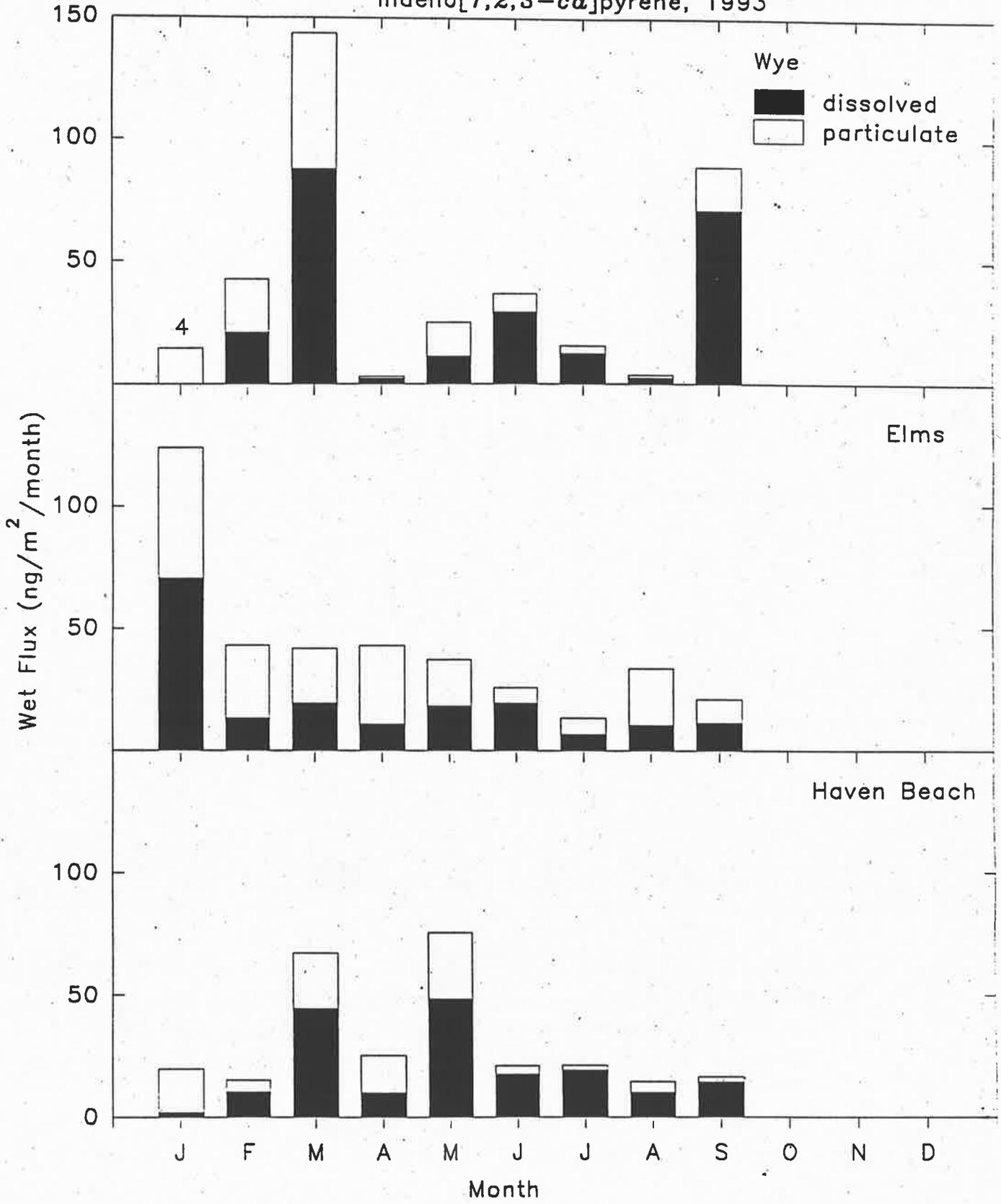


Indeno[1,2,3-cd]pyrene, 1992

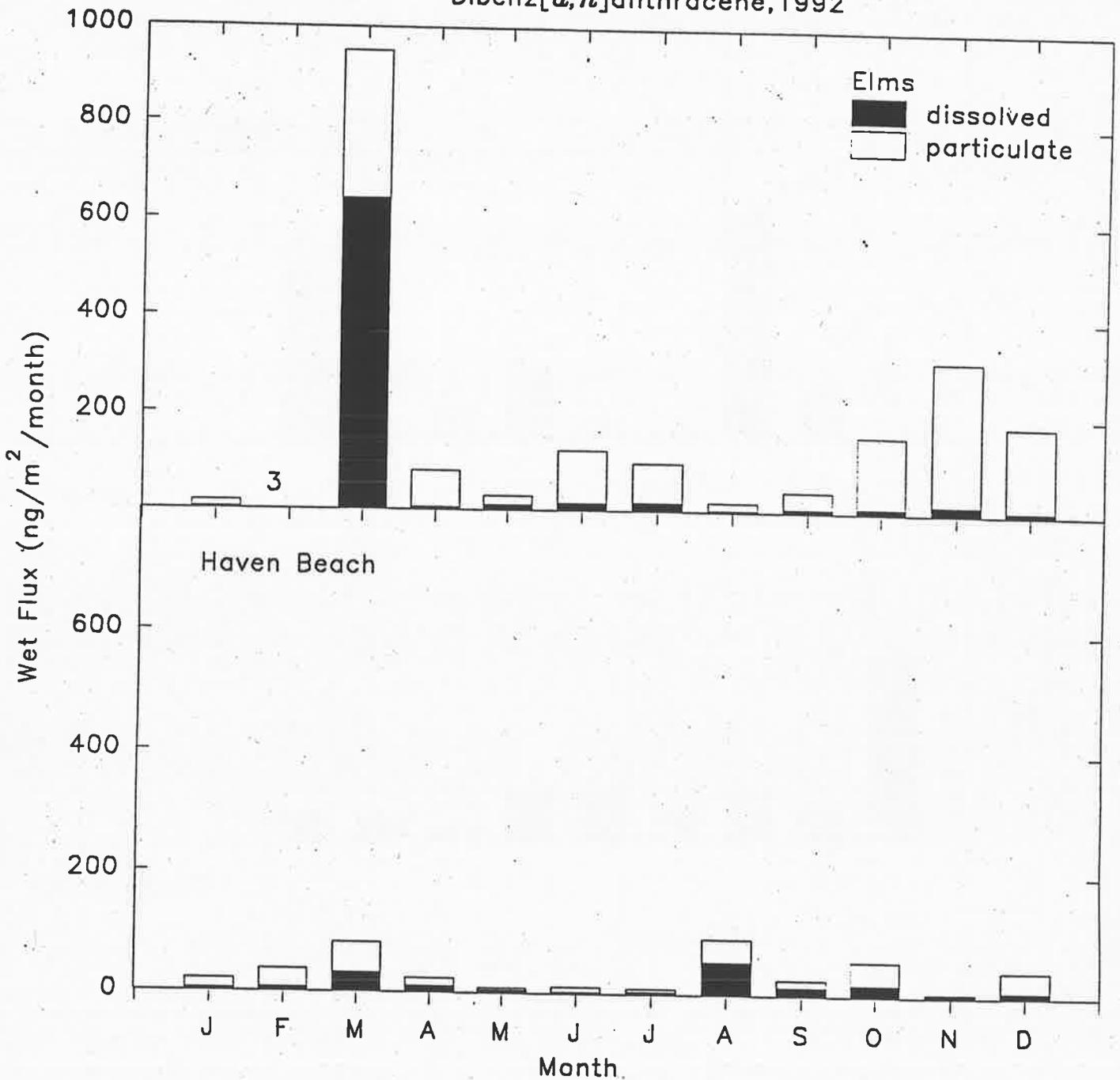


A1.59 Indeno[1,2,3-cd]pyrene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Indeno[1,2,3-cd]pyrene, 1993

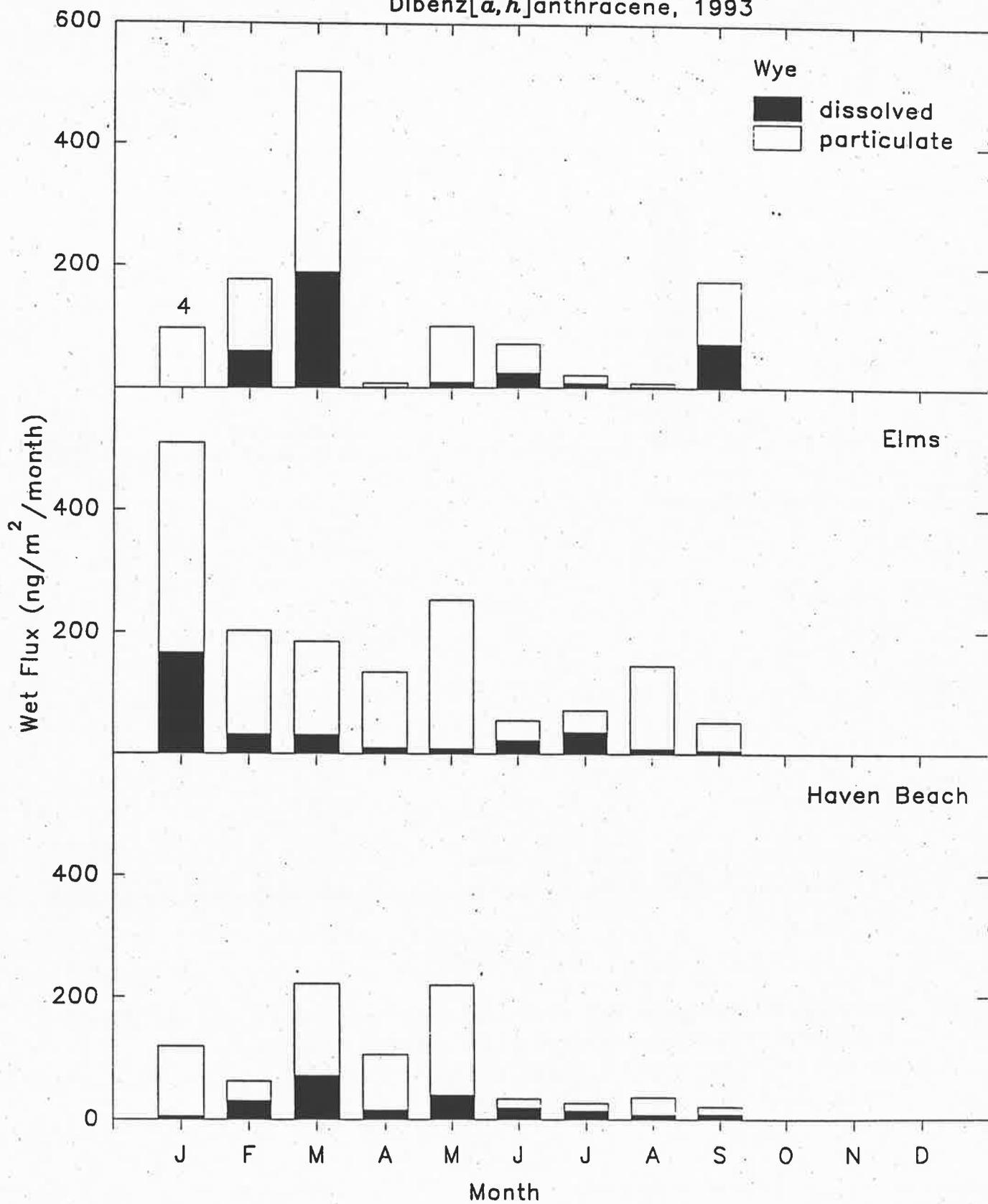


Dibenz[*a,h*]anthracene, 1992

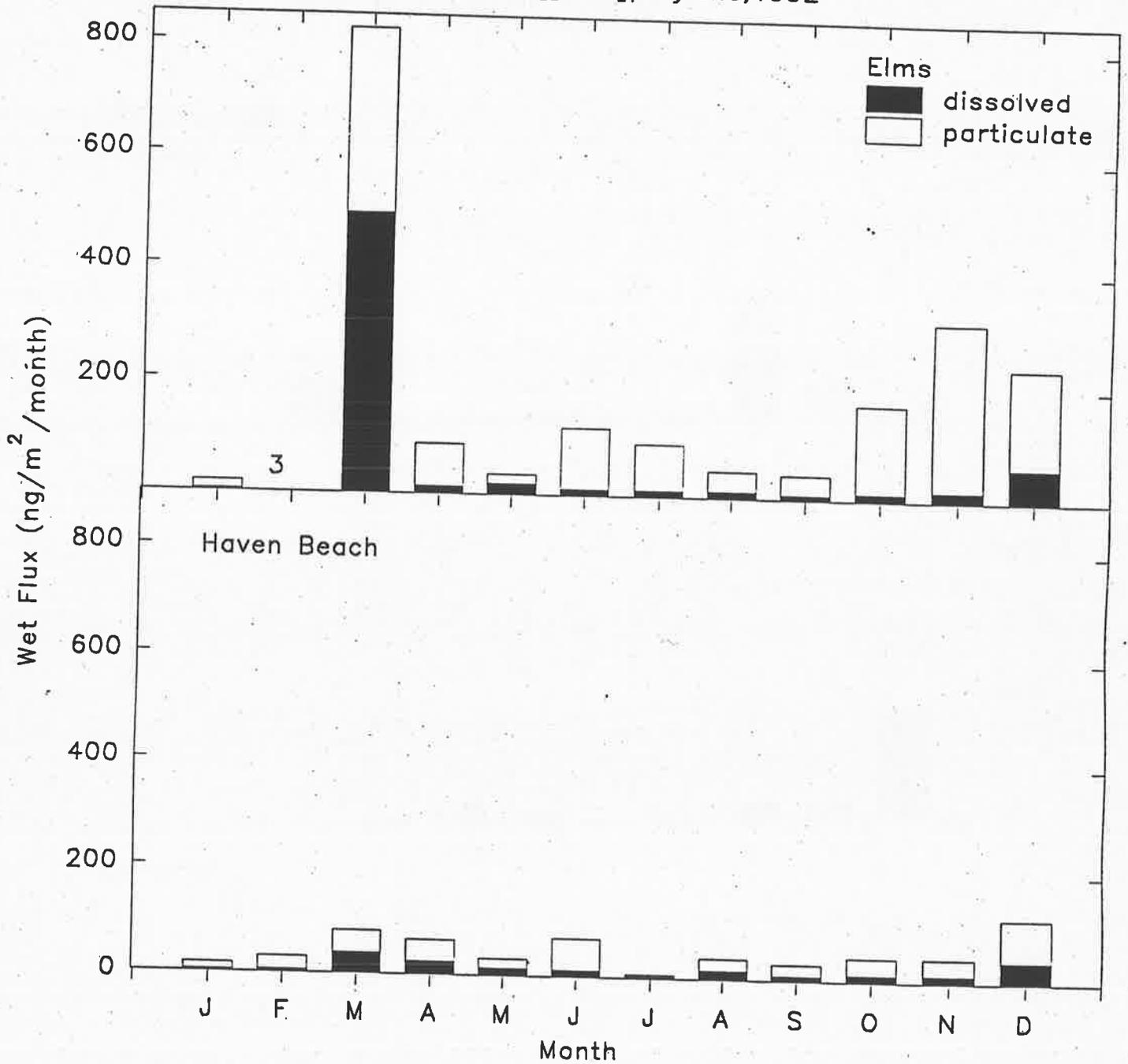


A1.60 Dibenz[*a,h*]anthracene wet depositional fluxes measured at the Wye, Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3= sample down, 4 = lost dissolved sample).

Dibenz[*a,h*]anthracene, 1993

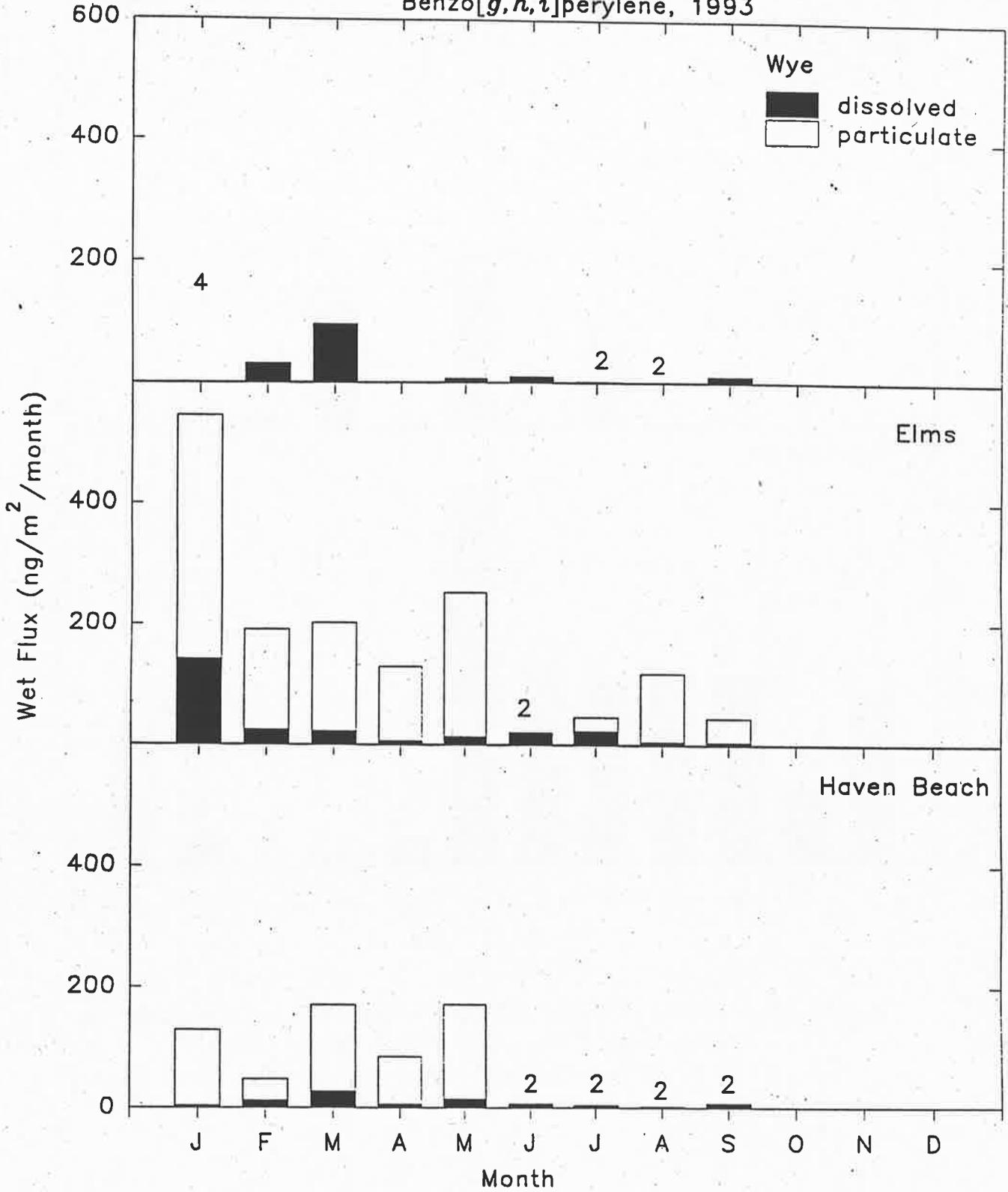


Benzo[*g,h,i*]perylene, 1992

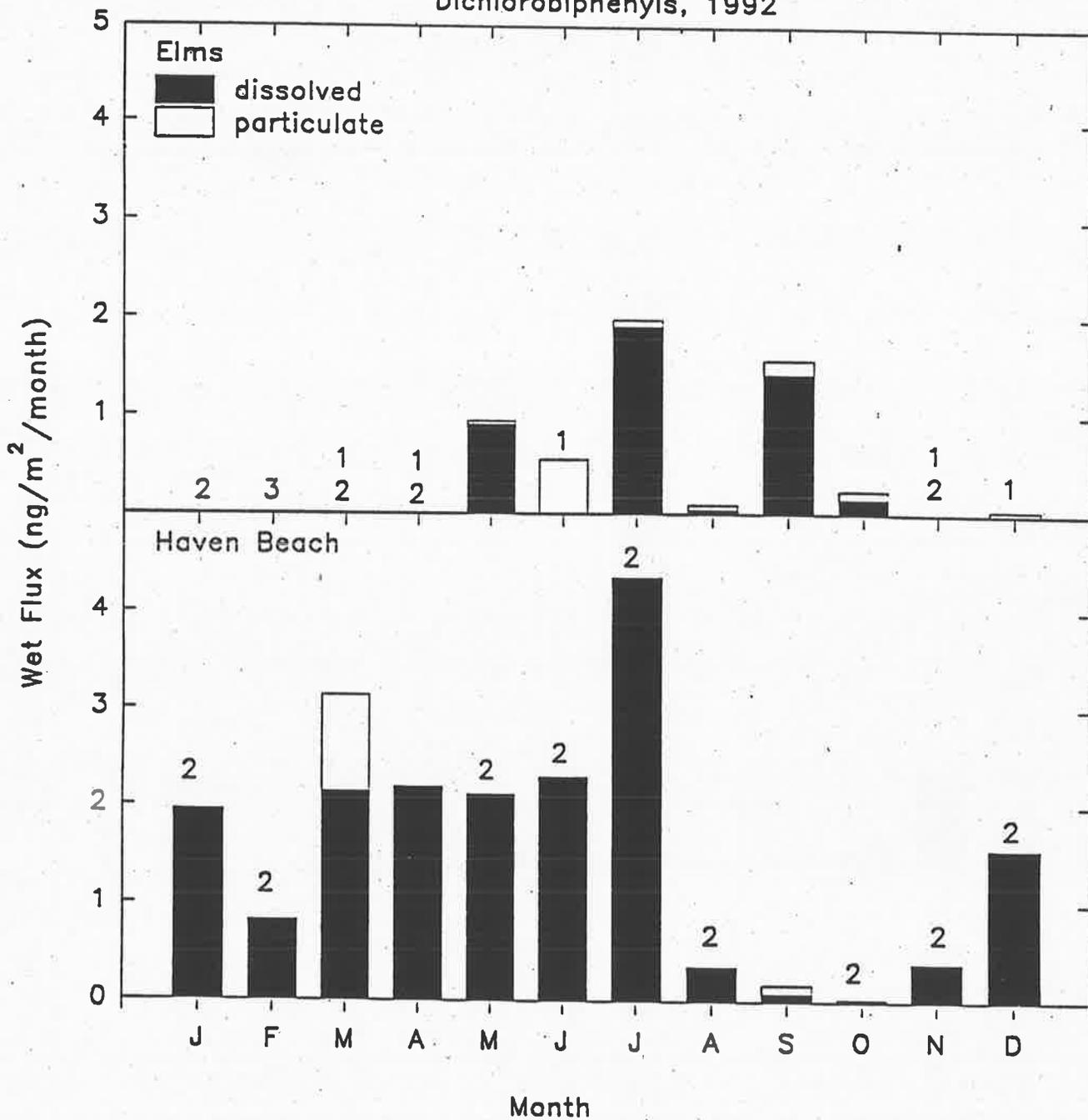


A1.61 Benzo[*g,h,i*]perylene wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1=not quantifiable in dissolved phase, 2=not quantifiable on filter, 3=sampler down, 4 = lost dissolved sample).

Benzo[*g,h,i*]perylene, 1993

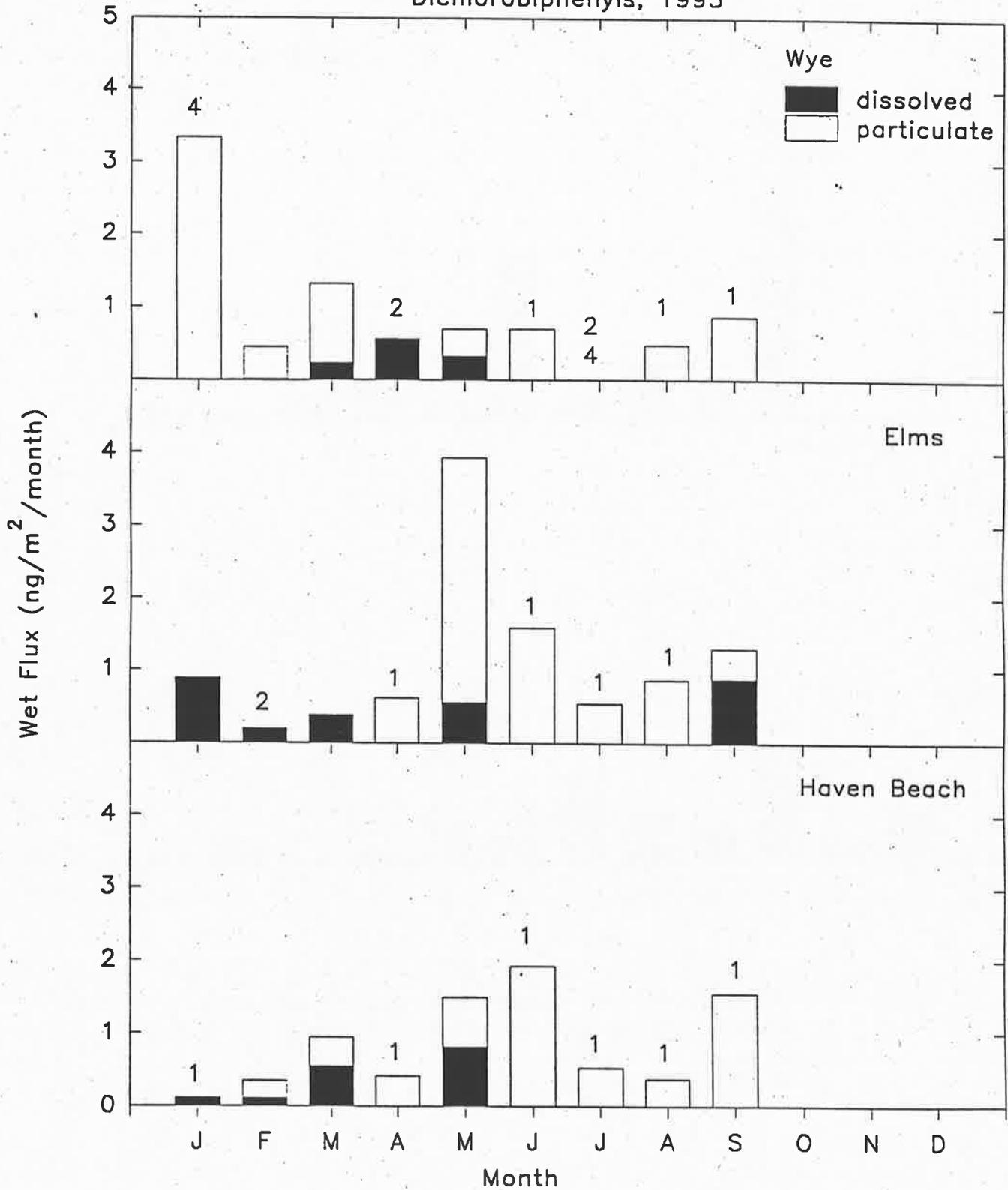


Dichlorobiphenyls, 1992

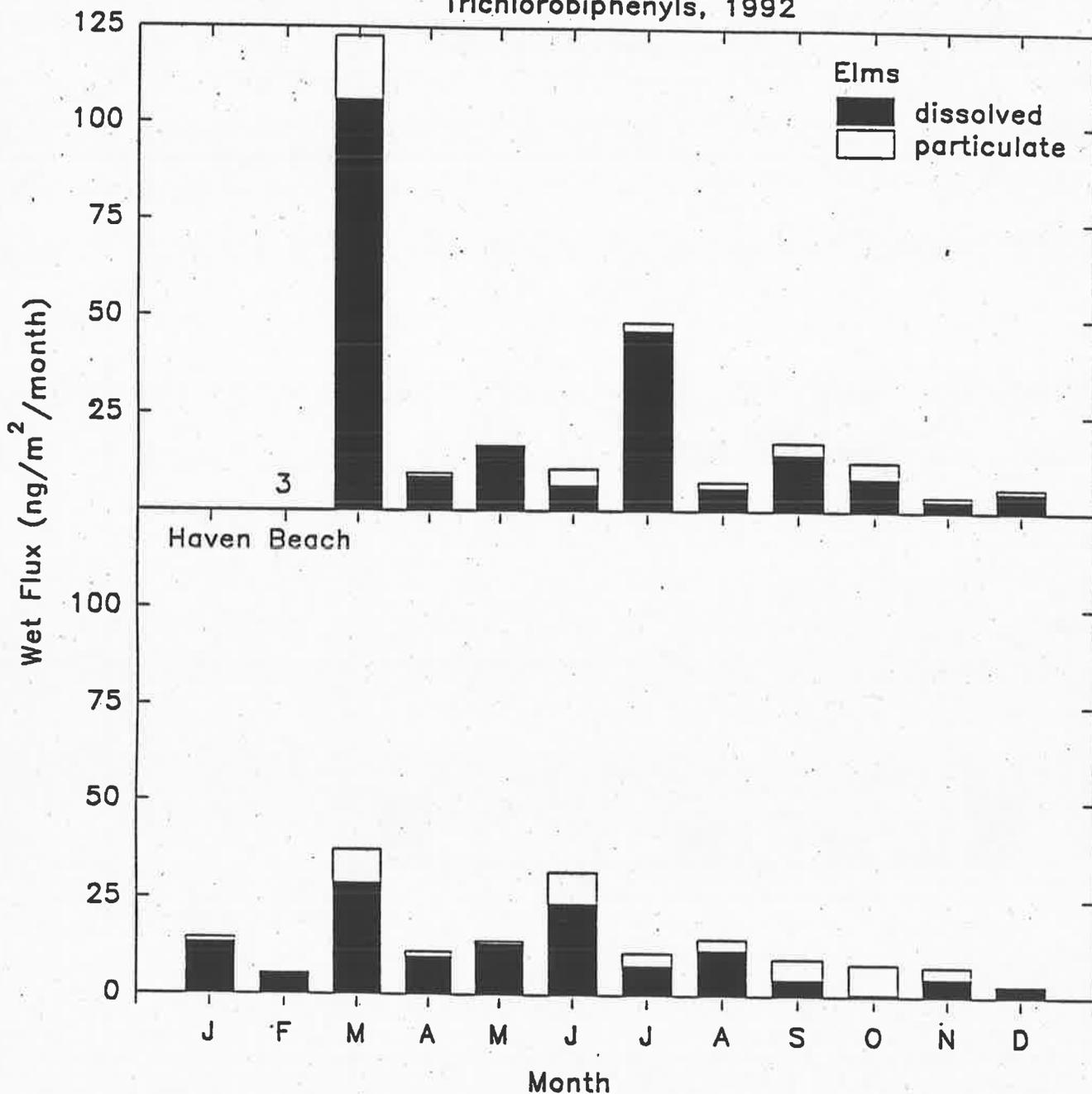


A1.62 Dichlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Dichlorobiphenyls, 1993

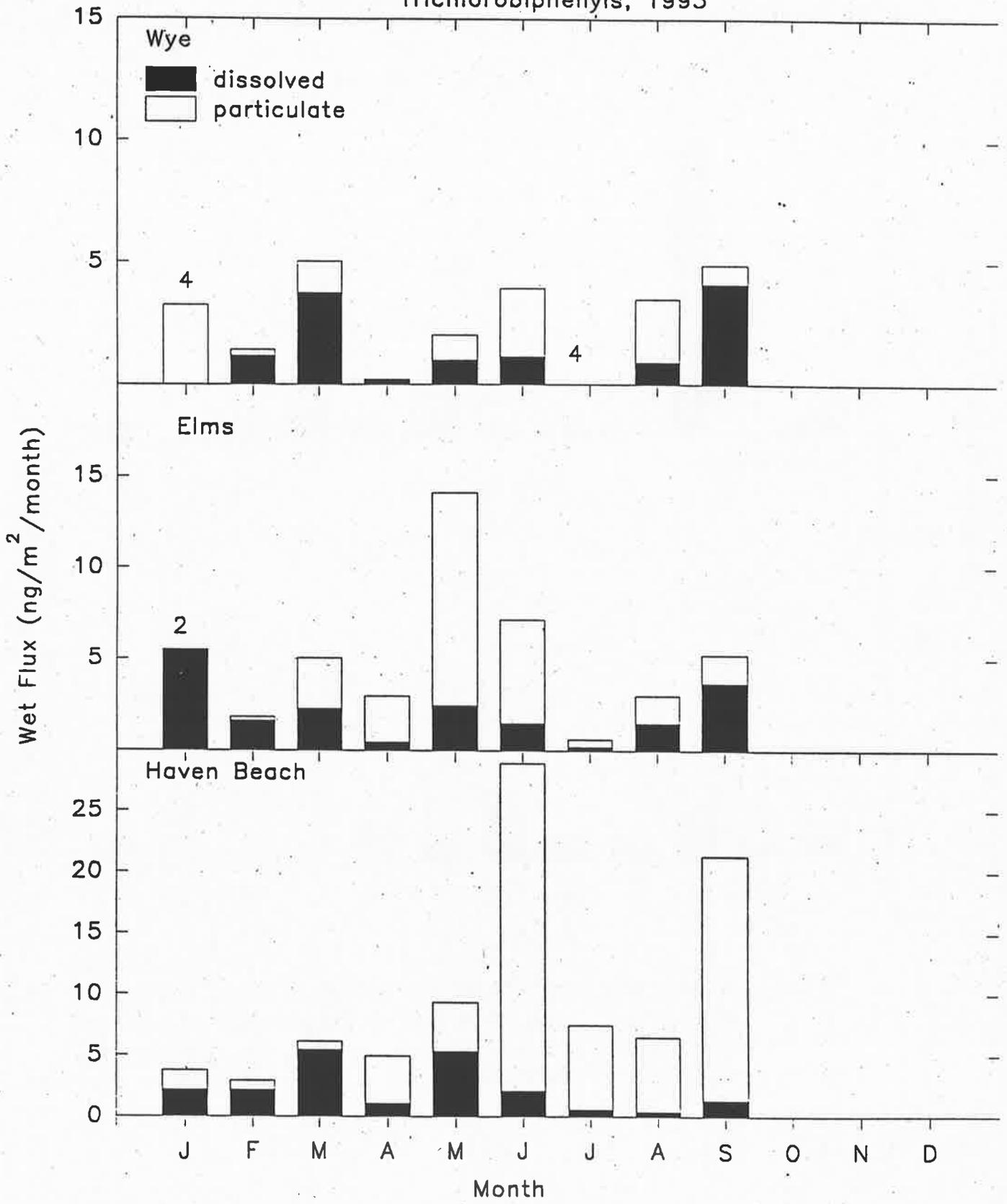


Trichlorobiphenyls, 1992

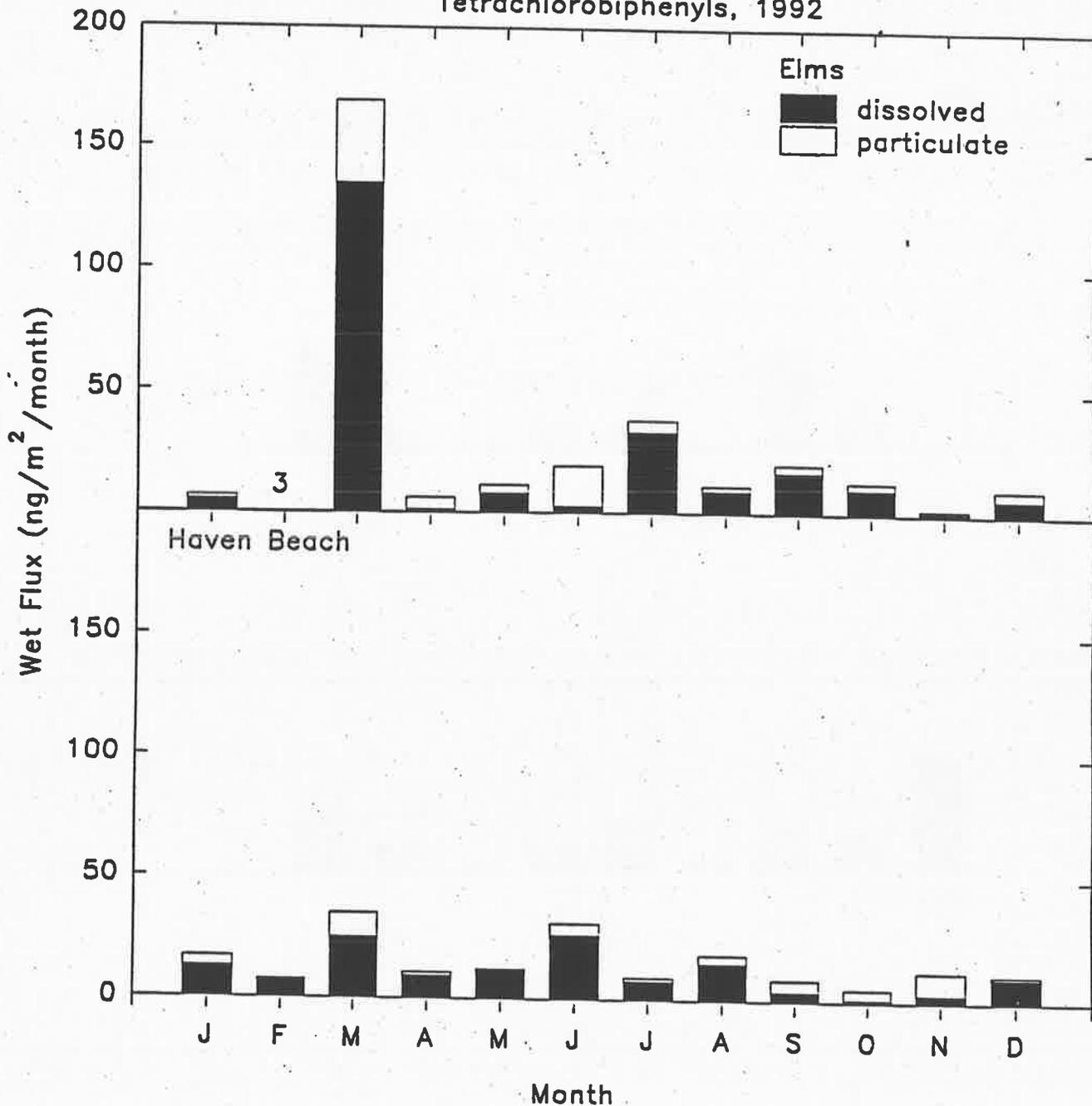


A1.63 Trichlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Trichlorobiphenyls, 1993

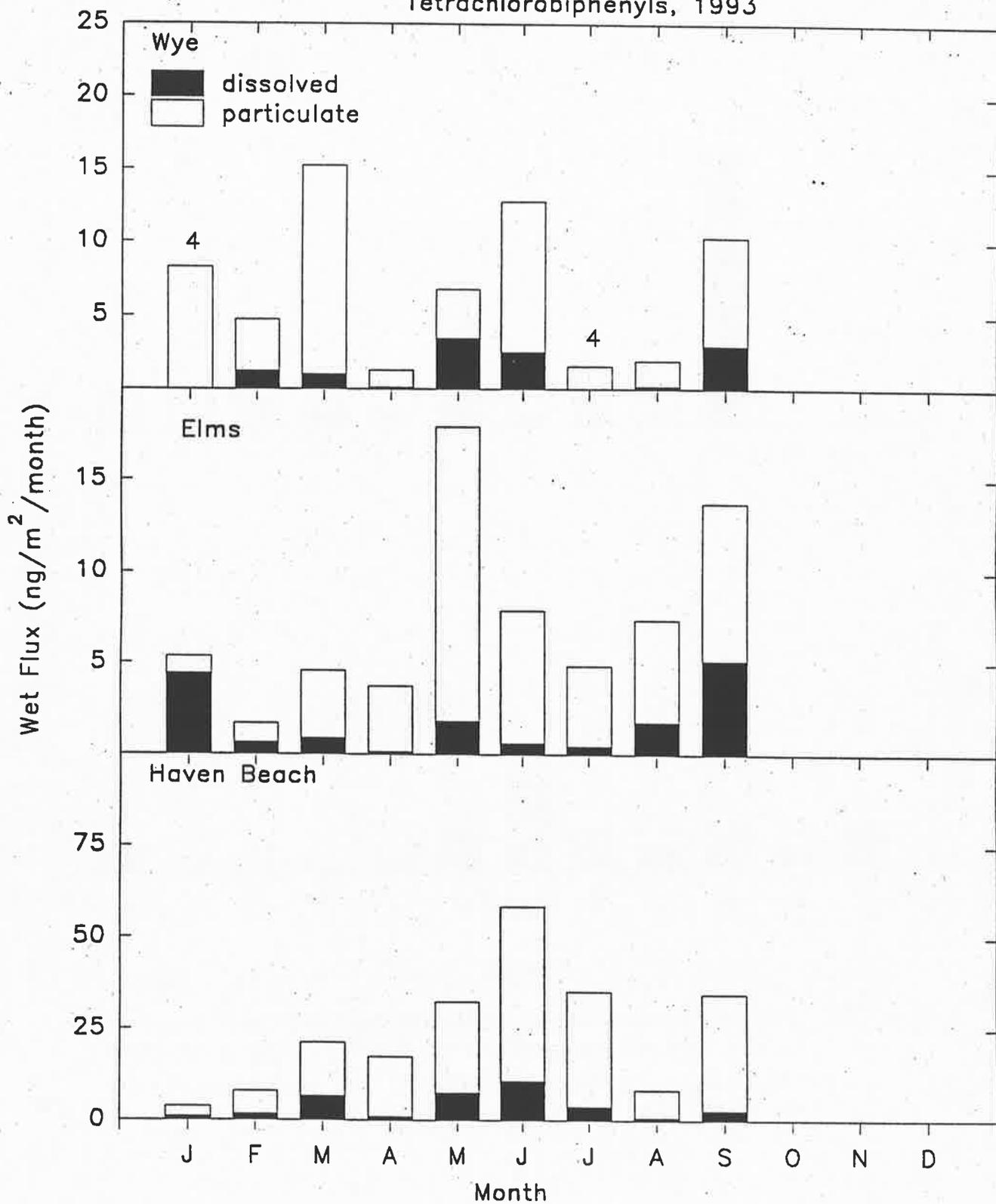


Tetrachlorobiphenyls, 1992

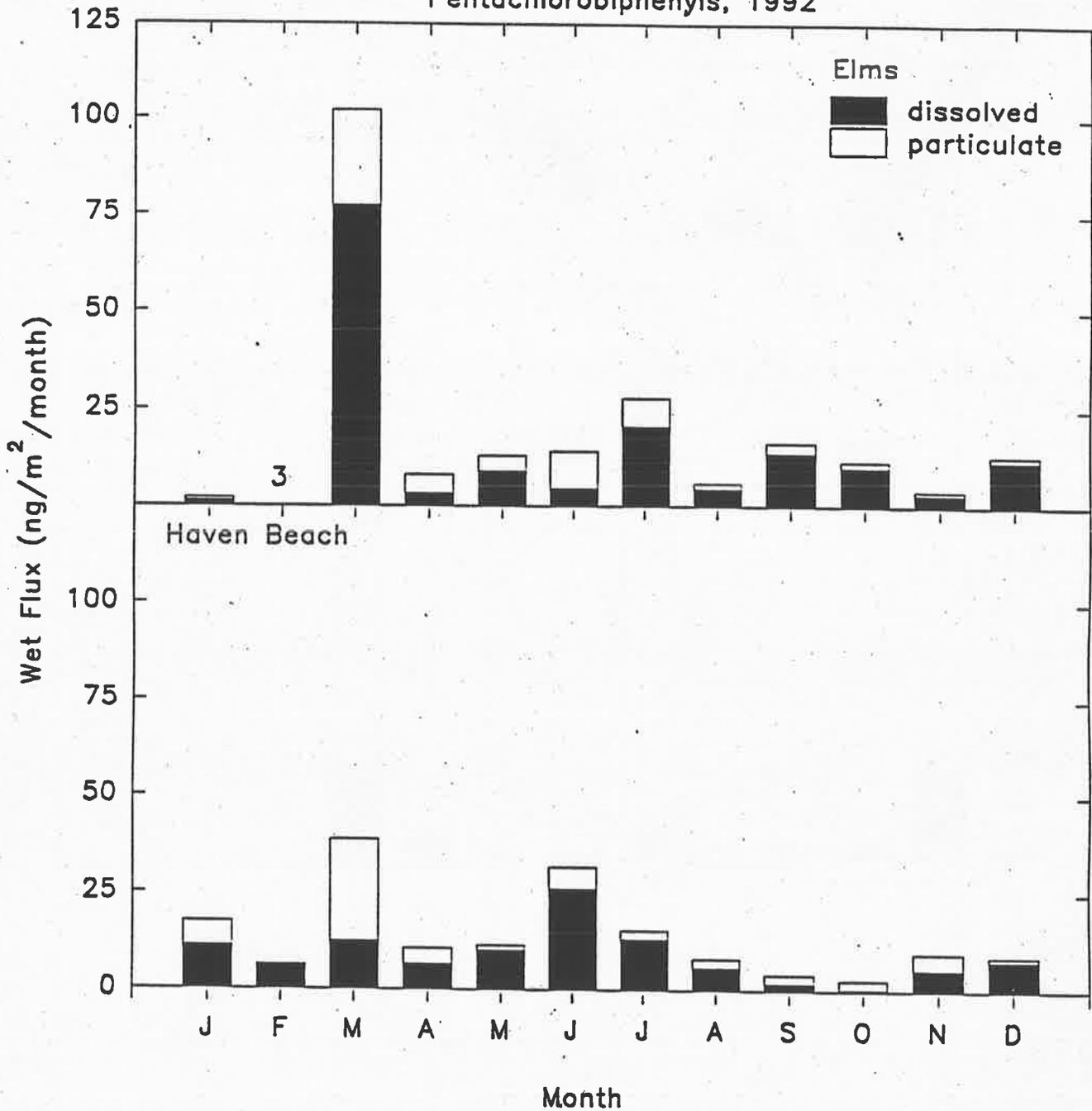


A1.64 Tetrachlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2= not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Tetrachlorobiphenyls, 1993

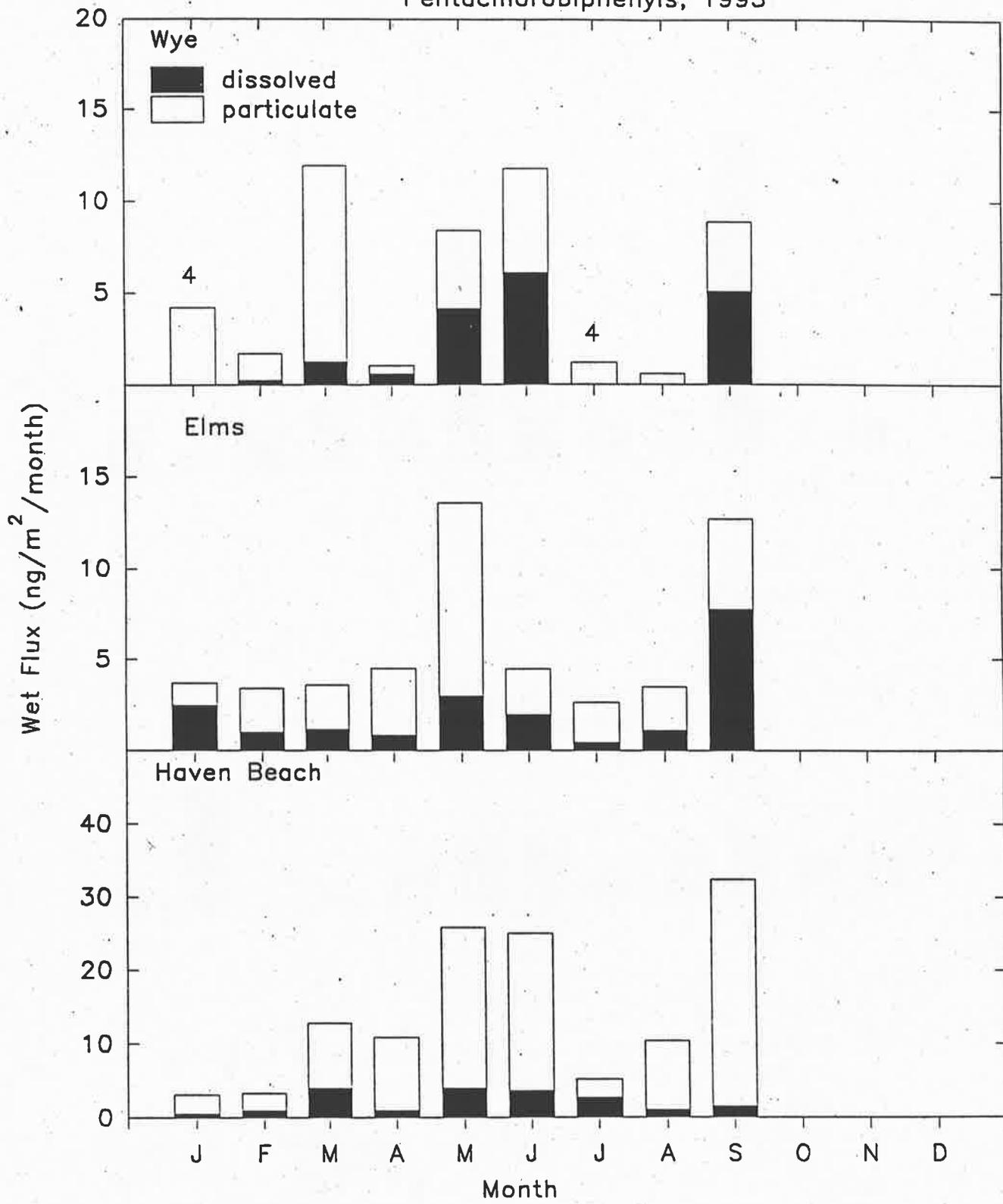


Pentachlorobiphenyls, 1992

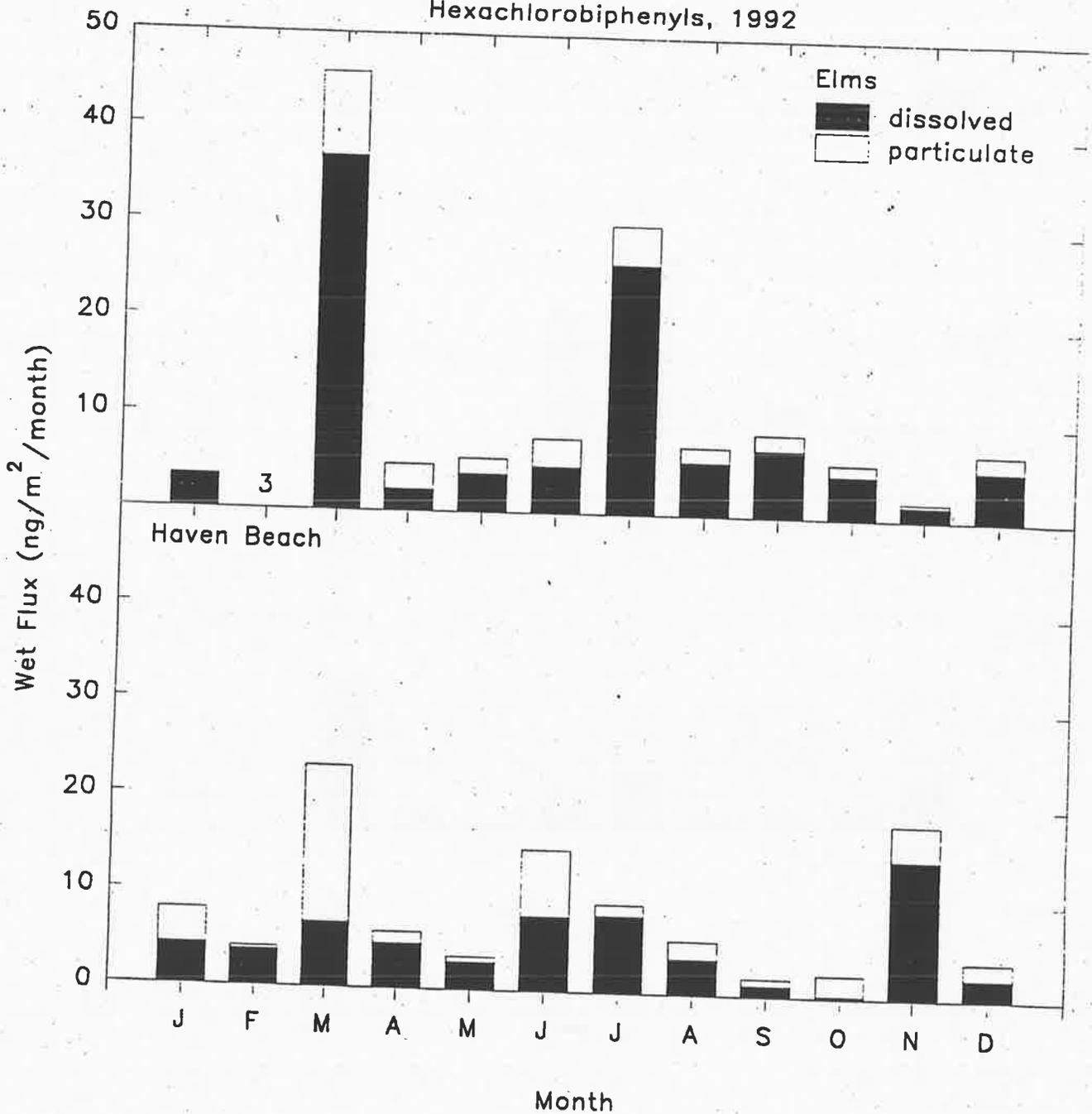


A1.65 Pentachlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample)..

Pentachlorobiphenyls, 1993

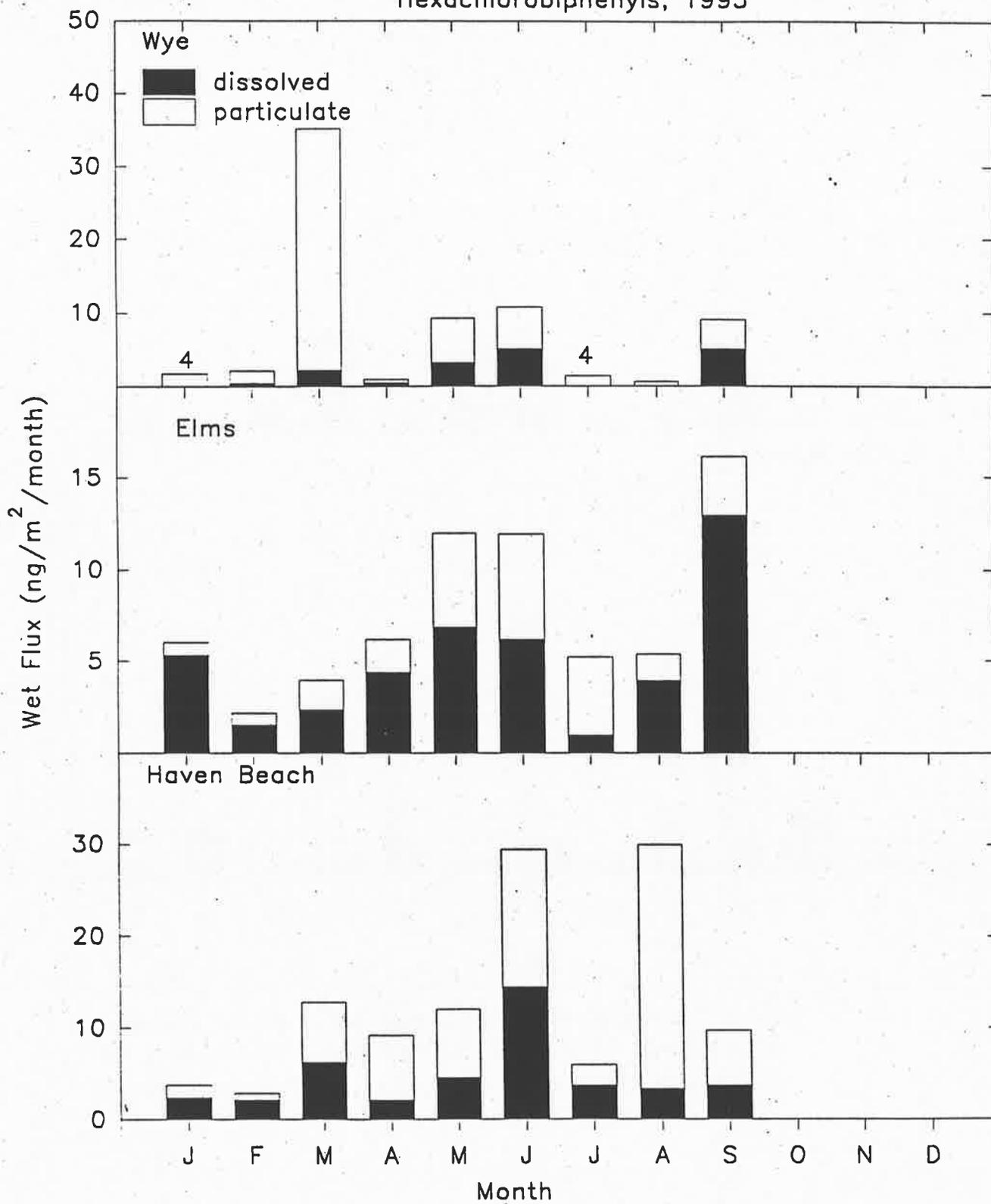


Hexachlorobiphenyls, 1992

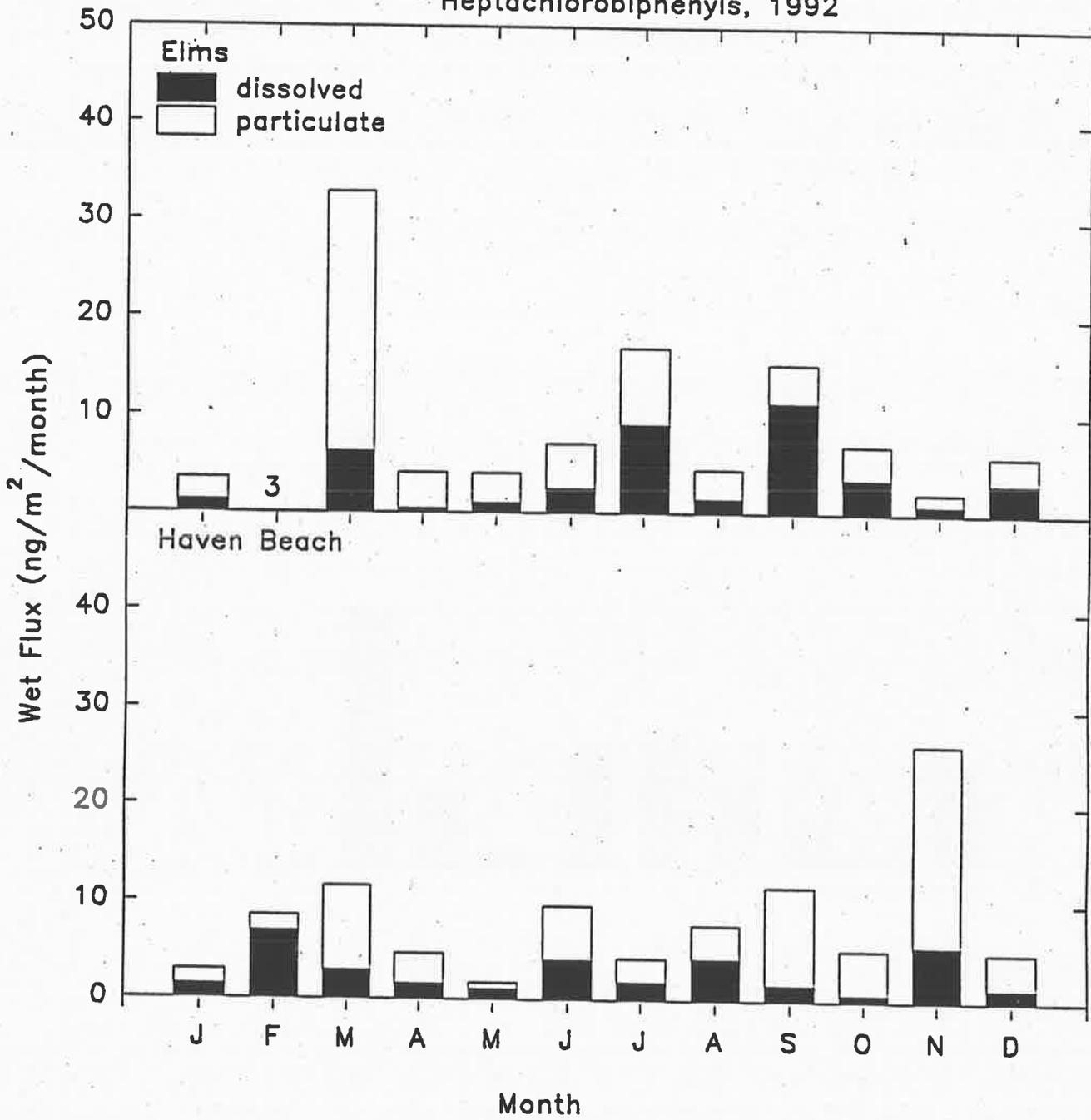


A1.66 Hexachlorobiphenyl wet depositional fluxes measured at the Elms and Haven Beach sites. (1=not quantifiable in dissolved phase, 2= not quantifiable on filter, 3= sampler down, 4 = lost dissolved sample).

Hexachlorobiphenyls, 1993

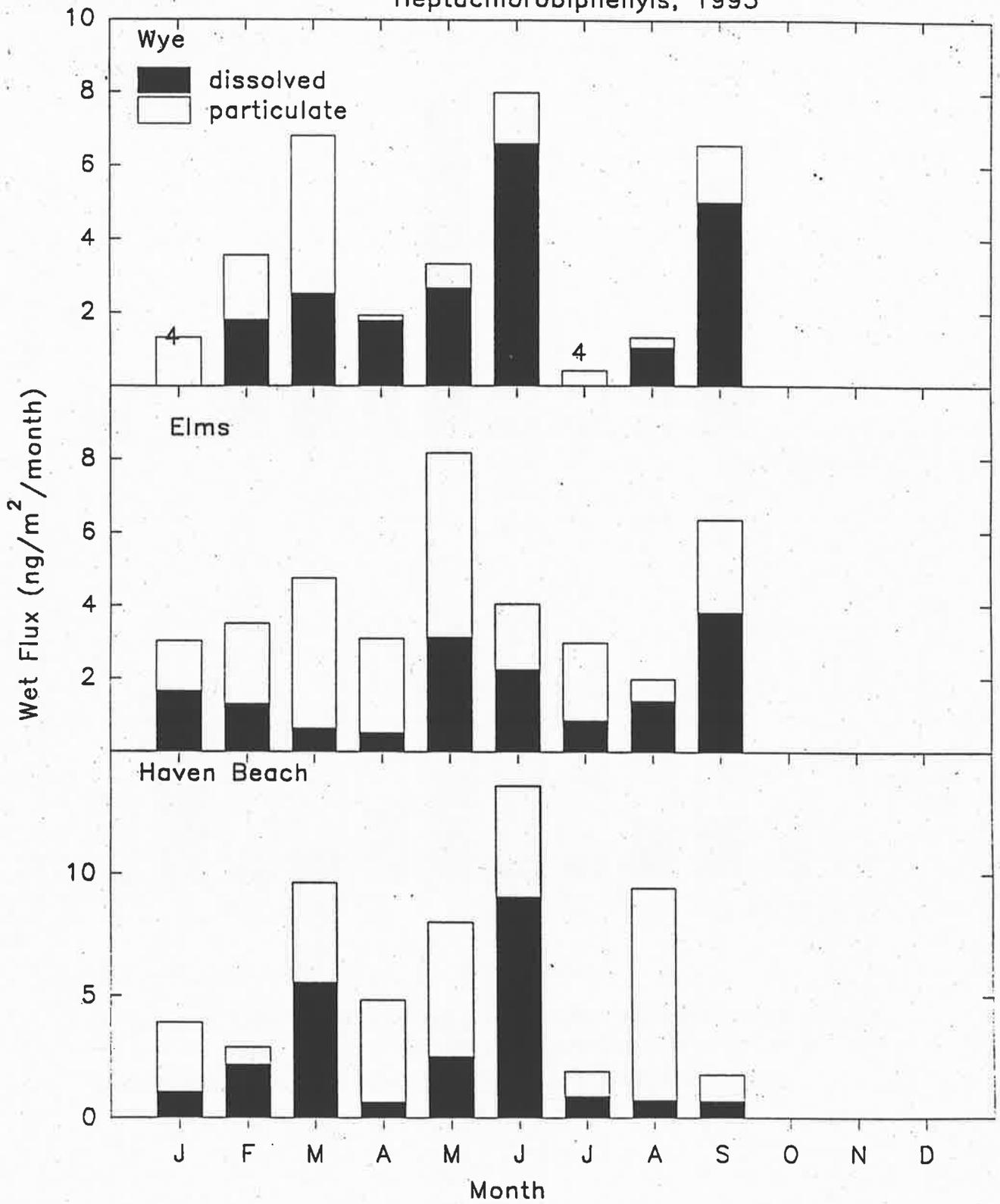


Heptachlorobiphenyls, 1992

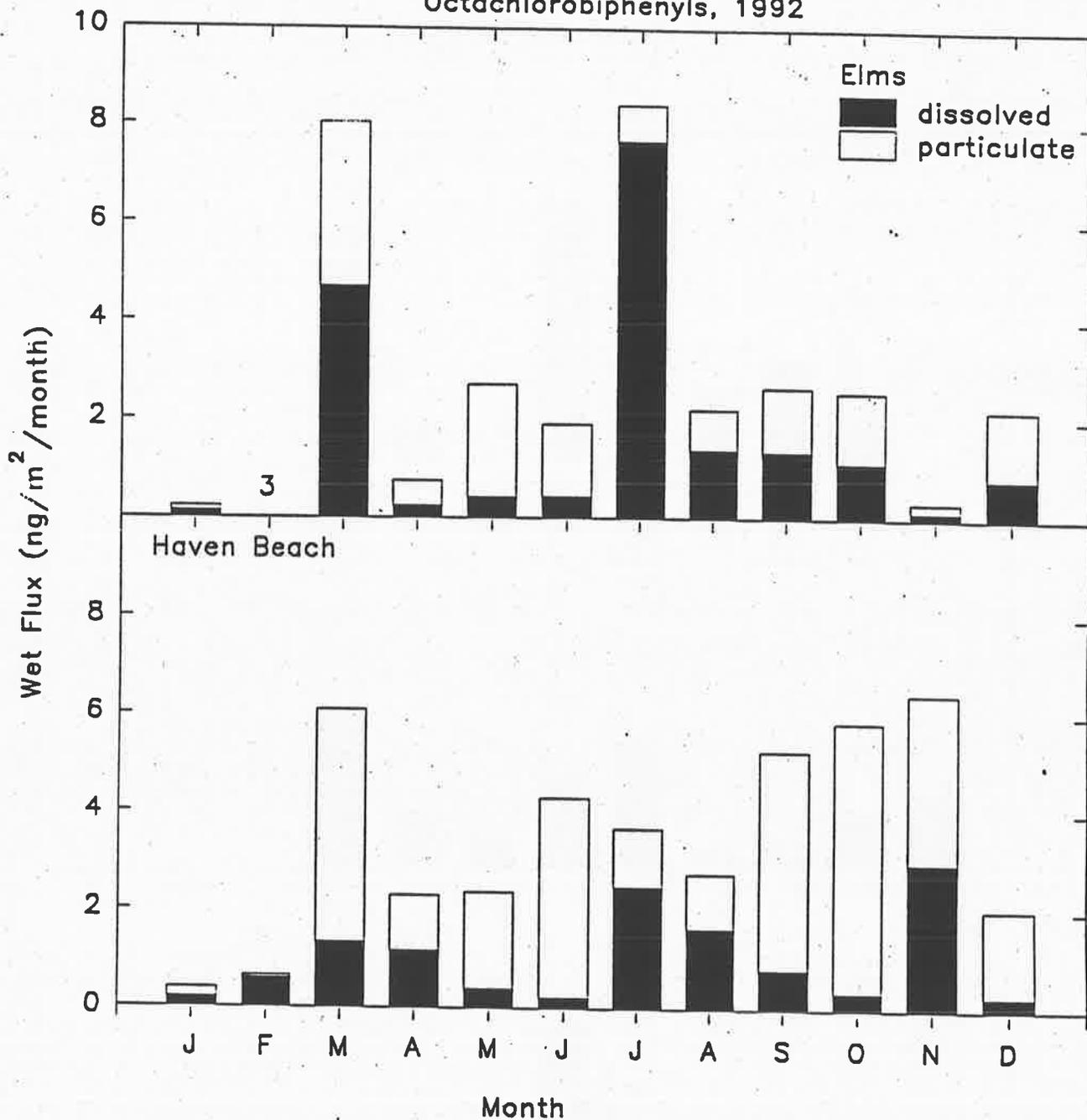


A1.67 Heptachlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Heptachlorobiphenyls, 1993

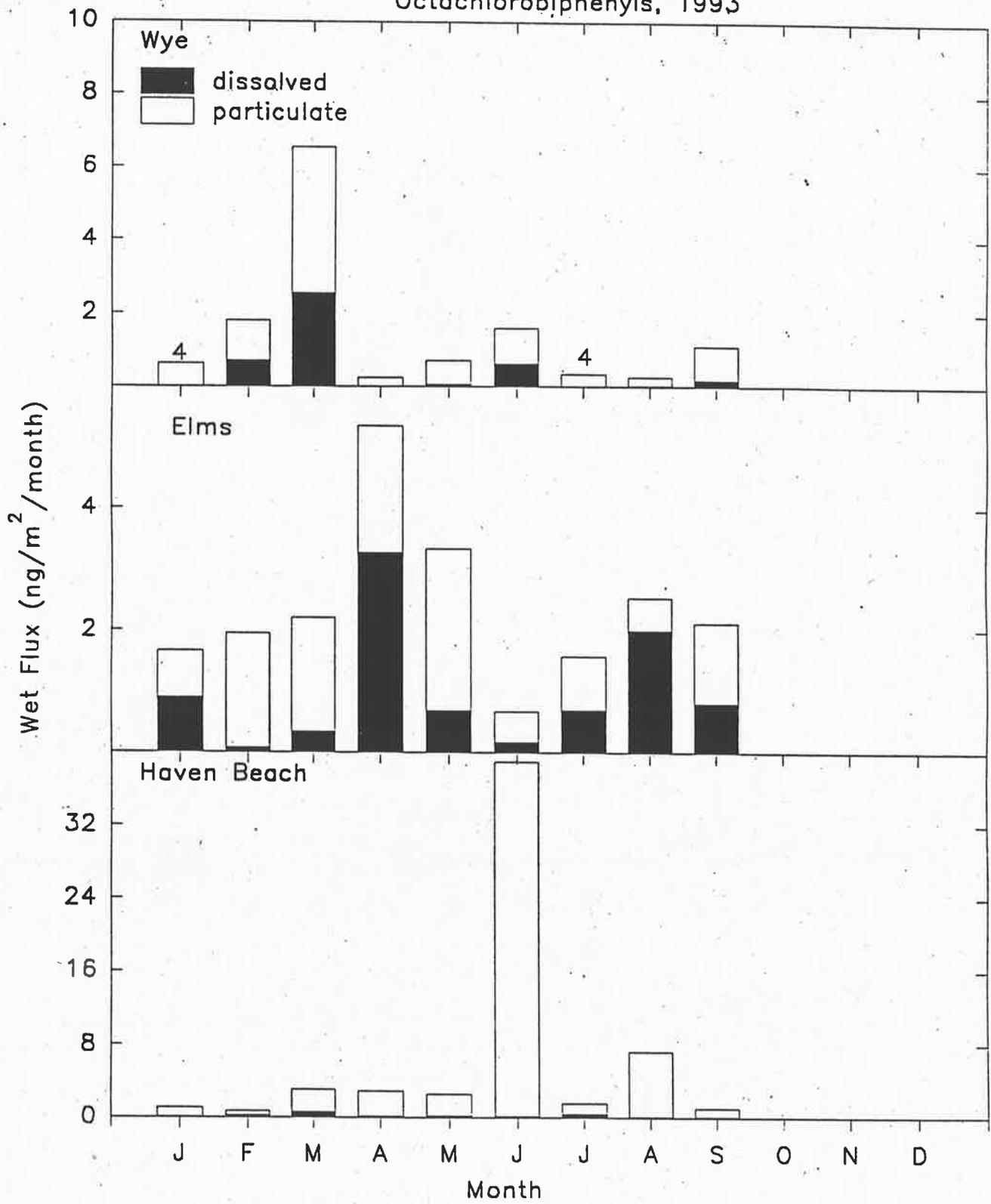


Octachlorobiphenyls, 1992

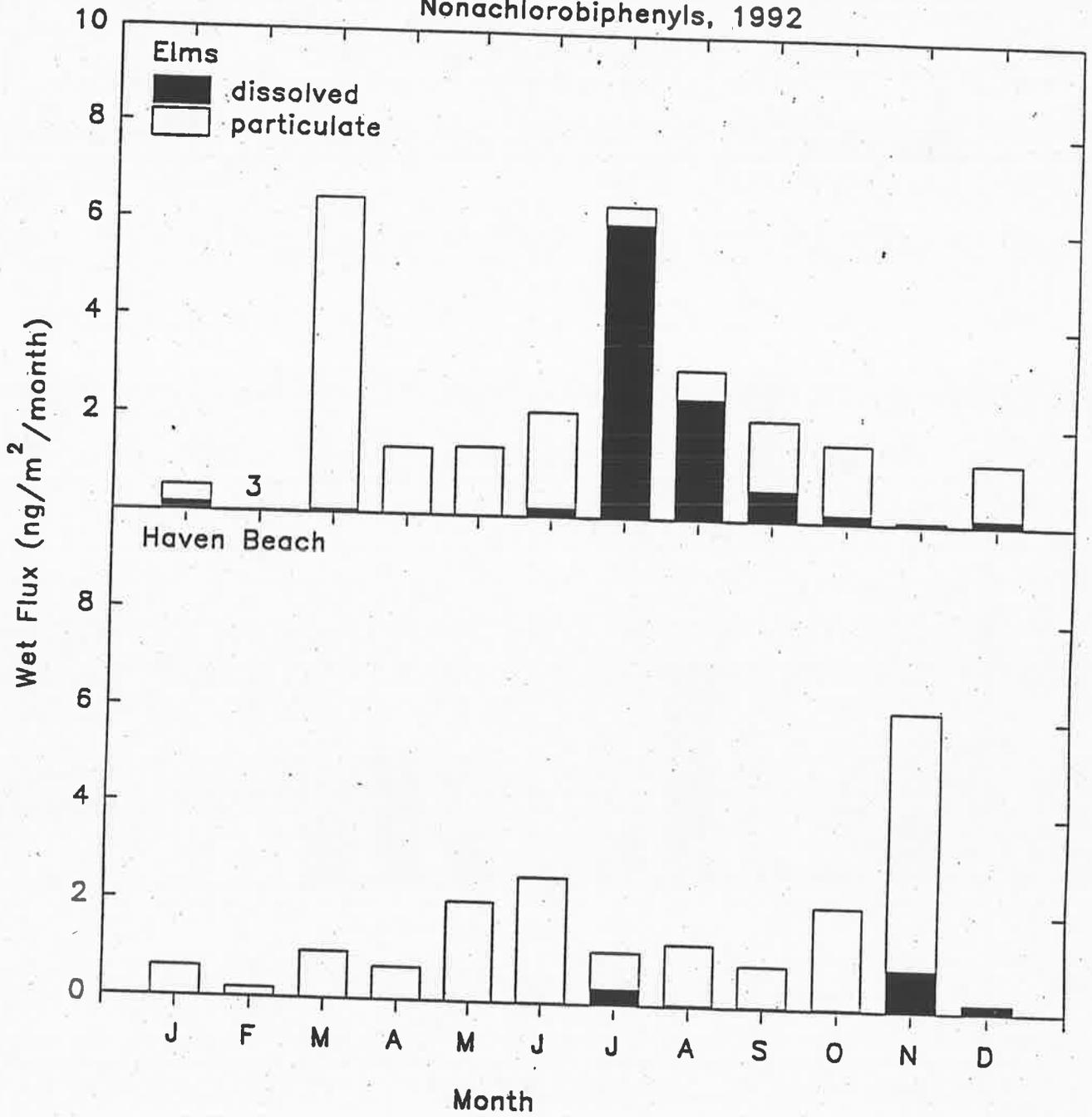


A1.68 Octachlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Octachlorobiphenyls, 1993

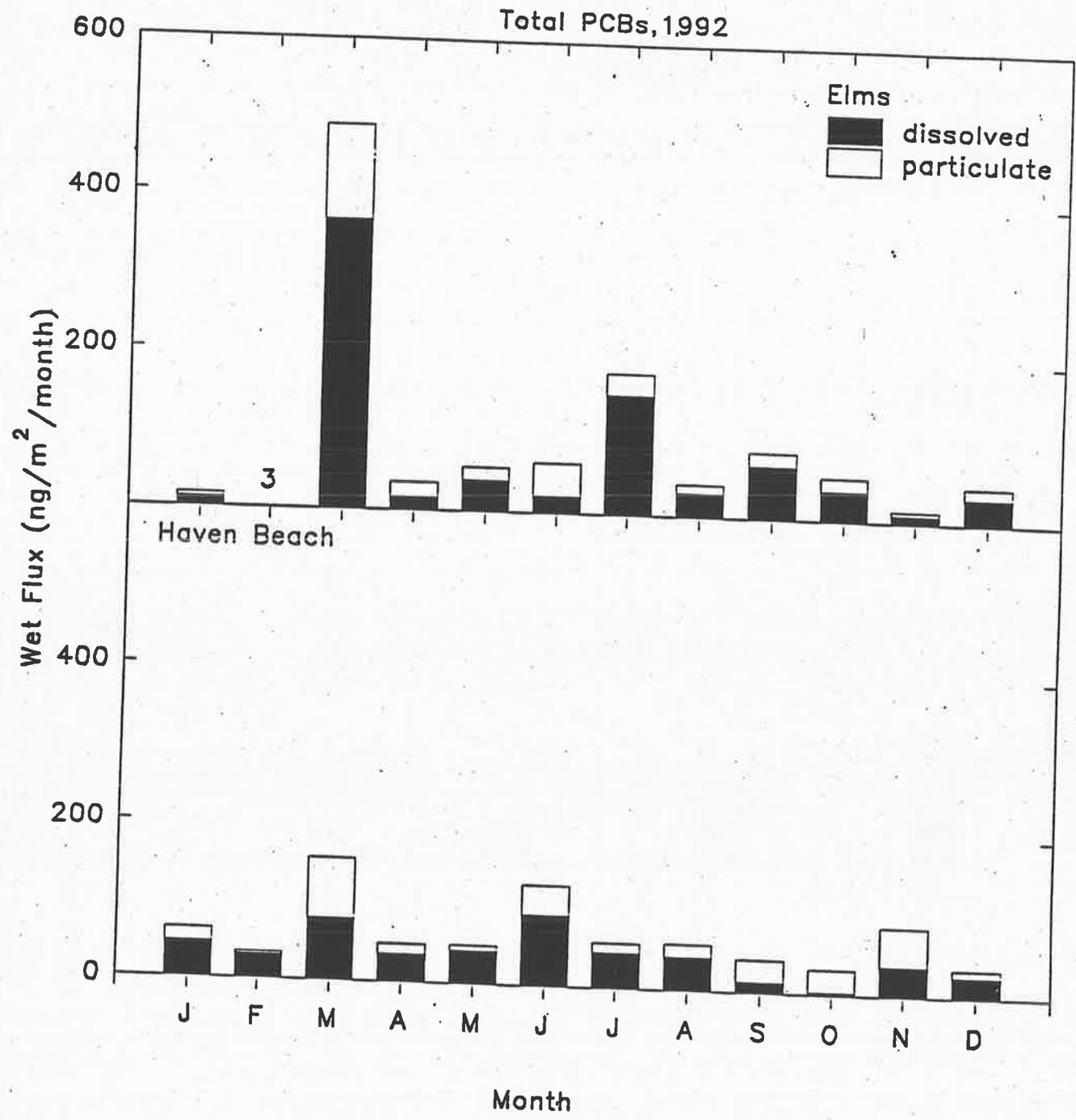


Nonachlorobiphenyls, 1992



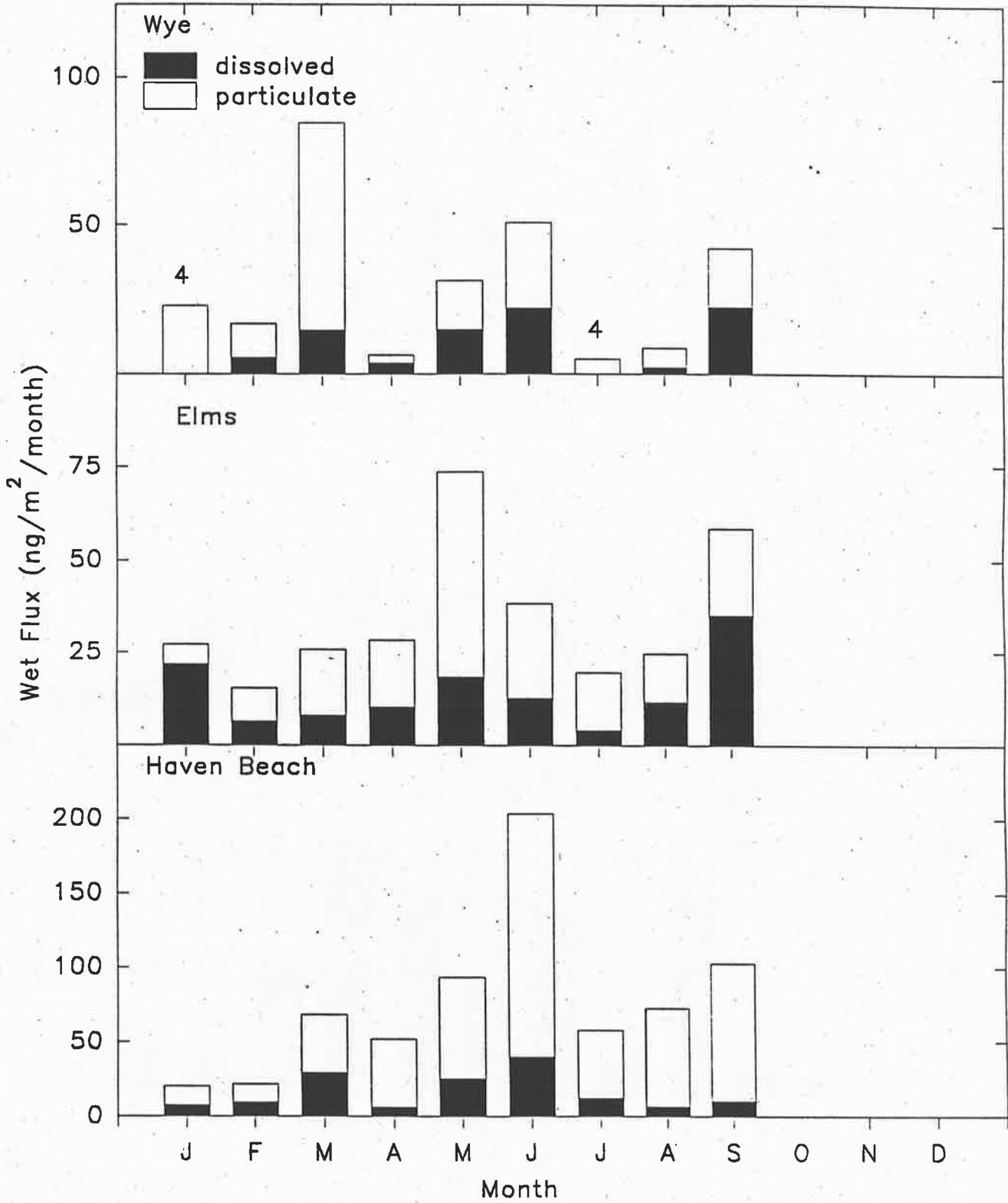
A1.69 Nonachlorobiphenyl wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Total PCBs, 1992

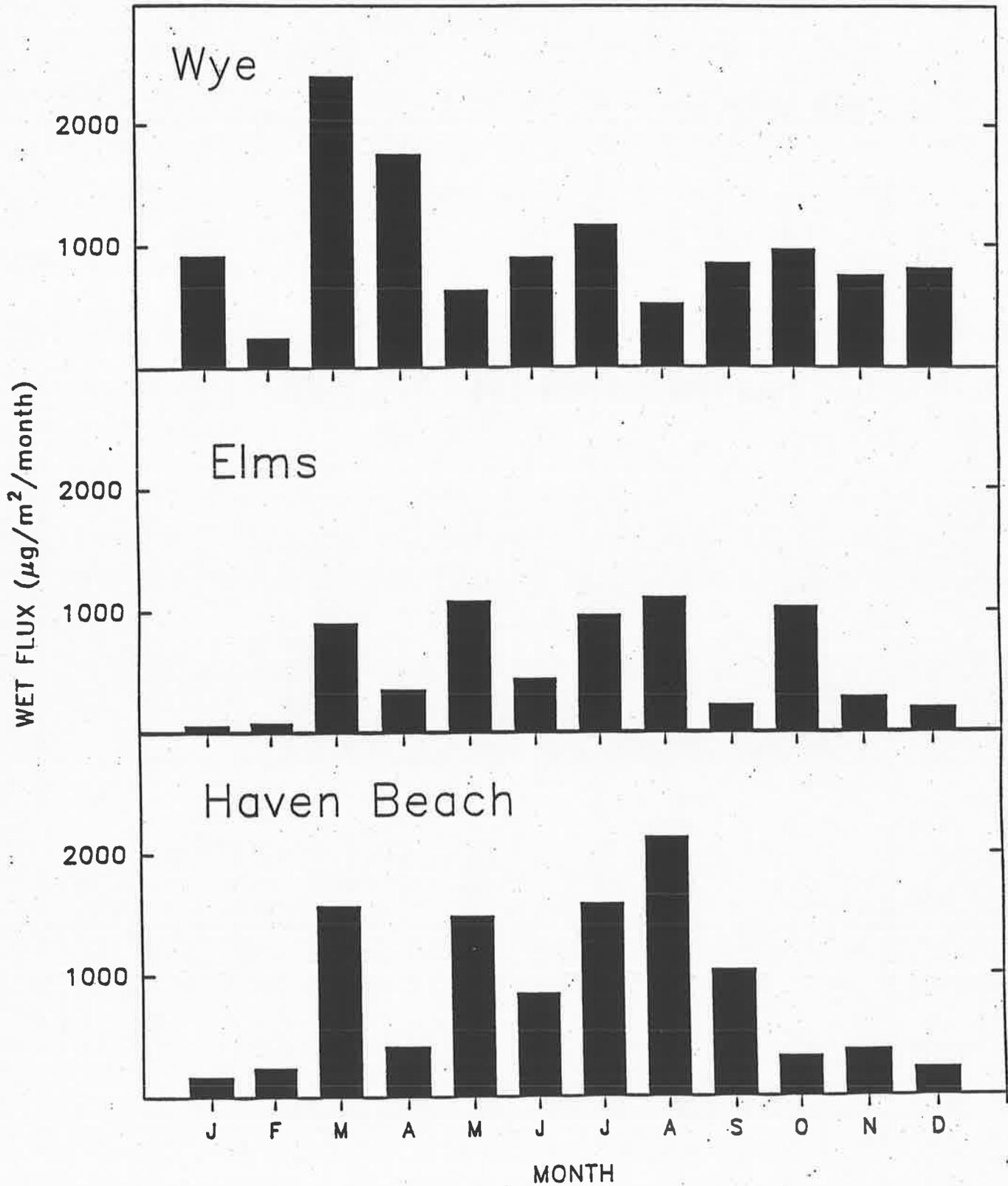


A1.70 Total PCB wet depositional fluxes measured at the Wye, Elms, and Haven Beach sites. (1 = not quantifiable in dissolved phase, 2 = not quantifiable on filter, 3 = sampler down, 4 = lost dissolved sample).

Total PCBs, 1993

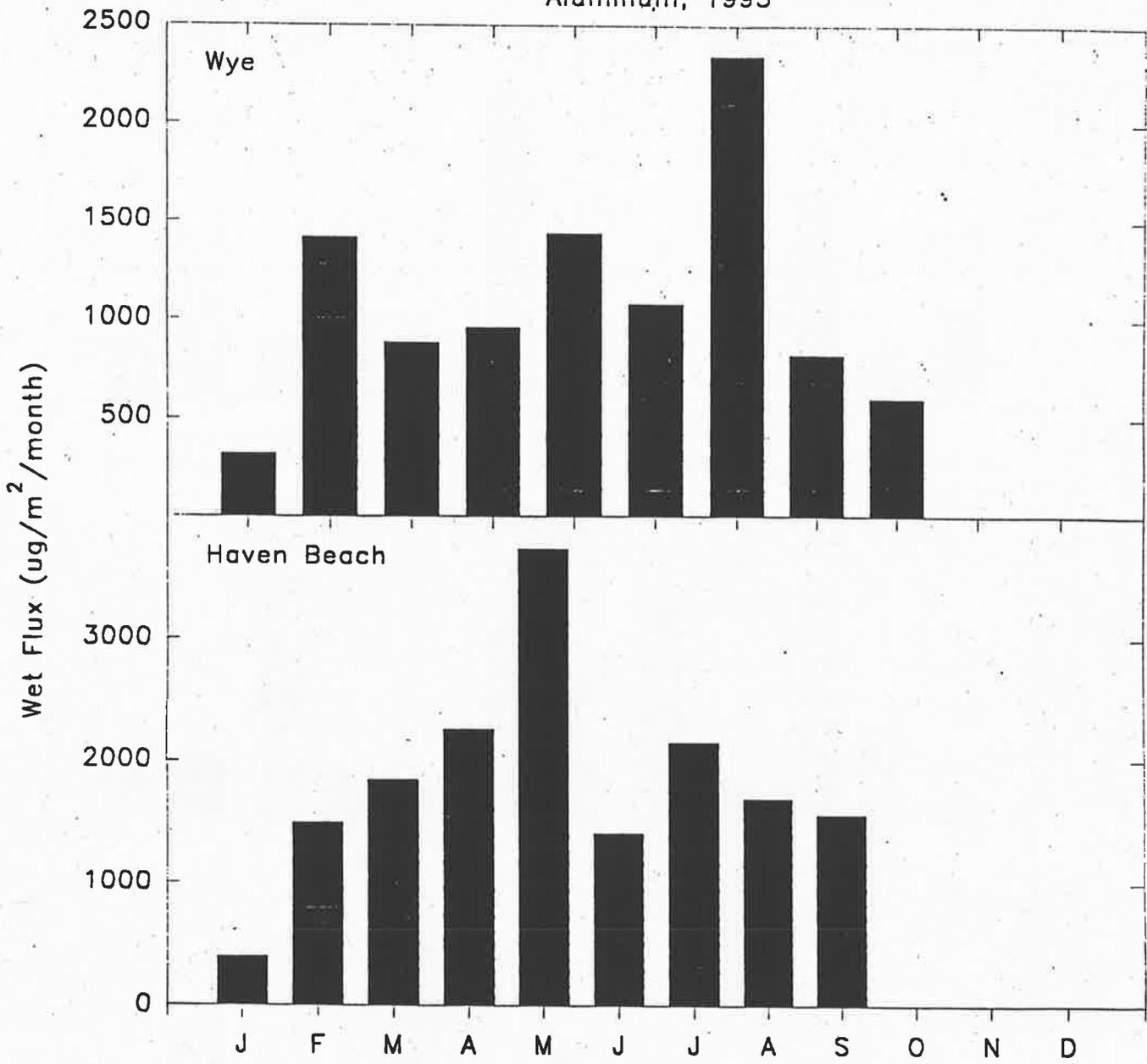


Aluminum, 1992

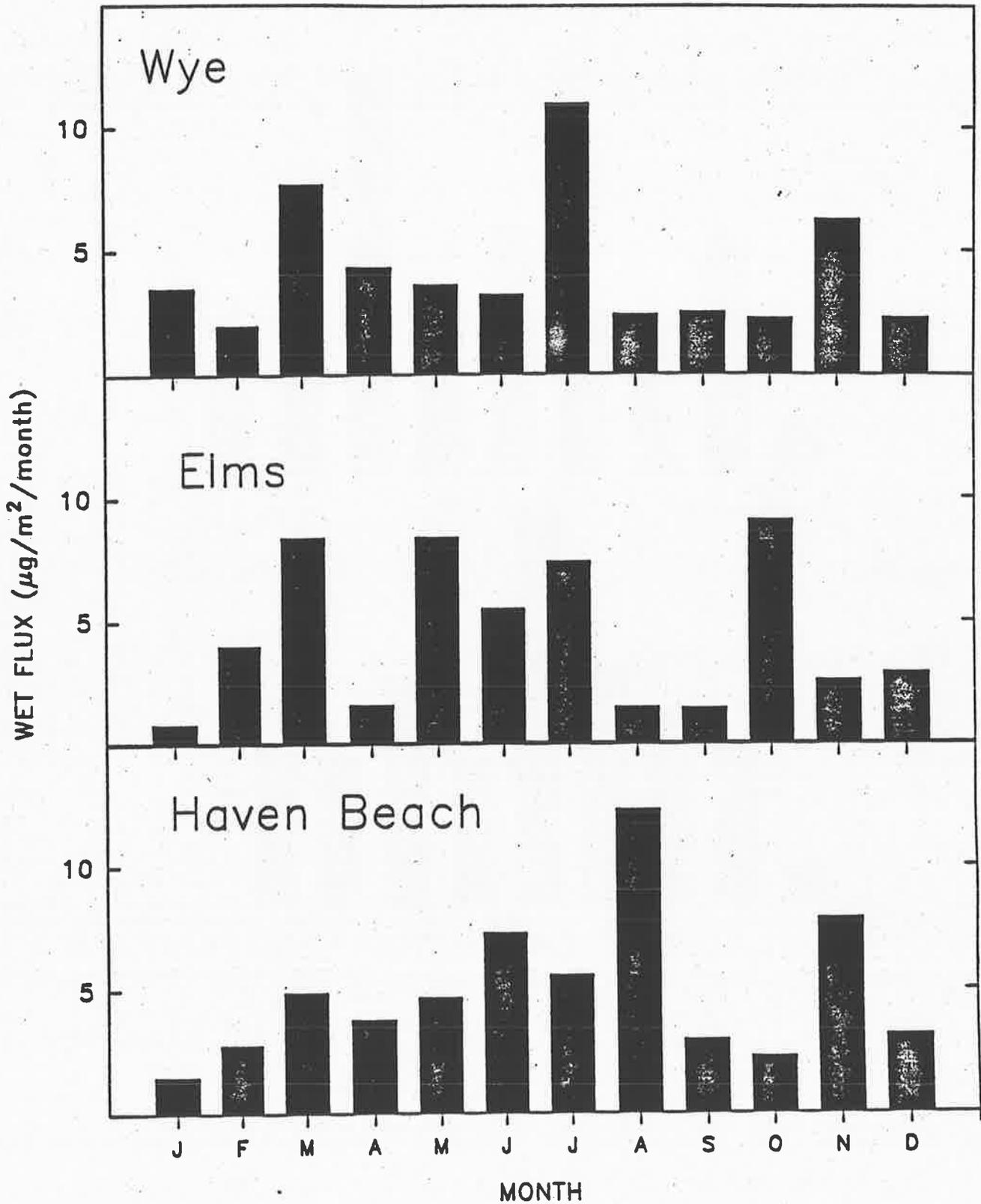


A 1.71 Monthly integrated wet deposition of Aluminum at each CBAD site.

Aluminum, 1993

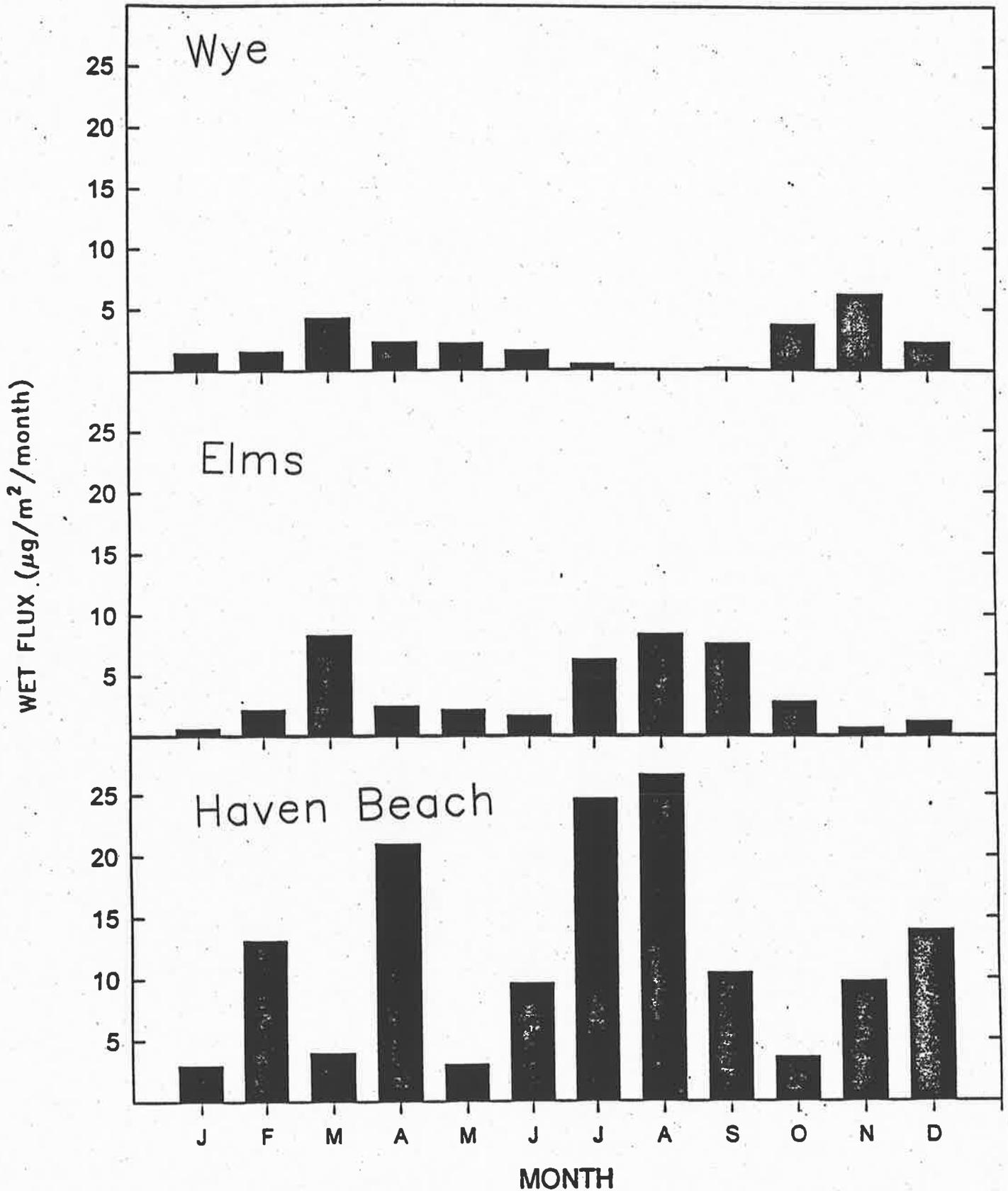


Arsenic, 1992

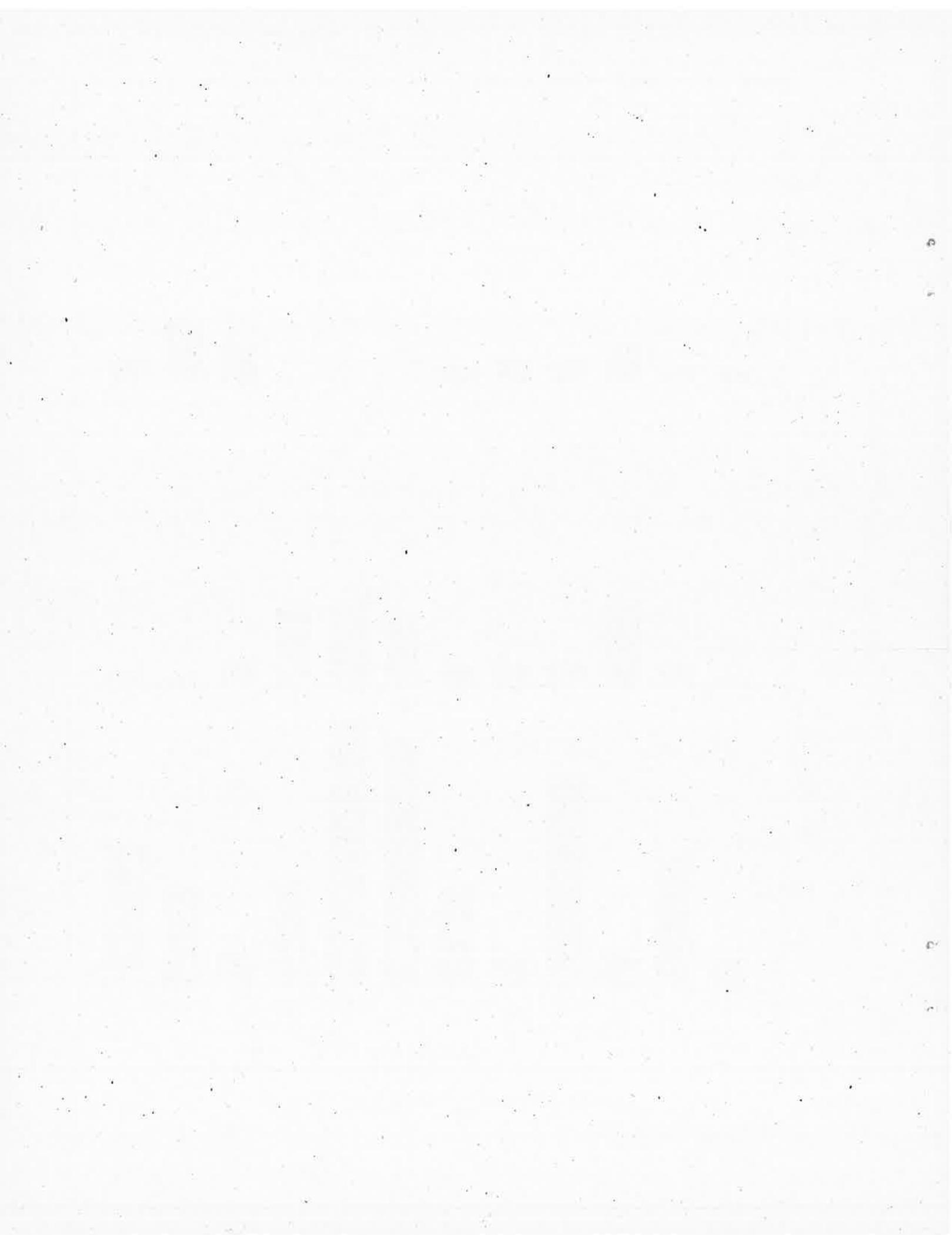


A1.72 Monthly integrated wet deposition of Arsenic at each CBAD site.

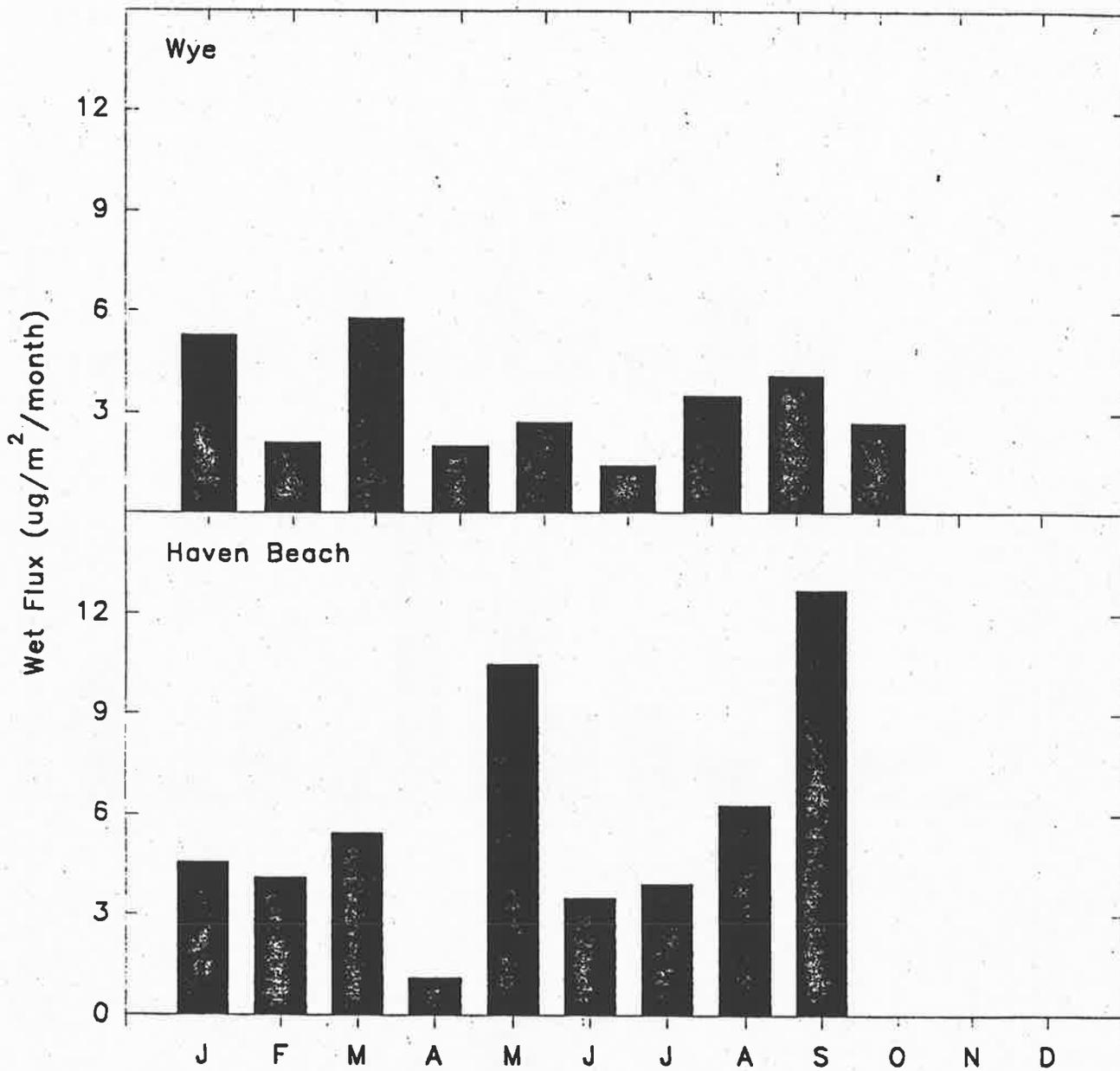
Cadmium, 1992



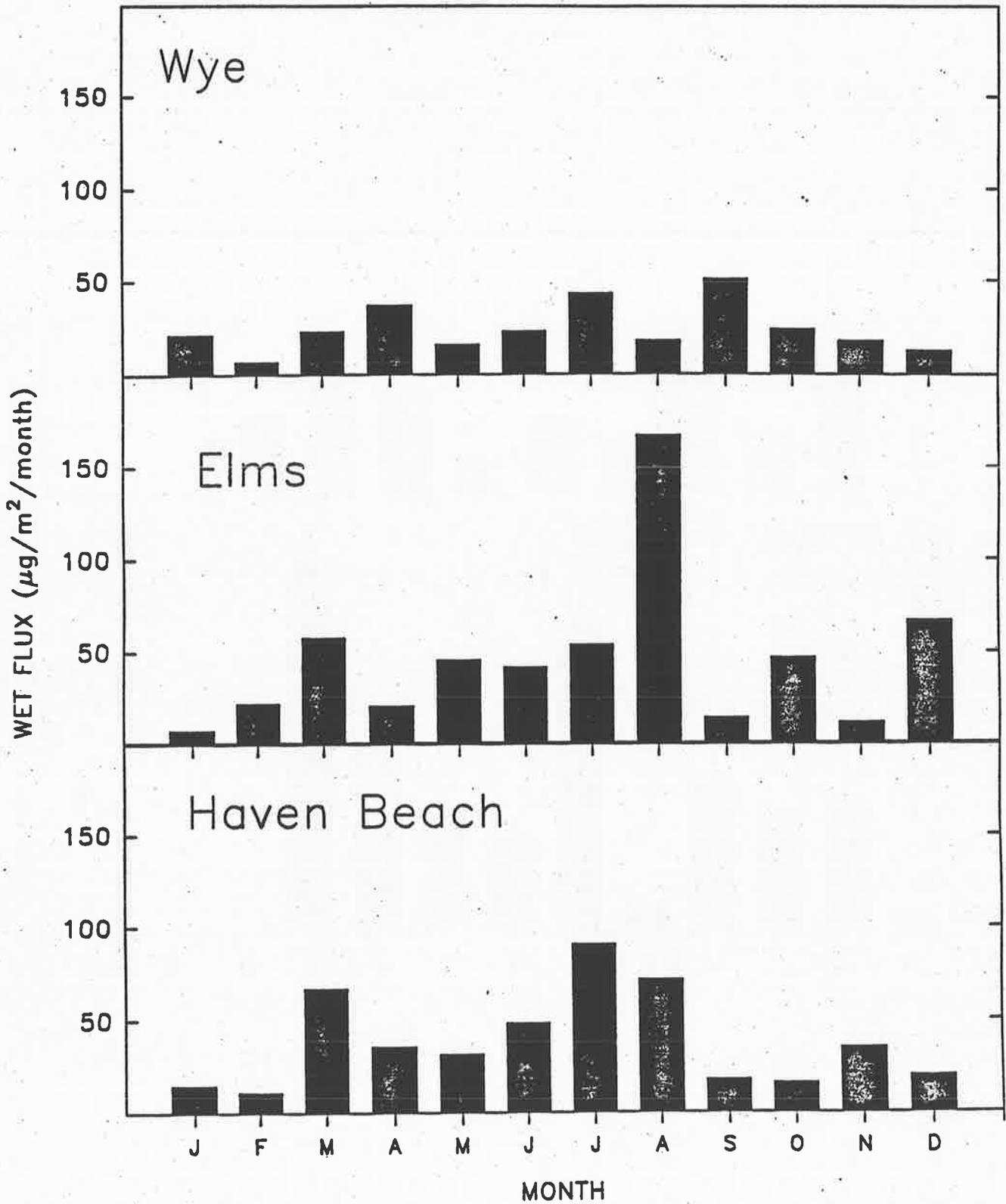
A1.73 Monthly integrated wet deposition of Copper at each CBAD site.



Cadmium, 1993

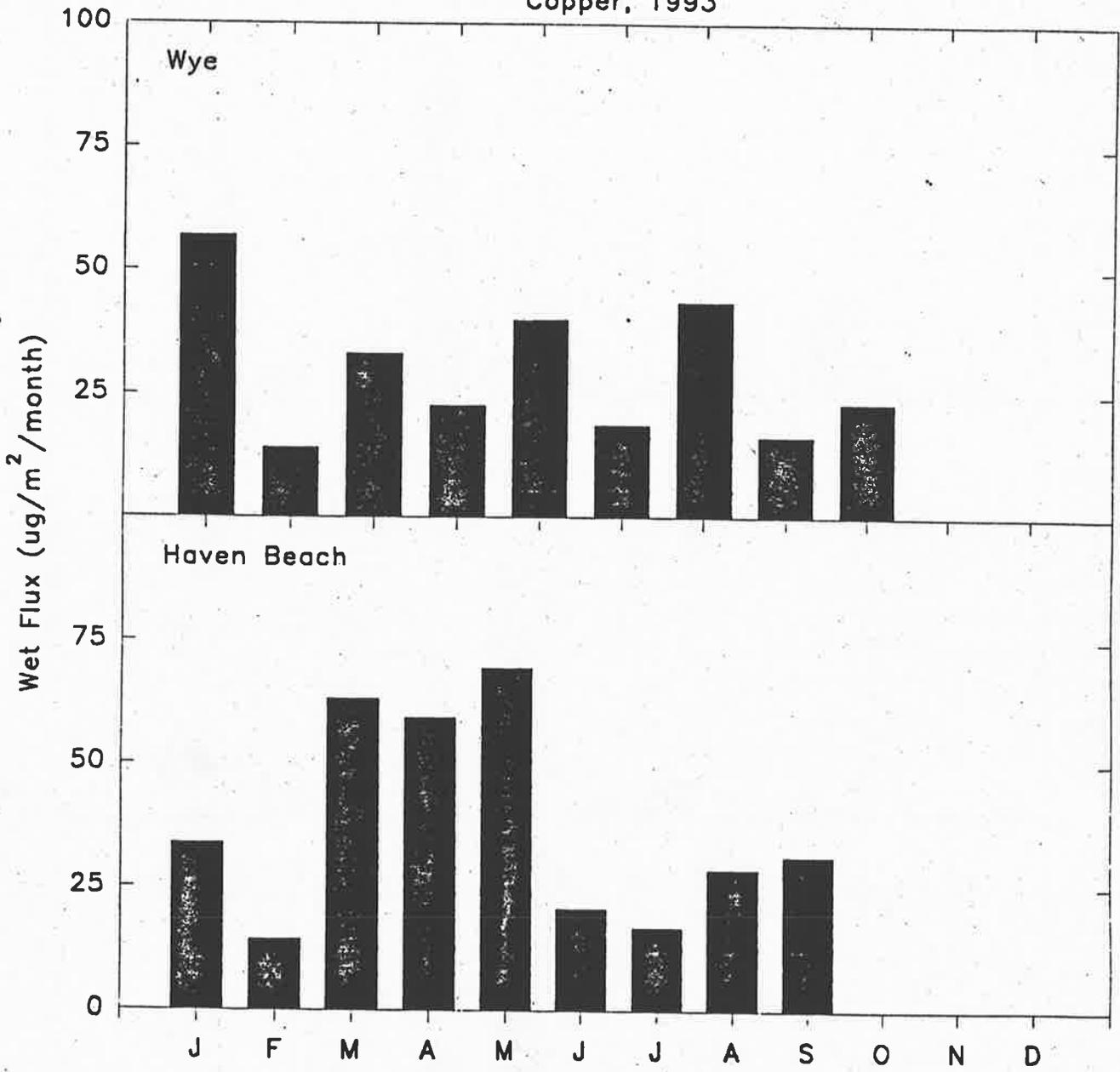


Copper, 1992

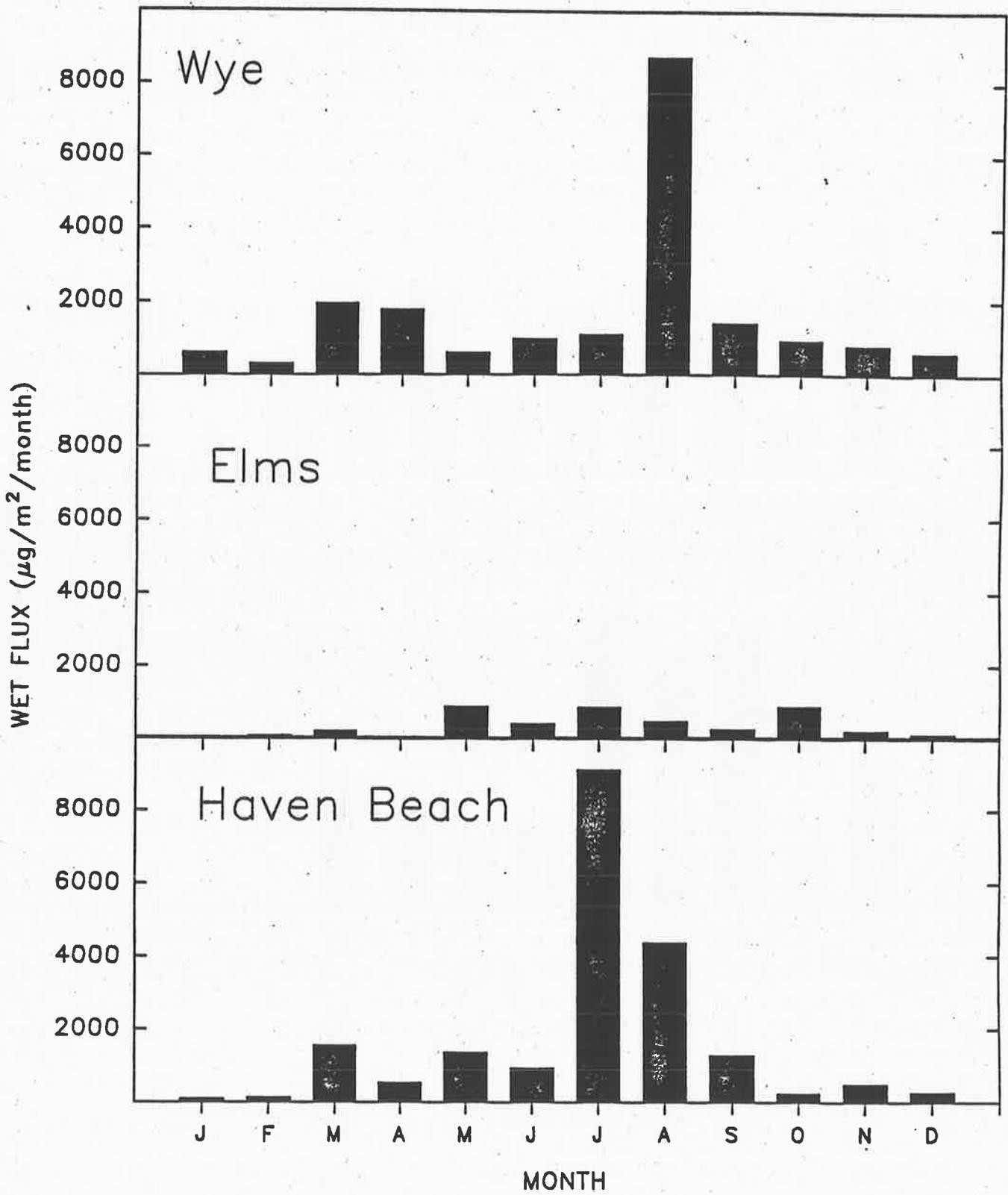


A1.74 Monthly integrated wet deposition of Copper at each CBAD site.

Copper, 1993

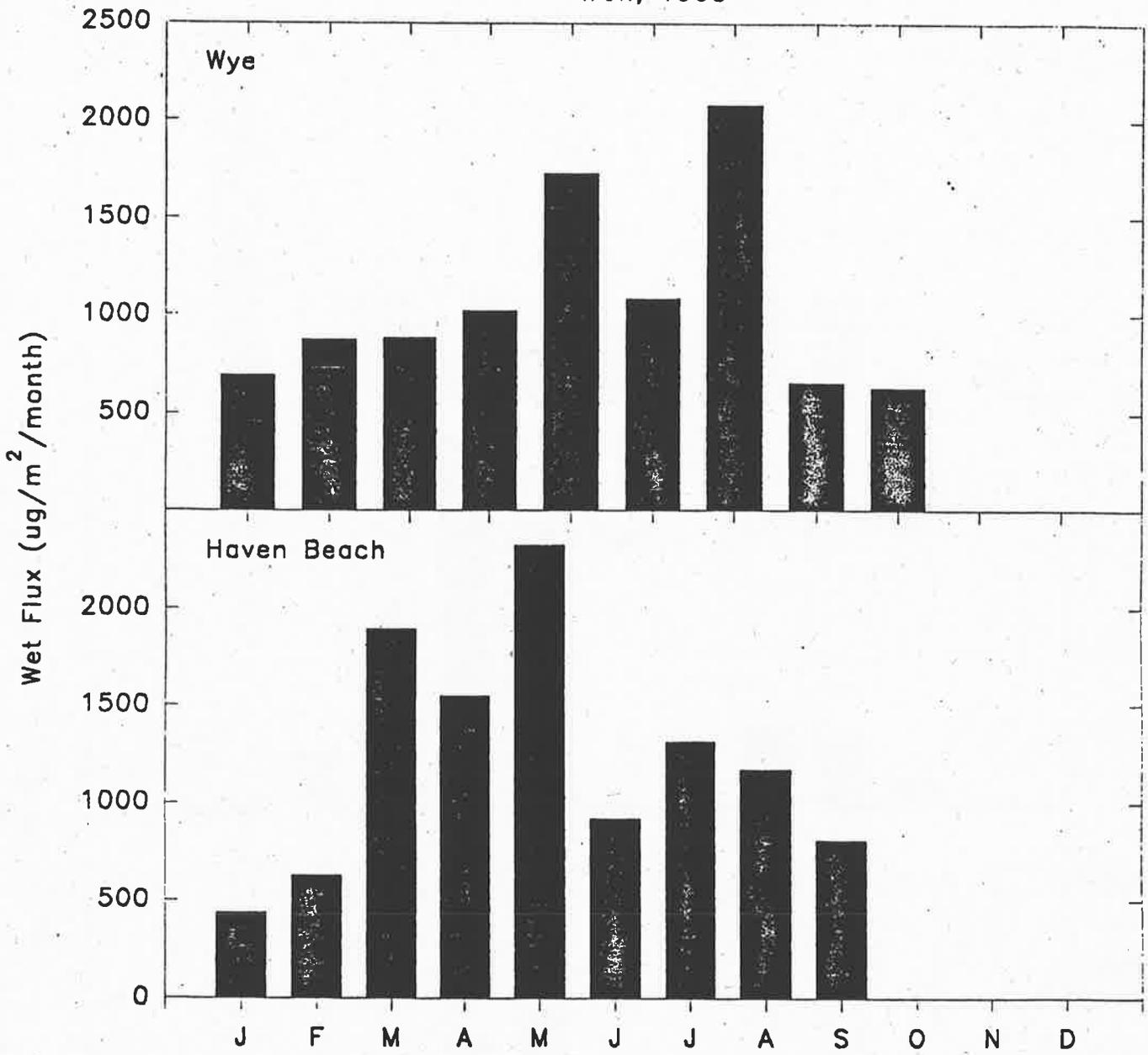


Iron, 1992

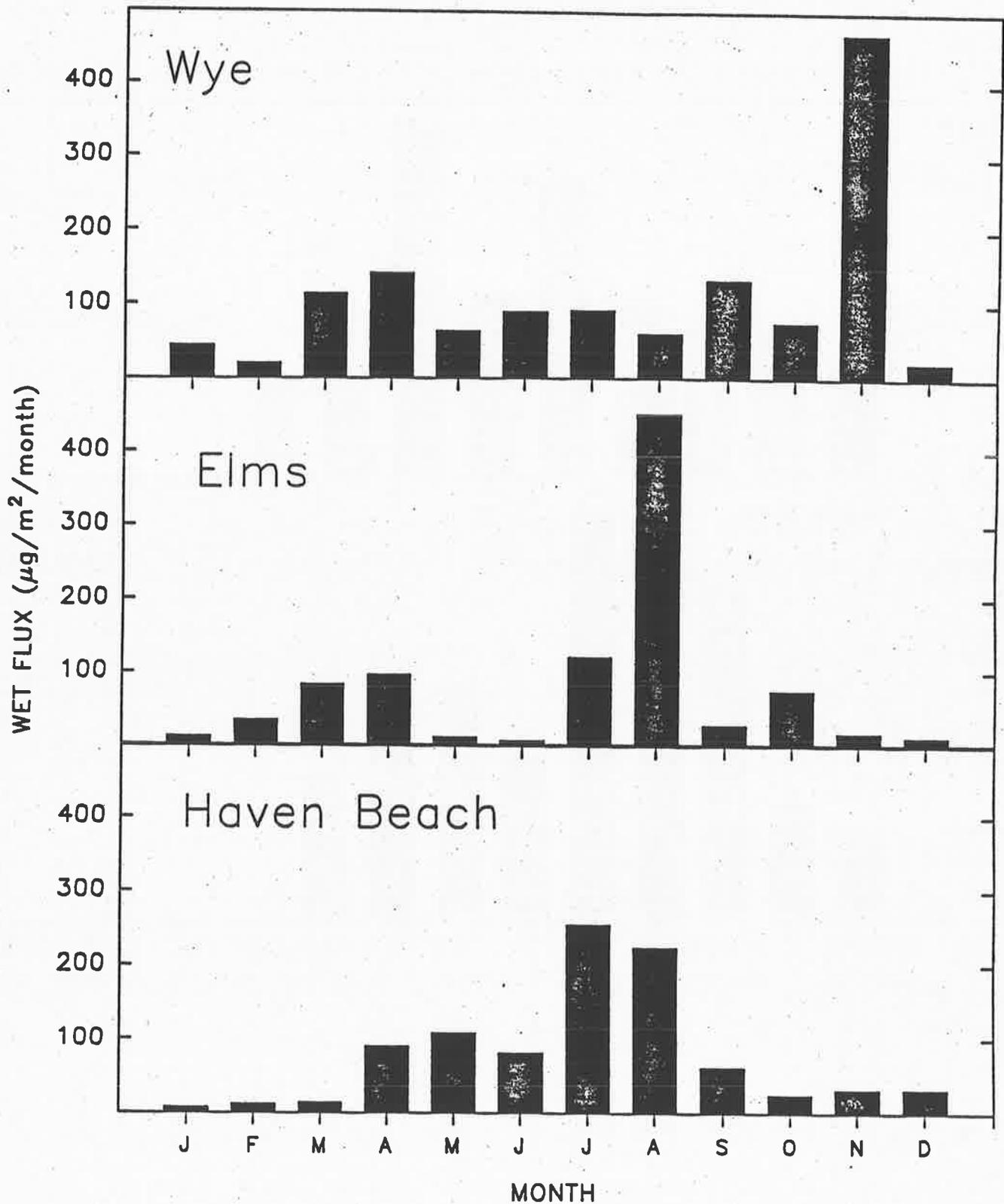


A1.75 Monthly integrated wet deposition of Iron at each CBAD site.

Iron, 1993

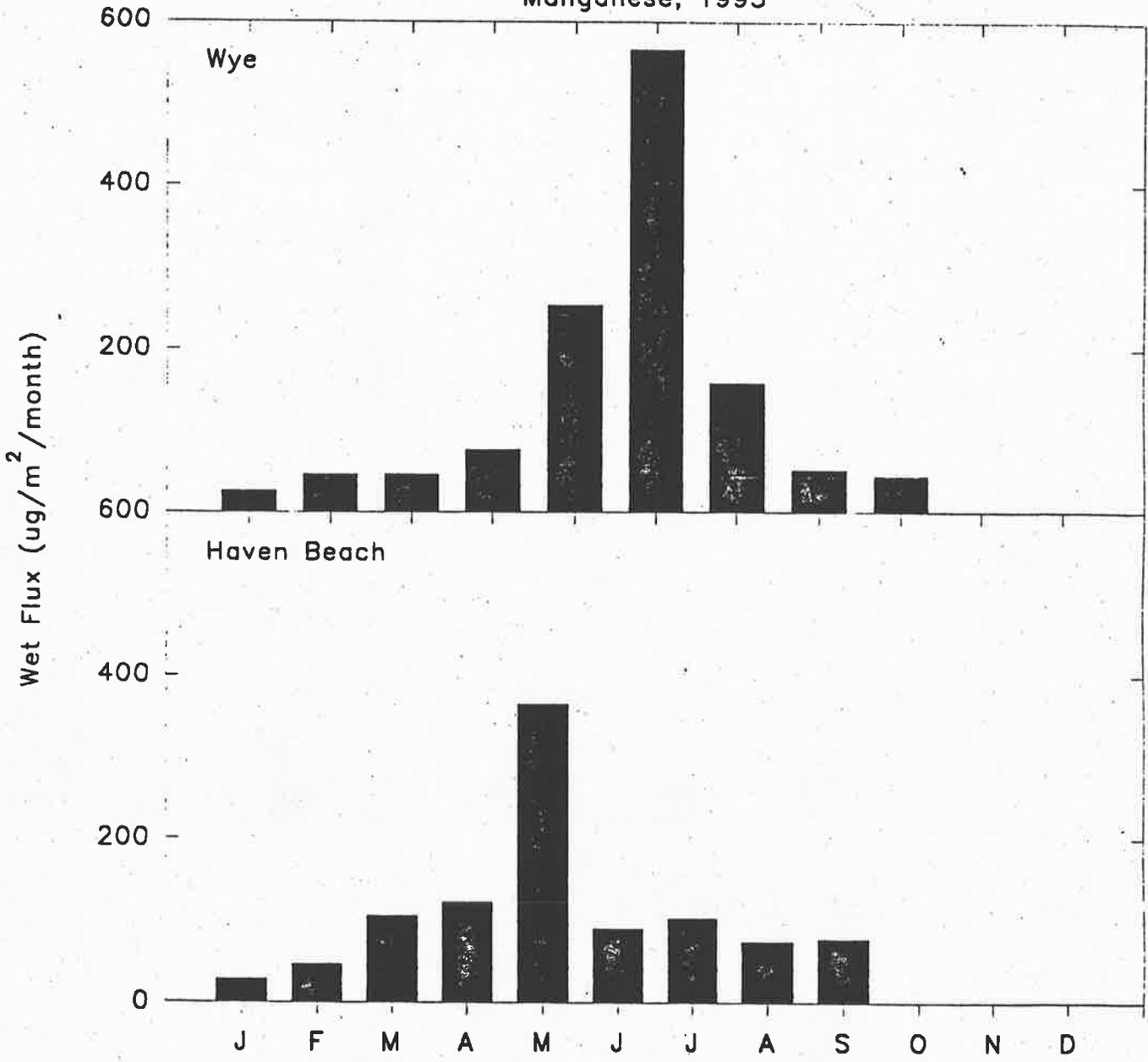


Manganese, 1992

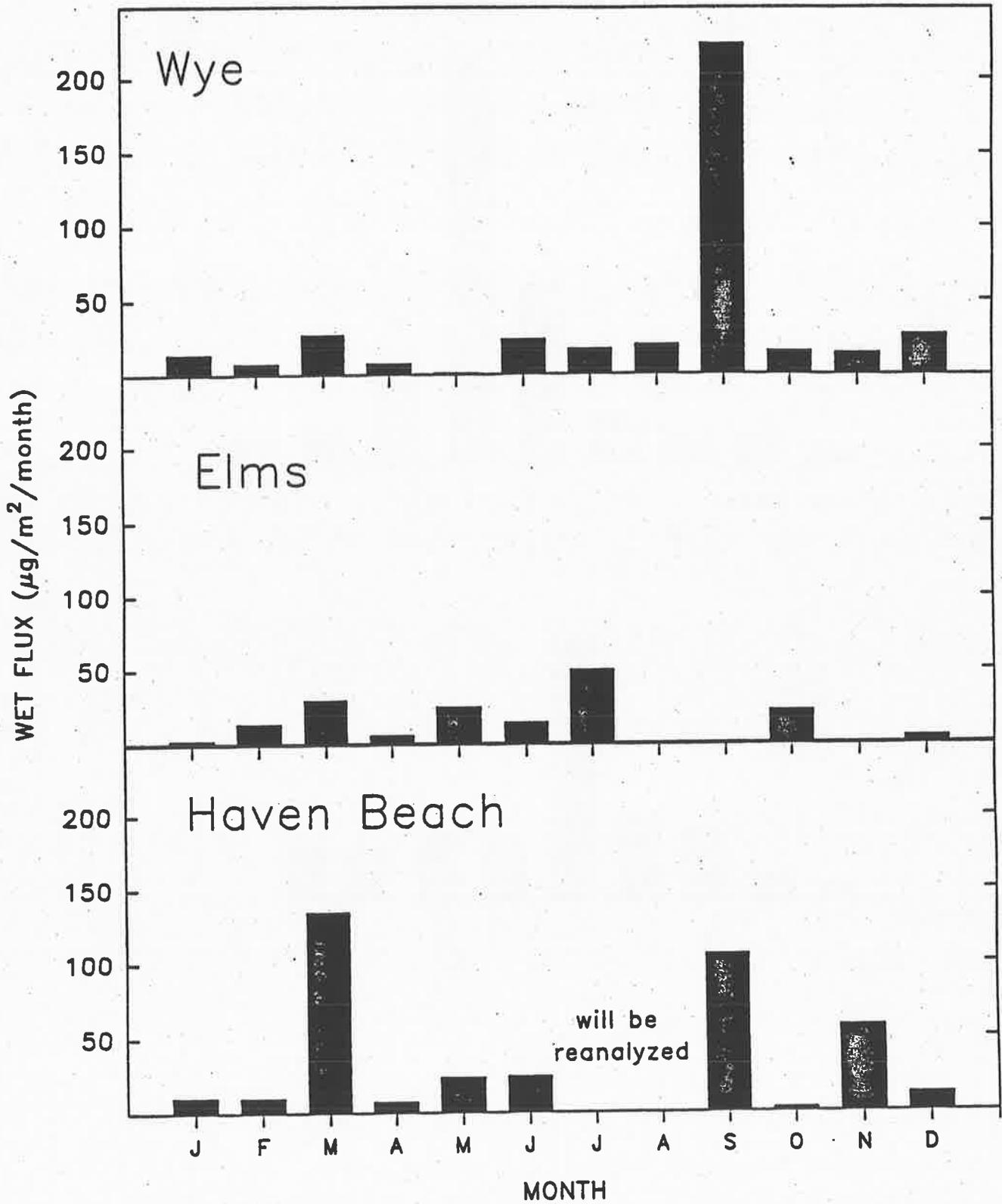


A1.76 Monthly integrated wet deposition of Manganese at each CBAD site.

Manganese, 1993

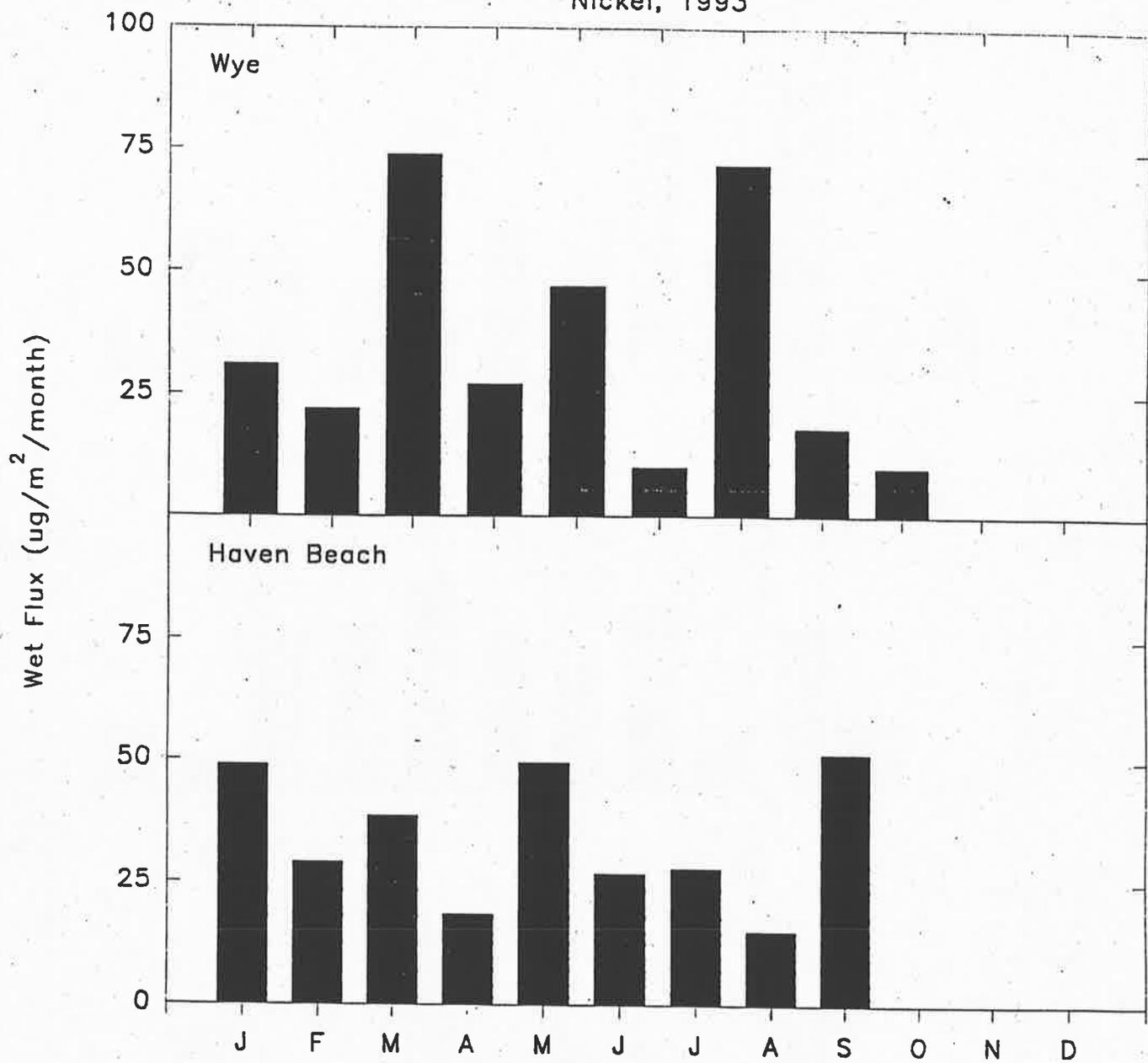


Nickel, 1992

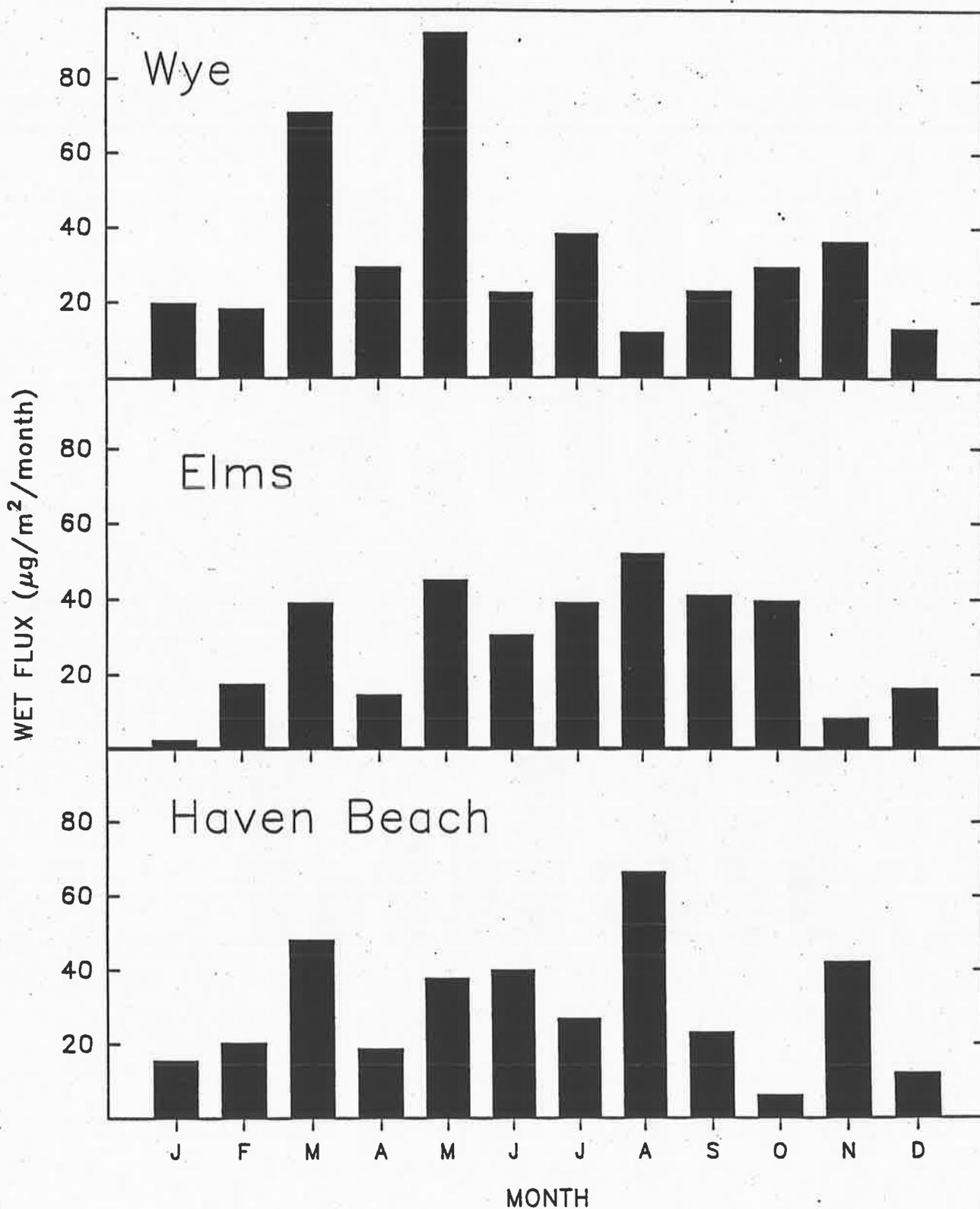


A1.77 Monthly integrated wet deposition of Nickel at each CBAD site.

Nickel, 1993

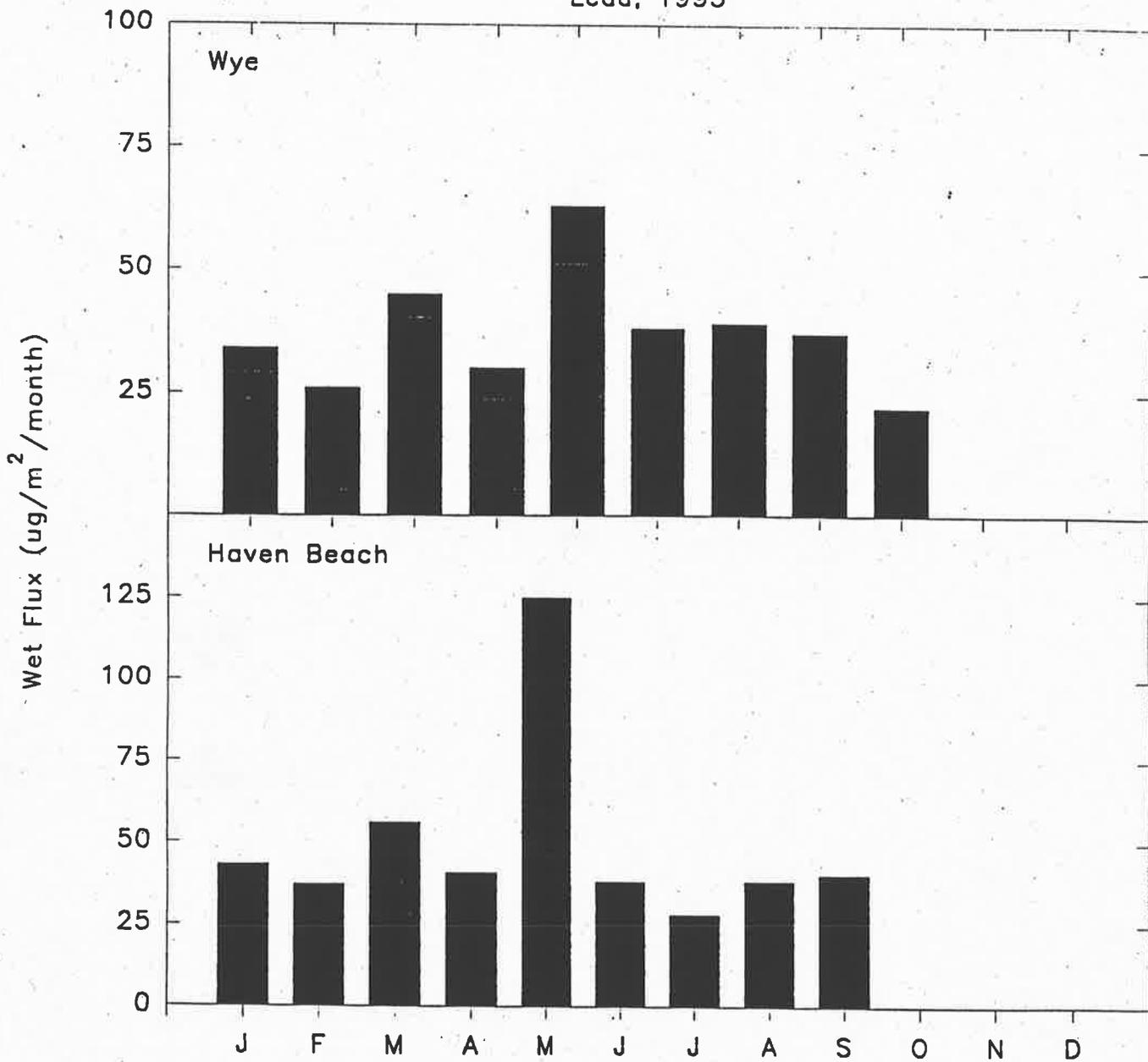


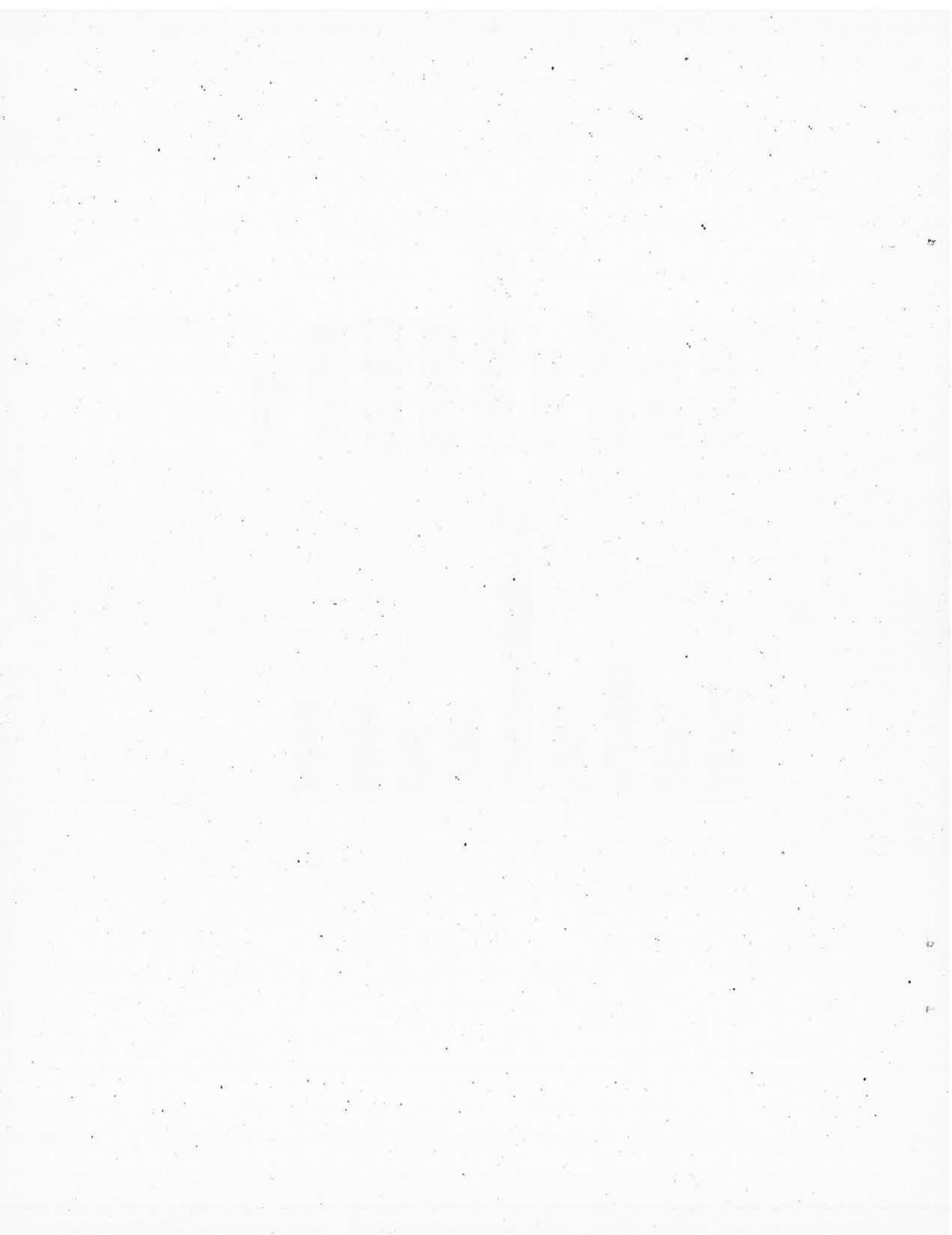
Lead, 1992



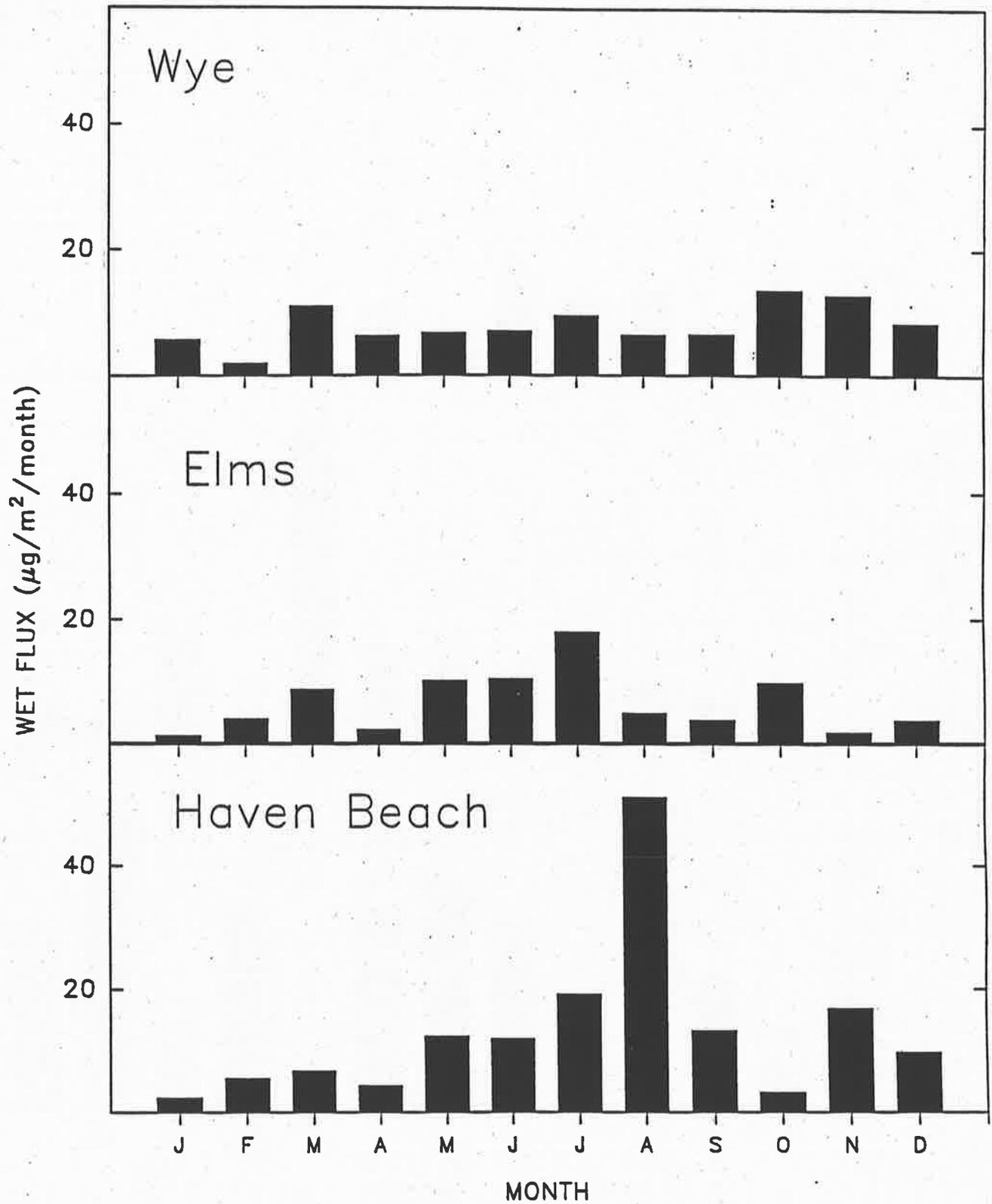
A1.78 Monthly integrated wet deposition of Lead at each CBAD site.

Lead, 1993



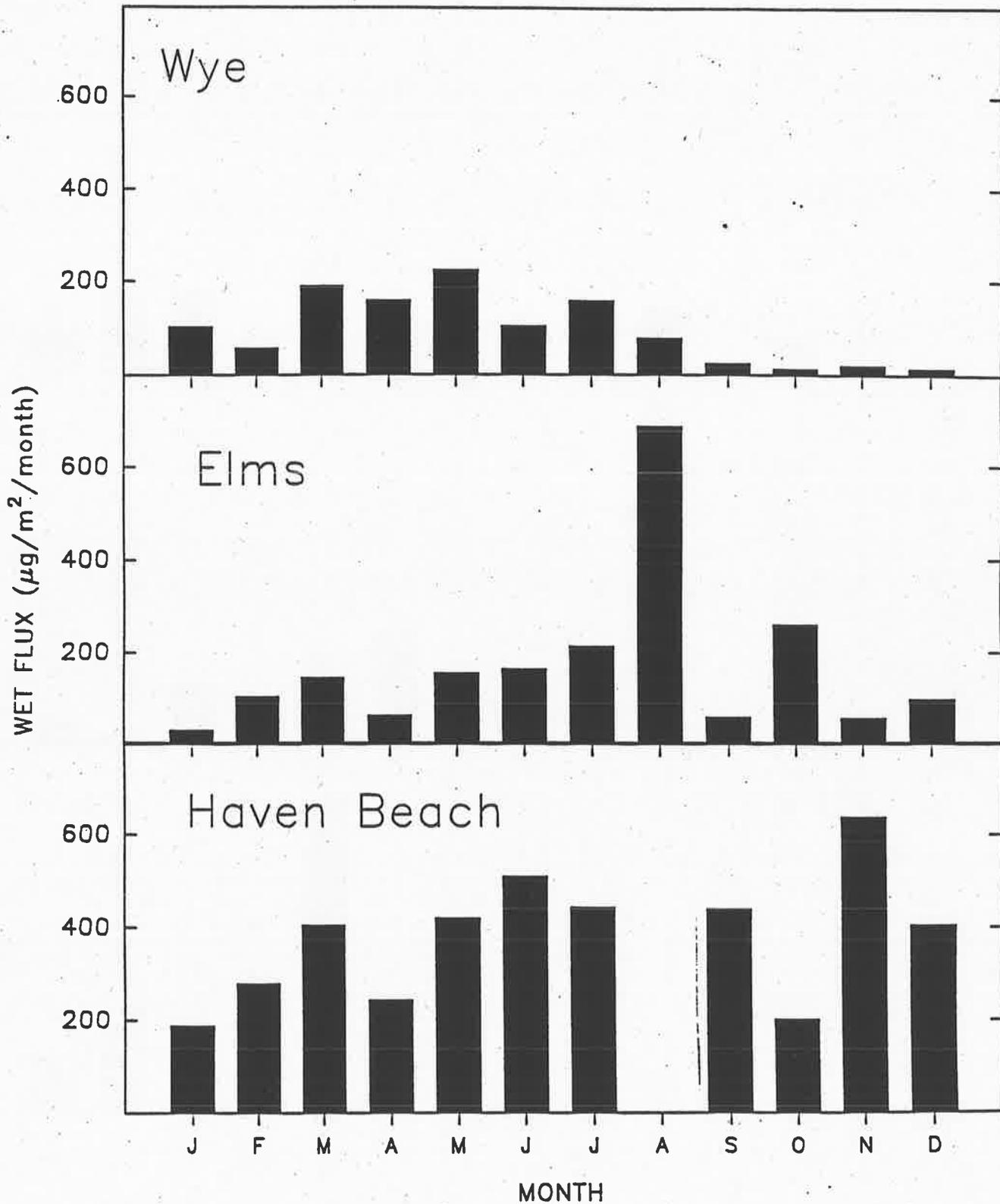


Selenium, 1992



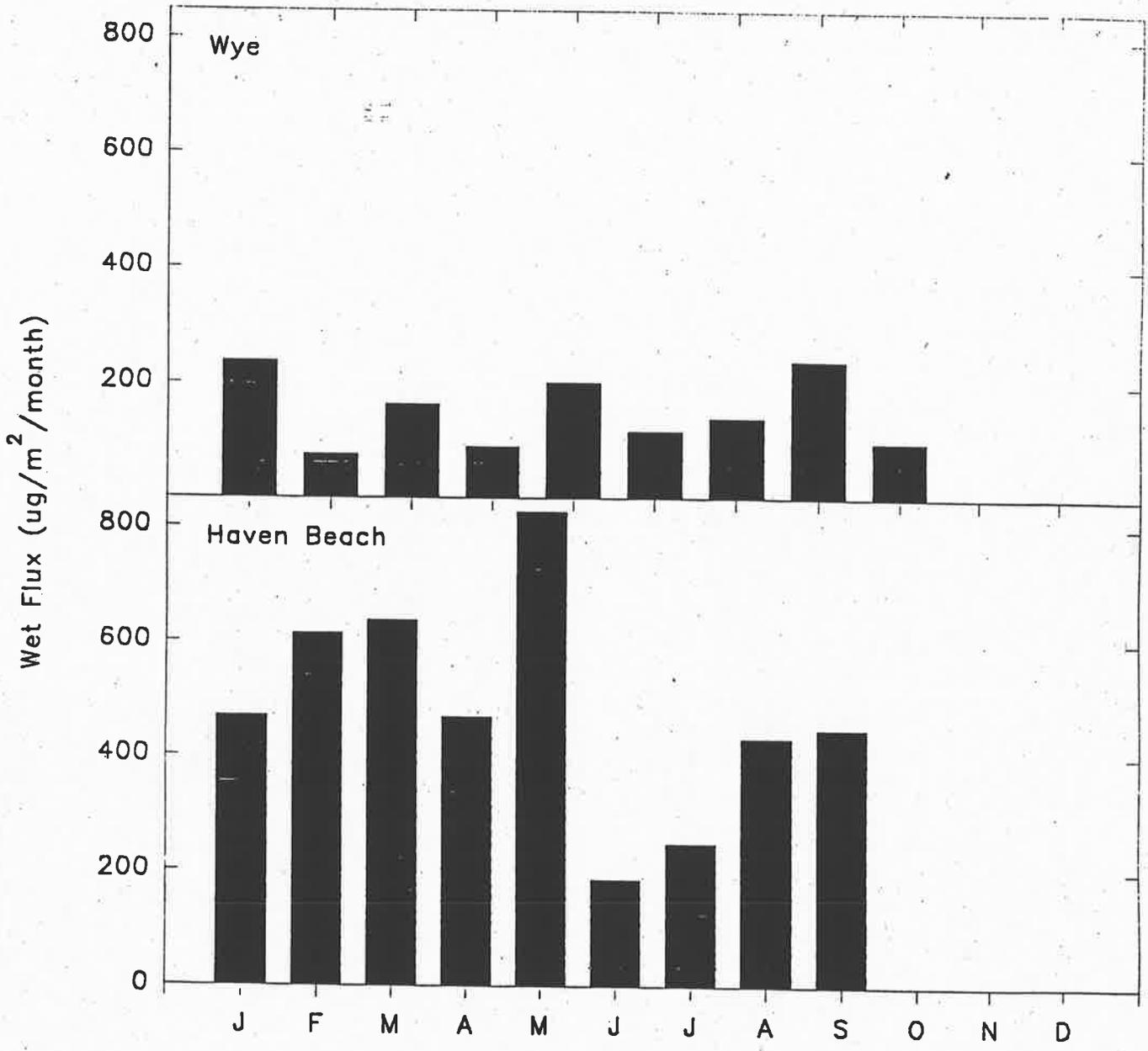
A1.79 Monthly integrated wet deposition of Selenium at each CBAD site.

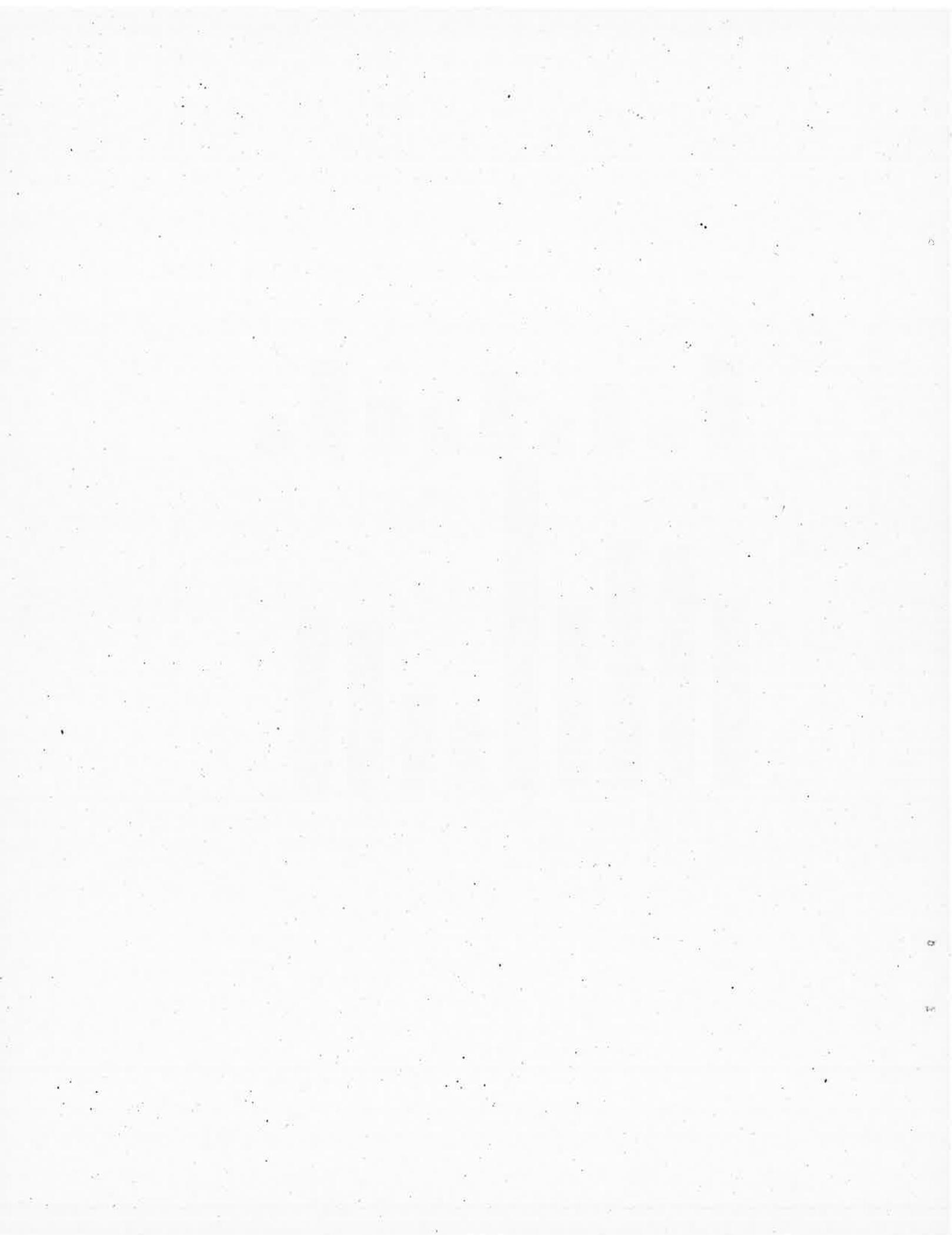
Zinc, 1992



A1.80 Monthly integrated wet deposition of Zinc at each CBAD site.

Zinc, 1993





Appendix A2

SampleID	Date	Week	Vol.	Fu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	B[b]F	B[k]F	B[a]P	Fl[1,23]P	B[a]A	B[ghi]P
908044F	1/15/92		2	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908045F	1/20/92		4	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908046F	2/12/92		6	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908047F	2/26/92		8	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908048F	3/11/92		10	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908049F	3/25/92		12	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908050F	4/07/92	1101	14	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.	n.q.
908051F	4/21/92	879	16	0.1462	1.5552	N.D.	0.2211	0.3017	N.D.	0.0230	0.0016	0.0013	0.0014	0.0004	0.0003	0.0007
908052F	5/05/92		18	0.4444	1.7680	N.D.	0.2108	0.3238	N.D.	0.0085	N.Q.	0.0009	0.0011	0.0004	0.0004	0.0005
908053F	5/19/92	1093	20	0.1480	2.0398	N.D.	0.4089	0.8372	N.D.	0.0152	0.0008	0.0007	0.0008	0.0005	0.0002	0.0009
908054F	6/03/92	922	22	0.4300	1.9734	N.D.	0.2035	0.8971	N.D.	0.0119	0.0009	0.0008	0.0002	N.D.	N.D.	N.D.
908055F	6/17/92	1077	24	0.1977	1.5488	N.D.	0.4815	0.8405	N.D.	0.0136	N.Q.	0.0020	0.0011	0.0005	N.D.	0.0008
908056F	7/01/92	924	26	N.Q.	1.7402	N.D.	0.8427	0.8682	0.0013	0.0205	N.Q.	0.0014	0.0010	0.0004	N.D.	0.0005
908057F	7/15/92	1100	28	N.Q.	0.0083	0.0499	0.2042	2.8737	0.0013	0.0217	0.8098	0.0014	0.0010	0.0082	0.0044	N.Q.
908058F	7/29/92	728	30	0.1719	1.4164	0.0000	NQ	0.3352	0.0001	0.0173	0.0019	0.0010	NQ	NQ	NQ	NQ
908059	8/12/92	983	32	0.1810	1.3688	N.D.	NQ	0.3905	0.0001	0.0276	0.0035	0.0016	NQ	NQ	NQ	0.0018
908061F	9/09/92	878	36	0.8227	2.8470	N.D.	0.2361	0.1150	N.D.	0.0084	0.0020	0.0010	NQ	NQ	NQ	0.0018
908062F	9/23/92	935	38	0.4928	2.7403	0.0530	NQ	0.5976	0.0024	0.0390	0.0064	0.0002	0.0004	0.0007	0.0004	0.0018
908063F	10/07/92	866	40	0.3888	1.8948	0.0498	NQ	0.2282	0.0049	0.0145	0.0003	0.0002	NQ	NQ	NQ	NQ
908064F	10/20/92	854	42	0.6905	1.8624	N.D.	0.3765	0.4108	N.D.	0.0118	0.0007	0.0008	NQ	NQ	NQ	NQ
908065F	11/03/92	739	44	0.7558	1.5624	N.D.	0.1744	0.1288	N.D.	0.0080	0.0007	0.0007	N.D.	N.D.	N.D.	0.0023
908066F	11/17/92	746	46	0.7778	0.9230	N.D.	N.D.	N.D.	N.D.	0.0100	0.0008	0.0007	N.D.	N.D.	N.D.	0.0008
908067F	11/17/92	822	48	8.7105	10.7430	0.4071	1.1890	0.8223	0.0123	0.0876	0.0099	0.0064	0.0040	0.0044	0.0044	0.0058
908068F	12/01/92	648	50	1.3784	2.0873	N.D.	0.2144	0.1143	0.0037	0.0976	0.0099	0.0064	0.0071	0.0044	0.0014	0.0014
908069F	12/28/92	482	52	7.8581	11.7721	0.8686	1.3128	0.9982	0.0403	0.1165	0.0103	0.0099	0.0017	0.0022	0.0014	0.0014

SampleID	Date	Week	Vol.	Fu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	B[a]P	B[a]F	B[a]P	B[a]P	Flu[2,3]P	a[1,2]A	B[ghi]P
906044F	1/15/92		2	0.00876	0.18912	0.01133	0.16642	0.13167	0.06382	0.12358	0.11284	0.10153	0.09230	0.06787	0.07756	0.01221	0.06177
906045F	1/30/92		4	0.00734	0.15762	0.01014	0.16220	0.17470	0.11050	0.30520	0.36163	0.26672	0.31763	0.16735	0.37130	0.02769	0.43061
906046F	2/12/92		6	0.00283	0.17107	0.01359	0.19698	0.17520	0.07074	0.16152	0.19756	0.15615	0.15615	0.10740	0.16338	0.02335	0.14953
906047F	2/26/92		6	0.00334	0.03360	0.00272	0.02204	0.02076	0.02123	0.05146	0.07928	0.06471	0.07769	0.02554	0.07019	0.00808	0.09855
906048F	3/11/92		10	0.00334	0.03659	0.00175	0.03456	0.02699	0.01319	0.03387	0.04150	0.03342	0.03768	0.01821	0.03624	0.00552	0.03463
906049F	3/25/92		12	0.01047	0.17719	0.00734	0.19073	0.14752	0.06440	0.14538	0.18040	0.12231	0.13368	0.08137	0.12967	0.01770	0.11762
906050F	4/07/92	1101		0.00704	0.06025	0.00488	0.10963	0.09828	0.03118	0.06029	0.16884	0.10271	0.12023	0.05301	0.09589	0.01316	0.06967
906051F	4/21/92	670		0.00222	0.01318	0.00057	0.06814	0.06923	0.02234	0.06930	0.06814	0.00483	0.00528	0.00258	0.00404	0.00072	0.00492
906052F	5/05/92		18	N.Q.	0.01233	0.00123	0.03338	0.03554	0.00731	0.02758	0.02877	0.01740	0.02055	0.00692	0.01348	0.00203	0.01371
906053F	5/19/92		20	N.Q.	0.01802	0.00689	0.01876	0.02215	0.00340	0.02014	0.02113	0.00801	0.01487	0.00457	0.00776	0.00105	0.00912
906054F	6/17/92	922		N.Q.	0.01586	0.00144	0.00704	0.00853	0.00329	0.00812	0.00932	0.00733	0.00670	0.00415	0.00863	0.00151	0.00668
906055F	7/01/92	1077		N.Q.	0.02851	0.00111	0.02317	0.02802	0.00365	0.01621	0.02545	0.00903	0.00870	0.00474	0.00730	0.00151	0.00668
906056F	7/15/92	924		N.Q.	0.02720	0.00205	0.03070	0.02490	0.00620	0.01824	0.02143	0.01488	0.01722	0.00474	0.00730	0.00151	0.00668
906057F	7/29/92	1109		N.Q.	0.02774	0.00192	0.02435	0.02588	0.00598	0.01836	0.02021	0.01228	0.01916	0.00822	0.01457	0.00053	0.00942
906058F	8/12/92	728		nd	0.02490	0.00320	0.01470	0.02240	0.00710	0.02250	0.02021	0.01228	0.01916	0.00822	0.01457	0.00053	0.00942
906059	8/26/92	983		nd	0.01640	0.00320	0.00760	0.00970	0.00540	0.01250	0.01470	0.00630	0.01230	0.00660	0.00460	0.00000	0.01030
906061F	9/09/92	573		nd	0.01962	0.00101	0.00766	0.00576	0.00356	0.00743	0.00715	0.00420	0.00630	0.00400	0.00600	nd	0.00900
906062F	9/23/92	935		0.00204	0.00760	0.00101	0.00766	0.00576	0.00356	0.00743	0.00715	0.00420	0.00630	0.00400	0.00600	0.00126	0.01018
906063F	10/07/92	36		0.00060	0.00760	0.00090	0.00420	0.00370	0.00300	0.00660	0.00810	0.00410	0.00560	0.00302	0.00717	0.00126	0.01018
906064F	10/20/92	866		0.00160	0.02030	0.00140	0.01350	0.02130	0.00800	0.02970	0.04460	0.01870	0.03600	0.00070	0.00350	nd	0.00510
906065F	11/03/92	654		0.00424	0.06067	0.00313	0.08108	0.06636	0.02319	0.07177	0.09948	0.05549	0.03993	0.00070	0.02410	nd	0.03710
906066F	11/17/92	739		0.00612	0.04384	0.00266	0.03544	0.03321	0.01089	0.02873	0.03890	0.03007	0.03319	0.01538	0.05510	0.00766	0.05722
906067F	11/30/92	748		0.01048	0.16223	0.00945	0.19898	0.16991	0.06405	0.21293	0.25101	0.14288	0.03319	0.01538	0.02768	0.00357	0.03777
906068F	12/01/92	822		0.02160	0.30146	0.03174	0.31439	0.29776	0.30556	0.66042	0.91054	0.64585	0.76319	0.6742	0.18416	0.02271	0.15920
906069F	12/15/92	648		0.00538	0.06208	0.00314	0.03721	0.04101	0.00768	0.06672	0.13370	0.12056	0.11915	0.08150	0.16471	0.02037	0.14228
906069F	12/29/92	462		0.01682	0.18731	0.02497	0.15908	0.16462	0.27291	0.49413	0.76965	0.63496	0.73869	0.36327	0.68825	0.09063	1.03982

PAH Concentrations (ng/m³) in Air PUF-Haven Beach, 1992

SampleID	Date	Week	Flu	Phen	Ant	Flt	Pyr	B[a]A	Chrysene	B[a]F	B[k]F	B[a]P	Fl[1,2,3]P	B[a]P	B[ghi]P
AB	1/15/92	2	0.9122	1.9132	N.D.	N.Q.	0.1205	N.D.	0.0101	0.0017	0.0010	0.0014	0.0009	0.0009	0.0031
AC	1/30/92	4	2.2102	2.7791	0.1653	0.2795	0.2226	0.0157	0.1050	0.0061	0.0050	0.0180	N.D.	N.D.	0.0033
AD	2/12/92	6	1.5622	2.6492	N.D.	0.4687	0.2499	0.0158	0.0681	0.0244	0.0191	0.0281	0.0144	0.0011	0.0009
AE	2/28/92	8	0.9747	1.6140	N.D.	0.2180	0.1053	N.D.	0.0387	0.0020	0.0017	0.0025	0.0004	N.Q.	0.0026
AF	3/11/92	10	0.8070	3.3298	N.D.	0.3377	0.4530	N.D.	0.1313	0.0084	0.0052	0.0119	0.0080	0.0024	0.0007
AG	3/25/92	12	0.8628	1.0912	N.D.	0.1578	0.1578	N.D.	0.0814	N.Q.	N.Q.	0.0047	0.0021	N.Q.	N.Q.
AH	4/08/92	14	0.8457	2.5839	N.D.	0.3779	0.3952	N.D.	0.0234	N.Q.	N.Q.	0.0009	N.D.	N.Q.	0.0007
AI	4/22/92	16	0.5813	2.3036	N.D.	0.4518	0.7638	N.D.	0.1029	0.0156	0.0129	0.0311	0.0178	0.0177	0.0452
AJ	5/06/92	18	0.2839	0.6859	N.D.	0.0725	0.0856	N.D.	0.0123	0.0010	0.0004	0.0009	0.0176	N.Q.	0.0041
AK	5/21/92	20	0.2954	1.4995	N.Q.	0.2476	0.5880	N.D.	0.0098	0.0007	0.0004	N.Q.	N.Q.	N.Q.	N.Q.
AL	6/03/92	22	N.Q.	N.Q.	N.Q.	0.9521	2.4629	N.D.	0.0327	0.0012	0.0010	N.D.	N.D.	N.D.	N.Q.
AM	6/17/92	24	0.2848	2.6982	N.D.	0.3119	0.3260	N.D.	0.0359	0.0008	0.0007	0.0013	N.Q.	0.0005	0.0010
AN	7/02/92	26	4.7220	16.5090	N.D.	1.9580	3.8198	N.D.	0.8780	0.0249	0.0204	0.0262	0.0120	0.0154	0.1218
AO	7/15/92	28	0.6961	5.5090	N.D.	2.2109	4.4042	N.D.	0.0794	0.0016	0.0013	0.0015	0.0004	N.D.	0.0018
AP	7/29/92	30	0.2135	2.6522	N.D.	0.6874	1.7166	N.D.	0.0223	0.0007	0.0008	0.0015	0.0003	N.D.	0.0083
AQ	8/12/92	32	0.2025	4.2525	N.D.	1.1549	3.5277	N.D.	0.0403	0.0007	0.0008	N.Q.	0.0003	0.0002	0.0015
AR	8/26/92	34	N.D.	N.D.	N.D.	N.Q.	N.D.	N.D.	0.0548	0.0015	0.0012	0.0028	0.0016	0.0017	0.0068
AS	9/09/92	36	0.5003	3.6098	N.D.	0.7393	2.8130	N.D.	0.0437	0.0011	0.0009	0.0012	N.D.	N.D.	0.0009
AT	9/25/92	38	0.3065	0.4652	N.D.	1.6288	3.6475	N.D.	0.0429	0.0024	0.0019	0.0019	0.0006	0.0007	0.0013
AU	10/07/92	40	0.2036	0.6581	N.D.	0.1448	0.2733	N.D.	0.0085	0.0010	0.0008	0.0008	0.0002	0.0003	0.0006
AV	10/21/92	42	0.8304	1.9920	N.D.	0.2029	0.1687	N.D.	0.0046	0.0005	0.0004	0.0009	N.D.	N.D.	0.0004
AW	11/04/92	44	0.6751	1.9568	N.D.	0.5325	0.6571	N.D.	0.0129	0.0026	0.0021	0.0022	0.0005	0.0007	0.0010
AX	11/18/92	46	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.
AY	12/02/92	48	2.0124	2.9378	N.D.	0.3302	0.2559	N.D.	0.0208	N.Q.	N.Q.	N.Q.	N.Q.	N.D.	N.Q.
AZ	12/16/92	50	1.3721	1.7732	N.D.	0.3161	0.2845	0.0039	0.0178	0.0029	0.0024	0.0024	N.D.	0.0003	0.0009
AA	12/31/92	52	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.	N.A.
			21.51	66.08	0.17	13.79	27.80	0.04	1.68	0.10	0.08	0.14	0.08	0.08	0.23
			0.8272	2.5022	0.0064	0.5304	1.0615	0.0014	0.0840	0.0039	0.0031	0.0083	0.0022	0.0022	0.0060

SampleID	Date	Week	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	BbF	BkF	BeP	BiP	Fl23P	B[a]A	B[a]P
AB	1/15/92	2	N.Q.	0.03958	0.00169	0.03052	0.02898	0.01607	0.02270	0.03207	0.02880	0.02781	0.02020	0.03724	0.00630	0.05193
AC	1/30/92	4	0.00740	0.12110	0.00971	0.09853	0.10115	0.19063	0.20865	0.40161	0.37015	0.33653	0.36749	0.58186	0.09431	0.49947
AD	2/12/92	6	0.00297	0.03748	0.00192	0.03473	0.02768	0.01480	0.02970	0.05361	0.03950	0.03955	0.02515	0.06323	0.00662	0.04498
AE	2/28/92	8	0.00334	0.07817	0.00438	0.06252	0.05395	0.03069	0.05229	0.10124	0.07648	0.07778	0.05300	0.11037	0.01745	0.09073
AF	3/11/92	10	0.00315	0.02193	0.00148	0.01119	0.00954	0.00483	0.01220	0.02124	0.01255	0.01654	0.00695	0.01408	0.00211	0.01608
AG	3/25/92	12	0.00353	0.02458	0.00163	0.01258	0.01070	0.00553	0.01368	0.02378	0.01411	0.01853	0.00779	0.01587	0.00237	0.01602
AH	4/08/92	14	0.00401	0.05142	0.00254	0.04735	0.04115	0.01374	0.07833	0.08782	0.07165	0.08009	0.02325	0.05680	0.00753	0.05519
AI	4/22/92	16	N.Q.	0.02001	N.D.	0.00709	0.01213	0.00283	0.01083	0.00551	0.00413	0.00550	0.00213	0.00300	0.00066	0.00362
AJ	5/06/92	18	N.Q.	0.01503	N.D.	0.00816	0.00731	0.00154	0.01597	0.01207	0.00987	0.00929	0.00281	0.00855	0.00075	0.00538
AK	5/21/92	20	N.Q.	N.Q.	N.D.	0.00722	0.01245	0.00154	0.00815	0.01040	0.00597	0.00638	0.00208	0.00568	0.00062	0.00453
AL	6/03/92	22	N.Q.	0.02470	0.00100	0.00180	0.04059	0.00280	0.02301	0.01803	0.01050	0.01597	0.00314	0.00568	0.00075	0.00453
AM	6/17/92	24	N.Q.	0.03069	0.00068	0.04823	0.00378	0.00393	0.03118	0.02805	0.01543	0.01597	0.00314	0.00568	0.00062	0.00453
AN	7/02/92	26	N.Q.	N.Q.	N.D.	0.12887	0.11517	0.04862	0.17127	0.07542	0.05353	0.02018	0.00501	0.00869	0.00111	0.00959
AO	7/15/92	28	N.Q.	0.02638	0.00157	0.01731	0.02251	0.00412	0.01721	0.02805	0.01543	0.02018	0.00501	0.00869	0.00111	0.00959
AP	7/29/92	30	N.Q.	0.01690	0.00107	0.01250	0.02158	0.00334	0.01361	0.00822	0.00839	0.01314	0.00468	0.04399	0.00852	0.05410
AQ	8/12/92	32	N.Q.	0.01924	N.D.	0.00823	0.01085	0.00374	0.01361	0.00822	0.00575	0.00730	0.00312	0.00876	0.00142	0.01186
AR	8/26/92	34	N.Q.	0.01685	0.00103	0.01082	0.02363	0.00274	0.01145	0.00524	0.00428	0.00658	0.00251	0.00570	0.00112	0.00722
AS	9/09/92	36	0.00273	0.01629	N.D.	0.01213	0.02740	0.00338	0.01500	0.00713	0.00448	0.00718	0.00305	0.00347	0.00060	0.00499
AT	9/25/92	38	N.Q.	0.02418	0.00120	0.01275	0.02371	N.D.	0.01685	0.00836	0.00448	0.00718	0.00305	0.00472	0.00078	0.00568
AU	10/07/92	40	N.Q.	0.04189	0.00206	0.01275	0.02371	0.00435	0.01825	0.00836	0.00684	0.00718	0.00305	0.00472	0.00078	0.00568
AV	10/21/92	42	N.Q.	0.05239	0.00273	0.02663	0.08489	0.00584	0.02736	0.01194	0.00908	0.01471	0.00384	0.00513	0.00085	0.00844
AW	11/04/92	44	N.Q.	0.02219	0.00110	0.06252	0.06877	0.01528	0.04142	0.03645	0.02982	0.03925	0.00823	0.01482	0.00138	0.01853
AX	11/18/92	46	N.Q.	0.11601	0.00898	0.02001	0.03165	0.01091	0.03884	0.06583	0.03131	0.03602	0.01891	0.04008	0.04356	0.04880
AY	12/02/92	48	N.Q.	0.08344	0.00448	0.13289	0.11971	0.04808	0.18020	0.18207	0.04881	0.03725	0.01893	0.05053	0.00799	0.04598
AZ	12/16/92	50	N.Q.	0.03125	0.00188	0.08313	0.08055	0.04653	0.12625	0.18435	0.15083	0.12449	0.06348	0.10829	0.01741	0.10565
AA	12/31/92	52	N.Q.	0.03046	0.00188	0.03870	0.04139	0.03419	0.08288	0.14859	0.13321	0.16173	0.08950	0.15517	0.01704	0.17351
						0.02710	0.02885	0.01978	0.07310	0.06048	0.05870	0.07051	0.02278	0.07205	0.00800	0.08287

Concentrations (ng/m³) of PAHs in Air - Vapor - Haven Beach, 1993

Sample ID	Date	Week	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chr	B[b]F	B[k]F	B[a]P	B[<i>a</i>]P	I[123]P	D[<i>a</i>]A	B[ghi]P
AC	1/12/93	2	1.37116	2.90879	0.90879	N.D.	0.33687	0.28431	0.00547	0.03842	0.00547	N.D.	0.00504	N.D.	N.D.	N.D.
AD	1/28/93	4	0.38082	0.92511	0.92511	N.D.	0.12985	0.05787	N.D.	0.00710	0.00062	0.00057	N.D.	N.D.	N.D.	N.D.
AE	2/09/93	6	0.73512	0.83014	0.83014	N.D.	0.08420	0.05584	N.Q.	0.01172	0.00121	0.00089	N.Q.	N.Q.	N.D.	N.D.
AF	2/23/93	8	0.52785	0.80588	0.80588	N.D.	0.06814	0.01310	N.D.	0.00343	N.D.	N.D.	N.D.	N.D.	N.D.	N.D.
AG	3/09/93	10	0.57703	0.86840	0.86840	N.D.	0.05450	0.05985	N.D.	0.00865	0.00172	0.00172	0.00148	N.D.	N.D.	0.00350
AH	3/23/93	12	0.52210	1.94445	1.94445	N.D.	0.11853	0.17595	N.D.	0.01573	N.D.	N.Q.	0.00330	N.D.	N.D.	0.00241
AI	4/07/93	14	0.86502	0.84078	0.84078	N.D.	0.06598	0.06437	N.Q.	0.00818	N.D.	N.D.	0.00127	N.D.	N.D.	N.D.
AJ	4/20/93	16	1.04388	1.15126	1.15126	N.D.	0.06098	0.07182	N.Q.	0.01427	N.D.	N.D.	0.00127	N.D.	N.D.	N.D.
AK	5/04/93	18	0.13953	0.72174	0.72174	N.D.	0.06282	0.11583	N.Q.	0.00358	N.Q.	N.Q.	N.D.	N.D.	N.D.	N.Q.
AL	5/18/93	20	0.38358	1.77881	1.77881	N.D.	0.13805	0.39838	N.Q.	0.03207	0.00157	0.00128	0.00421	N.Q.	N.D.	0.00237
AM	6/01/93	22	0.38753	N.Q.	N.Q.	N.Q.	0.10485	0.31088	N.Q.	0.00475	0.00020	0.00018	0.00421	N.Q.	N.D.	N.D.
AN	6/15/93	24	0.50880	1.92341	1.92341	N.D.	0.23782	0.42805	N.Q.	0.02742	0.00065	0.00054	0.00116	N.Q.	N.D.	N.D.
AO	6/29/93	26	0.37327	2.58224	2.58224	N.Q.	0.27121	0.70049	N.Q.	0.05327	0.00033	0.00027	0.00112	N.Q.	N.D.	N.Q.
AP	7/13/93	28	0.44019	3.43888	3.43888	0.08460	1.16813	2.68110	N.D.	0.03873	N.D.	N.D.	N.D.	N.D.	N.D.	N.D.
AQ	7/27/93	30	0.64408	3.68888	3.68888	0.05284	0.96347	1.32502	N.D.	0.03205	0.00163	0.00139	0.00163	N.D.	N.D.	0.00087
AR	8/11/93	32	0.24235	1.86810	1.86810	0.06116	0.89235	1.32557	N.D.	0.01561	N.D.	N.D.	N.D.	N.Q.	N.D.	N.D.
AS	8/24/93	34	0.41418	1.02108	1.02108	0.02483	0.16433	0.25872	N.Q.	0.02844	0.00105	0.00041	0.00105	N.Q.	N.D.	0.00108
AT	9/07/93	36	1.33433	3.80746	3.80746	0.30458	0.50383	0.81434	0.01023	0.04071	0.00430	0.00351	0.00647	N.D.	0.00241	0.00686
AU	9/21/93	38	N.D.	N.D.	N.D.	N.Q.	N.Q.	N.Q.	N.D.	0.05208	N.Q.	N.Q.	N.Q.	N.Q.	N.Q.	0.00090
AV	10/05/93	40	0.80978	1.76158	1.76158	0.20053	N.Q.	0.28283	N.Q.	N.Q.	N.Q.	0.00062	0.00062	N.Q.	N.D.	0.00055

Concentrations (ng/m³) of PAHs In Air - Filter - Haven Beach, 1993

Sample ID	Date	Week	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chry	B[b]F	B[k]F	B[a]P	B[a]P	Fl[1]P	D[ah]A	B[ghi]P
AC	1/12/93	2	0.00325	0.02605		N.D.	0.01211	0.01260	0.01189	0.05571	0.11577	0.11577	0.06585	0.01387	0.05930	0.06476
AD	1/26/93	4	0.00357	0.05003		N.D.	0.05389	0.04230	0.02329	0.06666	0.07664	0.04370	0.06426	0.04370	0.11637	0.06476
AE	2/09/93	6	0.00315	0.07150		N.D.	0.05935	0.04274	0.01236	0.08762	0.08372	0.02063	0.07617	0.02063	0.05975	0.07618
AF	2/23/93	8	0.00495	0.10691		N.D.	N.D.	0.00168	0.04695	0.10639	0.12979	0.06942	0.06942	0.05013	0.07973	0.05228
AG	3/09/93	10	0.00254	0.07513		N.Q.	0.04001	0.04017	0.03632	0.07875	0.10461	0.05673	0.06942	0.05013	0.07973	0.05228
AH	3/23/93	12	0.00268	0.02862		N.D.	0.02539	0.01983	0.00353	0.02049	0.01875	0.00770	0.01794	0.00676	0.01265	0.08729
AI	4/07/93	14	0.00605	0.03560		N.D.	0.03428	0.03684	0.00904	0.02889	0.02882	0.02150	0.02783	0.01326	0.02253	0.01507
AJ	4/20/93	16	0.00507	0.02166		N.Q.	0.01617	0.01569	N.Q.	0.02119	0.01458	0.00740	0.01653	0.00740	0.01174	0.03119
AK	5/04/93	18	0.00293	0.01309		N.D.	0.00756	0.01293	N.Q.	0.00608	0.00209	0.00246	0.00246	0.00080	0.00072	0.00175
AL	5/18/93	20	0.00348	0.01644		N.Q.	0.00982	0.01285	0.00274	0.01143	0.00774	0.00580	0.00781	0.00399	0.00583	0.00792
AM	6/01/93	22	0.00631	0.03861		N.Q.	0.04903	0.07874	0.00877	0.03183	0.02919	0.00947	0.00947	0.00352	0.02449	0.02546
AN	6/15/93	24	0.00537	0.02100		N.Q.	0.01342	0.01785	0.00308	0.01378	0.00792	0.00326	0.00326	0.00352	0.00607	0.00771
AO	6/29/93	26	N.Q.	0.01739		N.Q.	0.00871	0.01359	0.00308	0.01086	0.00493	0.00512	0.00512	0.00211	0.00298	0.00379
AP	7/13/93	28	N.Q.	N.Q.		N.Q.	0.01724	N.D.	N.Q.	0.01682	0.00365	0.00299	0.00491	0.00173	0.00156	0.00379
AQ	7/27/93	30	N.Q.	0.01838		N.Q.	0.01665	0.01843	0.00284	0.01717	0.01644	0.01005	0.00491	0.00173	0.00156	N.Q.
AR	8/11/93	32	0.00236	0.00874		N.D.	0.00456	0.00849	N.Q.	0.00576	0.00304	0.00197	0.00313	0.00462	0.01035	N.Q.
AS	8/24/93	34	0.00844	0.02669		N.D.	0.01553	0.01511	0.00489	0.01274	0.01535	0.01072	0.01608	0.00605	0.00182	0.01179
AT	9/07/93	36	0.03818	0.07449		N.D.	0.01834	0.02801	0.01501	0.05602	0.09116	0.05681	0.09044	0.00787	N.D.	0.00260
AU	9/21/93	38	0.01475	0.03809		N.Q.	0.01399	0.01680	0.00320	0.01453	0.00850	0.00778	0.01047	0.00408	0.06200	0.06091
AV	10/05/93	40	0.01554	0.04196		N.Q.	0.02583	0.02410	0.00651	0.01806	0.01749	0.01431	0.01881	0.00986	0.00575	0.00850

Concentrations (ng/m³) of PAHs in Air - Vapor - Elms, 1993

Sample ID	Date	Week	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chry	B[b]F	B[k]F	B[e]P	B[a]P	I[123]P	D[ghi]A	B[ghi]P
BV	1/12/93	2	2.09083	8.40199	0.12877	0.92196	0.42157	0.00966	0.04590	0.00572	0.00581	0.00193	0.00035	0.00072	N.D.	0.00111
BW	1/26/93	4	3.07978	4.08630	0.10841	0.80952	0.24210	N.Q.	0.02502	0.00727	0.00739	0.00282	0.00092	0.00131	N.D.	0.00350
BX	2/08/93	6	1.04211	1.89077	N.D.	0.13371	0.04528	N.D.	0.00734	0.00200	0.00164	0.00189	0.00103	0.00129	N.D.	0.00183
BY	2/23/93	8	0.91698	1.08192	N.D.	0.05308	0.03233	0.00611	0.01179	0.00845	0.00773	0.00905	0.00538	0.00518	N.D.	0.00786
BZ	3/09/93	10	0.23327	0.23311	N.D.	0.01037	0.00587	N.Q.	0.00042	N.Q.	N.Q.	N.D.	N.D.	N.D.	N.D.	N.Q.
BAA	3/23/93	12	N.A.													
BAB	4/07/93	14	0.46278	0.49600	N.D.	0.05018	0.03467	N.Q.	0.00087	0.00087	0.00055	0.00089	0.00034	N.D.	N.D.	N.Q.
BAC	4/20/93	16	0.35717	1.09244	N.D.	0.14962	0.10038	N.Q.	0.00742	0.00093	0.00068	0.00072	0.00030	N.Q.	N.D.	0.00113
BAD	5/04/93	18	0.53656	1.72129	N.D.	0.20099	0.19482	0.00102	0.00903	0.00258	0.00211	0.00308	0.00157	N.Q.	N.D.	0.00555
BAE	5/18/93	20	0.26584	1.06102	N.D.	0.07933	0.09786	N.Q.	0.00216	N.Q.	N.Q.	0.00045	N.D.	N.D.	N.D.	N.Q.
BAF	6/01/93	22	0.00870	0.01468	N.D.	0.00124	0.00134	N.Q.	0.00037	N.D.	N.D.	N.Q.	N.D.	N.D.	N.D.	0.00129
BAG	6/15/93	24	0.20949	0.51127	N.D.	0.02803	0.02652	N.D.	0.00094	N.Q.	N.Q.	N.Q.	N.Q.	N.D.	N.D.	N.D.
BAH	6/29/93	26	1.24248	4.37940	N.D.	N.Q.	N.D.	N.D.	0.01312	N.Q.	N.Q.	N.Q.	N.Q.	N.D.	N.D.	N.D.
BAI	7/13/93	28	0.35137	3.13315	N.Q.	0.21195	0.64272	N.Q.	0.03350	0.00147	0.00120	0.00347	N.Q.	N.Q.	N.D.	0.00072
BAJ	7/27/93	30	0.45931	1.39715	N.D.	0.14603	0.10491	N.Q.	0.00445	0.00042	0.00034	0.00047	N.Q.	N.D.	N.D.	0.00028
BAK	8/10/93	32	0.30038	1.13248	N.D.	0.09540	0.15857	N.Q.	0.00309	0.00043	0.00035	0.00048	N.D.	N.D.	N.D.	N.D.
BAL	8/24/93	34	0.28680	1.11817	0.28680	0.12231	0.09263	N.Q.	0.00430	0.00039	0.00031	0.00043	N.D.	N.D.	N.D.	N.Q.
BAM	9/07/93	36	0.22731	0.84791	N.D.	0.07106	0.08428	N.Q.	0.00173	N.Q.	N.Q.	N.Q.	N.D.	N.D.	N.D.	N.Q.

Concentrations (ng/m³) of PAHs in Air - Filter - Elms, 1993

Sample ID	Date	Week	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chry	B[b]F	B[k]F	B[e]P	B[a]P	Fl[2,3]P	D[ah]A	B[ghi]P
BV	1/12/93	2	0.00972	0.10342	0.00514	0.07228	0.05548	0.03971	0.09949	0.12162	0.10680	0.11358	0.04047	0.13289	0.01463	0.13578
BW	1/26/93	4	0.00761	0.16086	N.Q.	0.15587	0.12843	0.09543	0.29873	0.37110	0.30363	0.29438	0.18364	0.39849	0.04545	0.34372
BX	2/08/93	6	0.00998	0.16771	N.Q.	0.19328	0.11586	0.07983	0.18248	0.15354	0.12563	0.13003	0.05559	0.10084	0.00272	0.09879
BY	2/23/93	8	0.00928	0.18942	N.D.	0.17445	0.11899	0.04751	0.10776	0.07689	0.07578	0.07457	0.05785	0.11184	0.00678	0.07757
BZ	3/09/93	10	0.01041	0.11858	N.D.	0.06657	0.05318	0.02732	0.05674	0.07879	0.06808	0.07305	0.04458	0.09752	0.01101	0.06835
BAA	3/23/93	12	N.D.	0.03345	N.D.	0.03494										
BAB	4/07/93	14	0.00410	0.05786	N.D.	0.07815	0.05275	0.01336	0.05212	0.05014	0.04103	0.04986	0.02068	0.02858	0.00222	0.02958
BAC	4/20/93	16	0.00252	0.02223	N.D.	0.01818	0.01357	0.00528	0.01293	0.01588	0.01068	0.01443	0.00779	0.01488	0.00145	0.01633
BAD	5/04/93	18	0.00282	0.01943	0.00097	0.01675	0.01804	0.00291	0.00935	0.00477	0.00460	0.00590	0.00299	0.00394	N.D.	0.00672
BAE	5/18/93	20	0.00425	0.03294	N.Q.	0.02708	0.02726	0.00232	0.01014	0.01030	0.00858	0.01082	0.00400	0.00588	0.00059	0.00777
BAF	6/01/93	22	N.Q.	N.Q.	N.D.	N.Q.	N.Q.	N.Q.	0.00065	0.00083	0.00083	0.00074	0.00038	0.00034	N.D.	N.Q.
BAG	6/15/93	24	0.00846	N.Q.	0.06229	0.03640	0.03394	0.00748	0.02413	0.03753	0.02889	0.03974	0.01221	0.01553	N.D.	0.02030
BAH	6/29/93	26	N.Q.	0.04608	0.00273	0.04598	0.04595	0.01060	0.03710	0.03527	0.02186	0.03168	0.01374	0.02132	0.00338	0.02928
BAI	7/13/93	28	N.Q.	0.02741	N.Q.	0.02835	0.03630	0.00328	0.01369	0.01116	0.00714	0.01068	0.00478	0.00647	0.00068	0.00959
BAJ	7/27/93	30	N.Q.	0.02878	0.00145	0.02240	0.01823	0.00540	0.01787	0.02148	0.01594	0.02011	0.00987	0.01208	0.00210	0.01525
BAK	8/10/93	32	0.00301	0.02278	N.Q.	0.01822	0.01948	N.Q.	0.00856	0.00811	0.00436	0.00764	0.00307	0.00507	N.D.	0.00952
BAL	8/24/93	34	0.00225	0.01744	N.D.	0.00982	0.00780	0.00290	0.00884	0.01045	0.00641	0.01138	0.00472	0.00810	0.00131	0.01099
BAM	9/07/93	36	0.00320	0.03117	N.D.	0.01715	0.01976	0.00257	0.01048	0.01121	0.00817	0.01277	0.00437	0.00512	N.Q.	0.00642

PAH Concentrations (ng/L) in Resins-Wye, 1992

SampleID	Date	Week	Vol. (L)	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	B[b]F	B[k]F	B[a]P	B[ghi]P	Fl[123]P	d[12h]A	B[ghi]P
072892WR	7/14/92-7/28/92		30	1.595	6.804	0.399	4.315	2.897	0.287	1.100	0.673	0.541	0.550	nd	nd	nd	nd
081192WR	7/28/92-8/11/92		32	0.895	3.297	0.441	3.055	1.448	0.148	0.917	0.495	0.274	0.330	0.126	0.120	nd	0.448
082592WR	8/11/92-8/25/92		34	0.834	2.147	0.181	1.119	0.463	0.004	0.308	0.240	nd	0.107	nd	0.131	nd	nd
090892WR	8/25/92-9/08/92		36	0.740	2.289	0.122	1.932	0.933	1.268	1.308	nd	nd	nd	nd	nd	nd	nd
092292WR	9/08/92-9/22/92		38	1.083	3.997	0.440	5.253	1.773	0.515	0.529	nd	nd	nd	nd	nd	nd	nd
WMFBR*	9/22/92-10/06/92		40	nd--	0.1090	nd	0.1380	0.0577	nd	nd	nd	nd	nd	nd	nd	nd	nd
JEB blk*				0.51	nd	nd	4.38	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd
JKR blk*				1.01	5.78	nd	1.06	0.81	0.31	0.58	1.40	0.44	nd	4.52	1.09	nd	16.29

* lab blank, ng
 * field blank, no precipitation, assume 60 L
 -- not quantifiable, i.e. blank mass > 33% sample mass
 -- -- not detected, i.e. no peak

PAH Concentrations (ng/L) in Filters-Wye, 1992

SampleID	Date	Week	Vol. (L)	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	B[b]F	B[k]F	B[a]P	B[ghi]P	Fl[123]P	d[12h]A	B[ghi]P
072892WF	7/14/92-7/28/92		30	0.106	1.038	0.067	1.842	1.348	0.268	0.805	0.423	0.000	0.841	ND	0.217	ND	0.565
081192WF	7/28/92-8/11/92		32	0.184	1.733	ND	2.794	2.188	0.458	1.197	0.934	ND	1.358	ND	0.908	ND	0.928
090892WF	8/25/92-9/08/92		36	0.208	2.223	0.173	3.357	2.486	0.508	1.762	1.398	2.291	0.259	0.000	0.000	0.000	0.818
092292WF	9/08/92-9/22/92		38	0.278	4.119	0.238	5.198	4.081	1.075	3.308	4.347	1.976	2.252	0.000	0.000	0.784	0.878
704081F*				ND--	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
704082F*				ND	ND	ND	ND	ND	ND	ND	ND						
7040FFB*				ND	ND	ND	ND	ND	ND	ND	ND						

* lab blank, ng
 -- -- not detected, i.e. no peak

PAH Concentrations (ng/L) in Filters-Heaven Beach, 1992

SampleID	Date	Week	Vol.(L)	Flu	Phen	Ant	Flu	Pyr	B[a]A	Chrysene	B[b]F	B[k]F	B[a]P	B[a]P	Fl[23]P	d[1h]A	B[ghi]P
NH001F	1/1/92-1/4/92	2	16	ND	0.451	ND	0.700	0.891	0.425	ND	ND	ND	ND	ND	ND	ND	ND
NH002F	1/14/92-1/29/92	4	16.02	ND	0.911	0.000	1.330	1.075	0.270	0.965	2.846	ND	ND	0.944	ND	ND	0.488
NH003F	1/29/92-2/11/92	6	0.6	ND	15.722	1.806	20.072	18.457	6.990	8.350	17.883	ND	12.386	ND	ND	ND	ND
NH004F	2/11/92-2/21/92	8	63	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
NH005F	2/27/92-3/10/92	10	22.8	ND	0.218	ND	0.236	3.21	ND	0.242	1.48	0.40	1.87	ND	ND	ND	0.85
NH006F	3/10/92-3/24/92	12	18.33	ND	1.976	0.000	4.36	0.900	0.516	0.834	2.344	2.648	2.320	10.173	0.000	0.000	0.000
NH008EF	3/24/92-3/27/92	14	32.85	0.395	2.263	0.724	3.976	3.942	ND	1.092	ND	ND	ND	ND	ND	ND	ND
NH007F	3/27/92-4/07/92	16	8.82	1.189	1.828	0.398	1.514	1.051	0.381	0.891	0.848	0.371	0.750	0.876	0.000	0.000	0.606
NH008F	4/07/92-4/21/92	18	10.18	1.839	1.437	ND	3.625	3.229	0.554	1.186	2.025	0.777	1.116	0.884	0.000	0.000	0.617
NH009F	4/21/92-5/05/92	20	55.6	0.188	0.374	0.105	0.398	0.283	0.095	0.227	0.517	ND	0.250	0.498	ND	ND	0.288
NH010F	5/05/92-5/20/92	22	27.90	0.051	0.280	0.058	0.374	0.239	0.072	0.172	0.408	ND	0.166	0.080	ND	ND	ND
NH011F	5/20/92-6/02/92	24	0.117	0.117	0.558	0.107	0.868	0.468	0.128	0.304	0.890	ND	0.258	0.652	ND	ND	0.580
NH012F	6/02/92-6/30/92	26	18.71	0.138	0.970	0.175	1.528	0.945	0.280	0.805	0.805	0.281	0.472	0.641	0.113	0.113	0.648
NH013F	6/18/92-6/30/92	28	74	0.085	0.301	0.088	0.384	0.287	0.143	0.349	0.492	ND	0.251	0.108	ND	ND	ND
NH014F	6/30/92-7/14/92	30	28.57	0.085	0.173	ND	0.147	0.118	ND	0.080	0.160	0.061	0.142	0.034	ND	ND	ND
NH015F	7/14/92-7/28/92	32	135.3	0.015	0.298	0.010	0.551	0.458	0.085	0.468	0.507	0.843	0.483	ND	ND	ND	ND
NH016F	7/28/92-8/11/92	34	113.9	0.043	0.241	0.018	0.207	0.155	0.050	0.153	0.105	0.130	0.053	0.000	0.000	0.000	0.000
NH018EF	8/11/92-8/12/92	34	42.535	0.028	0.146	0.013	0.164	0.139	0.028	0.087	0.079	0.115	0.082	ND	ND	ND	0.037
NH017F	8/12/92-8/28/92	36	142.24	0.028	0.220	0.033	0.532	0.483	0.084	0.224	0.627	ND	0.243	0.147	ND	ND	0.212
NH018F	8/28/92-9/08/92	36	63	ND	0.332	ND	0.723	0.871	0.170	0.483	0.960	0.113	0.428	0.380	ND	ND	ND
NH019F	9/08/92-9/22/92	38	20.85	ND	0.332	ND	0.723	0.871	0.170	0.483	0.960	0.113	0.428	0.380	ND	ND	ND
NH020F	9/22/92-10/06/92	40	97.3	0.000	11.948	0.000	0.000	0.000	6.770	14.124	NC-	NC-	NC-	NC-	NC-	NC-	NC-
NH021F	10/06/92-10/20/92	42	7.67	ND	0.725	0.190	1.408	1.191	0.314	0.516	0.875	0.331	0.343	0.242	ND	ND	ND
NH022F	10/20/92-11/03/92	44	23.12	0.114	1.192	0.100	3.334	0.134	0.501	1.355	2.555	0.872	0.875	1.680	1.152	ND	1.327
NH024F*	11/17/92-12/01/92	50	82.1	ND	0.374	ND	0.929	0.837	0.205	0.711	1.218	0.777	0.636	0.464	1.021	ND	0.790
NH025F	12/01/92-12/15/92	52	31.445	ND	0.458	ND	0.080	0.088	ND	ND	ND	7.844	ND	ND	ND	ND	ND
NH026F	12/15/92-12/30/92	52	31.445	ND	0.458	ND	0.080	0.088	ND	ND	ND	7.844	ND	ND	ND	ND	ND
NH028F				ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND
NH029F				ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND	ND

* lab blank, no precipitation, assume 50 L
 - field blank, no precipitation, assume 50 L
 - not quantifiable, i.e. blank mass > 33% sample mass
 - - - not detected, i.e. no peak

PAH Concentrations (ng/L) in Resins-Haven Bosch, 1992

SampleID	Date	Week	Vol.(L)	Flu	Phen	Ant	Fla	Pyr	B[a]A	Chrysene	B[bf]F	B[k]F	B[a]P	B[ef]P	B[ghi]P	Fl[1,2,3]P	d[ghi]A	B[ghi]P
NH001R	1/1/92-1/14/92	2	18	1.112	3.550	0.100	1.811	1.017	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd
NH002R	1/14/92-1/28/92	4	18.82	0.863	2.874	nd	0.849	0.327	0.056	0.157	0.049	0.062	0.062	0.062	nd	nd	nd	nd
NH003R	1/28/92-2/11/92	6	0.8	10.813	NQ	nd	14.825	10.707	6.211	16.242	nd	nd	nd	nd	nd	nd	nd	nd
NH004R	2/11/92-2/27/92	8	63	0.378	1.538	0.094	0.974	0.441	nd	0.238	nd	nd	nd	nd	nd	nd	nd	nd
NH005R	2/27/92-3/10/92	10	22.8	1.050	3.685	0.284	2.837	1.885	0.391	1.458	1.101	0.485	0.485	0.485	nd	nd	nd	0.874
NH006R	3/10/92-3/24/92	12	18.33	1.068	6.208	0.471	3.583	2.914	0.507	2.259	2.232	1.387	1.155	1.155	1.168	nd	nd	nd
NH007R	3/24/92-3/27/92	14	32.85	0.341	0.861	nd	NQ	nd	nd	NQ	0.562	0.551	0.335	0.335	nd	nd	nd	nd
NH008R	3/27/92-4/10/92	16	6.82	nd	2.081	nd	NQ	4.203	nd	1.019	nd	nd	nd	nd	nd	nd	nd	nd
NH009R	4/10/92-5/05/92	18	10.18	1.097	2.550	0.090	1.430	0.778	nd	1.019	nd	0.522	0.522	0.522	nd	nd	nd	nd
NH010R	5/05/92-5/20/92	20	55.6	0.671	0.446	0.112	1.750	1.187	0.200	0.967	0.905	0.441	0.481	0.481	0.285	0.480	nd	0.383
NH011R	5/20/92-6/02/92	22	57.27	0.446	1.493	0.090	1.582	0.931	0.092	0.327	0.170	0.161	0.074	0.074	0.083	0.167	nd	0.206
NH012R	6/02/92-6/16/92	24	27.90	0.858	1.712	0.104	0.968	0.542	0.125	0.488	nd	nd	nd	nd	nd	nd	nd	nd
NH013R	6/16/92-6/30/92	26	19.71	0.388	1.823	0.188	NQ	0.837	nd	0.275	0.145	0.138	0.369	0.369	nd	nd	nd	0.287
NH014R	6/30/92-7/11/92	28	74	0.358	1.137	0.070	1.060	0.571	0.041	0.238	0.109	0.085	0.075	0.075	0.019	0.020	nd	0.099
NH015R	7/11/92-7/28/92	30	28.57	0.422	1.116	0.122	1.042	0.585	0.060	0.348	0.228	0.160	0.131	0.131	nd	nd	nd	nd
NH016R	7/28/92-8/11/92	32	135.3	0.259	1.038	0.058	0.907	0.514	0.038	0.185	nd	nd	0.039	0.039	nd	nd	nd	nd
NH018ER	8/11/92-8/11/92	34	113.9	0.184	1.030	nd	0.935	0.580	nd	0.304	0.178	0.148	0.170	0.170	0.072	0.345	0.052	0.050
NH017R	8/11/92-8/28/92	34	42.535	0.357	0.775	nd	0.580	nd	nd	NQ	nd	nd	0.403	0.403	nd	nd	0.209	0.159
NH019R	8/28/92-9/08/92	36	142.24	0.106	0.384	nd	0.158	0.133	nd	NQ	nd	nd	0.025	0.025	nd	nd	0.054	0.030
NH019R	8/28/92-9/08/92	36	63	nd	0.683	0.282	0.655	nd	nd	NQ	nd	0.222	0.087	0.455	nd	nd	0.148	nd
NH019R	8/08/92-8/22/92	38	20.85	0.308	1.084	nd	0.737	nd	nd	NQ	0.182	0.404	0.384	0.384	nd	nd	nd	nd
NH020R	8/22/92-10/06/92	40	87.3	0.398	3.554	0.204	1.785	0.873	0.103	0.775	0.144	0.144	3.139	3.139	nd	nd	nd	nd
NH021R	10/06/92-10/20/92	42	7.87	0.457	NQ	0.131	1.232	0.828	nd	nd	0.817	0.817	nd	nd	nd	nd	nd	nd
NH022R	10/20/92-11/03/92	44	23.12	0.875	2.183	0.133	3.227	2.075	0.285	1.187	0.733	0.658	0.520	0.520	0.235	0.628	nd	0.597
NH023R	11/03/92-11/17/92	46	11.18	0.443	2.751	0.118	1.983	0.845	nd	0.478	nd	nd	nd	nd	nd	nd	nd	nd
NH024R*	11/17/92-12/01/92	48	nd	0.142	0.083	0.008	0.1370	0.0385	0.0304	0.0252	nd	nd	nd	nd	nd	nd	nd	nd
NH025R	12/01/92-12/15/92	50	82.1	1.483	8.781	0.204	8.131	4.417	0.383	1.818	0.865	0.558	0.578	0.578	0.395	0.423	0.085	0.472
NH026R	12/15/92-12/30/92	52	31.445	1.842	8.142	0.125	4.188	2.364	0.162	0.772	0.464	0.248	0.388	0.388	0.242	nd	nd	nd
NH8R*				nd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd	nd

* lab blank, ng

- field blank, no precipitation, assume 60 L

- - not quantifiable, i.e. blank mass > 33% sample mass

- - - not detected, i.e. no peak

Sample	Date	Vol. (L)	Week	Flu	Phen	Anth	Flt	Pyrene	B(a)A	Chrysene	B(a)P	B(k)P	B(e)P	Flt(a)P	B(a)P	B(a)A	B(ghi)P	B(b)P	
011303WR	022002-0172	28.002	2	0.64	2.09	0.68	1.06	0.88	0.97	0.20	0.14	0.11	0.44	0.19	0.19	0.19	0.12	0.12	0.10
020803WR	012703-2020	31.70	4	0.10	0.00	0.11	0.00	0.00	0.10	0.17	0.10	0.10	0.40	0.27	0.27	0.27	0.10	0.10	0.10
022303WR	022003-0223	1.54	8	23.70	90.00	1.81	0.00	20.74	4.33	1.02	0.10	0.40	0.40	0.40	0.40	0.40	0.40	0.40	0.40
030903WR	022003-0309	7.074	10	7.40	20.00	0.43	13.03	1.04	1.20	1.02	0.41	0.41	10.43	10.43	10.43	10.43	10.43	10.43	10.43
032303WR	030903-0323	07.912	13	1.20	4.00	0.10	2.04	1.12	0.23	0.40	0.21	0.21	0.40	0.40	0.40	0.40	0.40	0.40	0.40
040803WR	032003-0408	10	14	0.00	3.14	0.10	1.07	0.74	0.12	0.23	0.22	0.14	0.81	0.42	0.42	0.42	0.42	0.42	0.42
042003WR	040803-0420	90.000	10	0.00	0.00	0.01	0.07	0.02	0.01	0.01	0.01	0.01	0.00	0.00	0.00	0.00	0.00	0.00	0.00
051003WR	040803-0510	03	16	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
051803WR	051003-0518	03	20	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
060103WR	051003-0601	22	22	1.13	3.02	0.40	1.00	0.70	0.22	0.10	0.00	0.00	1.20	0.00	0.00	0.00	0.00	0.00	0.00
061003WR	060103-0610	23	24	0.00	2.43	0.12	0.02	0.04	0.24	0.20	0.10	0.10	2.07	0.44	0.44	0.44	0.44	0.44	0.44
061503WR	060103-0615	0.052	20	0.02	2.00	0.01	3.01	1.02	0.17	0.07	0.10	0.10	2.31	0.22	0.22	0.22	0.22	0.22	0.22
062903WR	060903-0629	0.052	30	0.02	2.00	0.01	3.01	1.02	0.17	0.07	0.10	0.10	2.31	0.22	0.22	0.22	0.22	0.22	0.22
072003WR	071003-0720	20.00	30	0.10	0.00	0.00	0.00	0.20	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
081203WR	071003-0812	07.10	32	0.00	0.00	0.04	0.37	0.00	0.01	0.01	0.01	0.01	0.74	0.00	0.00	0.00	0.00	0.00	0.00
090903WR	081003-0909	1.03	34	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00	0.00
092003WR	090003-0920	20.70	36	2.10	0.00	0.00	3.10	1.30	0.17	0.00	0.00	0.00	103.00	0.00	0.00	0.00	0.00	0.00	0.00
													4.20	0.11	0.11	0.11	0.33	0.33	0.33

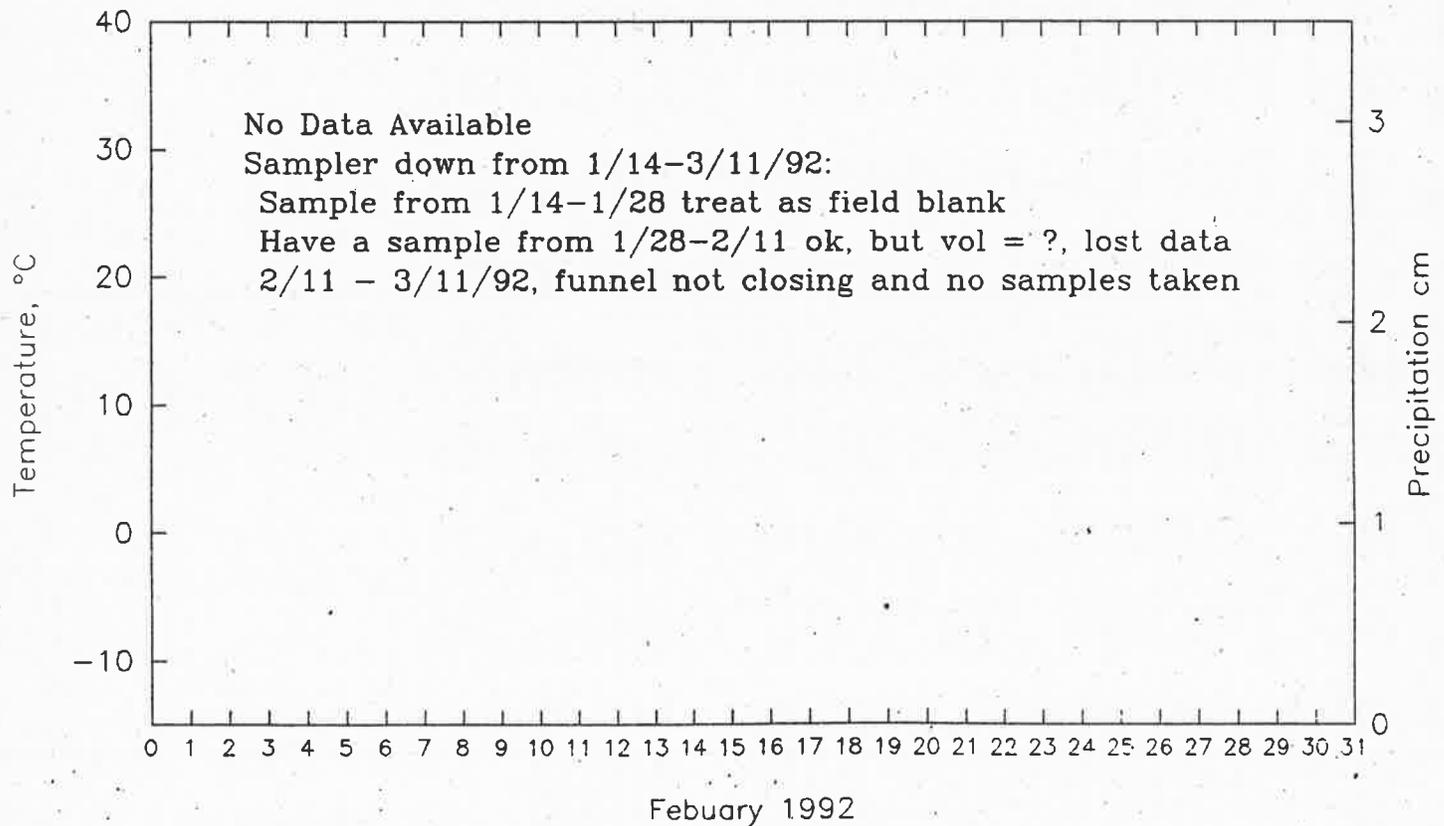
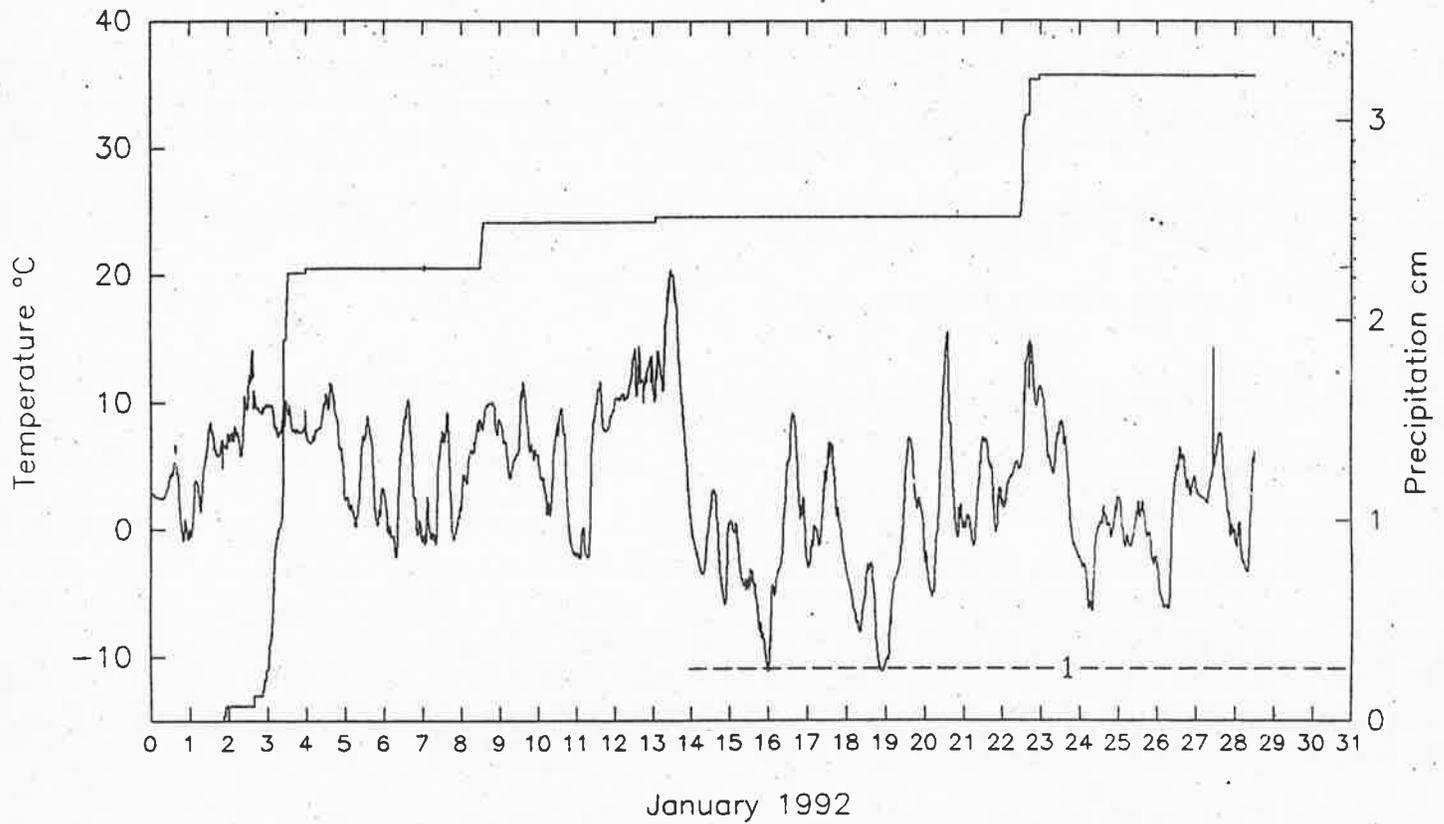
Sample	Date	Wind	Flu	Phen	Ant	Flu	Pyrene	Fluor	Chrysene	BaP	BaA	Fluor								
01275WF	12/29/92-01/12	35.652	3																	
02085WF	01/15/93-01/27	31.78	4	6.39	0.53	0.48	4.88	3.85	3.45	2.10	1.35	3.65	3.13	1.48	2.88	0.21	2.88	0.21	2.88	
02235WF	01/27/93-02/09	1.024	6	2.36	0.37	3.19	2.44	1.79	1.89	0.85	1.14	1.79	1.65	0.88	1.14	0.88	1.14	0.88	1.14	
02019WF	02/09/93-02/23	1.024	8	31.16	1.72	24.84	28.39	21.52	29.71	11.98	7.91	11.52	12.71	6.37	17.54	2.88	17.54	2.88	17.54	
02019WF	02/23/93-03/09	1.024	12	17.18	0.86	2.58	13.11	11.82	11.26	7.88	4.28	11.82	7.88	6.38	7.88	1.32	7.88	1.32	7.88	
04685WF	03/23/93-04/09	1.024	14	1.72	0.86	1.58	0.86	1.28	0.86	0.86	0.86	0.86	0.86	0.86	0.86	0.86	0.86	0.86	0.86	
04205WF	04/09/93-04/28	0.098	10	0.1553	0.0033	0.1757	0.1437	0.1234	0.1918	0.0097	0.0388	0.1234	0.0732	0.0453	0.0025	0.0097	0.0025	0.0097	0.0025	
05183WF	05/04/93-05/18	0	19																	
05183WF	05/18/93-06/01	62	29	1.81	0.14	1.88	1.82	1.28	1.71	0.78	0.38	1.28	0.74	0.71	0.88	0.19	0.88	0.19	0.88	
04133WF	06/01/93-06/15	22	22	2.89	0.24	3.08	3.87	2.83	2.71	1.48	1.84	2.83	1.81	1.28	1.84	0.88	1.84	0.88	1.84	
04293WF	06/15/93-07/15	0.621	26	5.28	0.36	7.27	0.84	19.35	4.88	3.38	1.01	19.35	3.37	4.18	0.73	3.38	0.73	3.38	0.73	
07175WF	07/15/93-07/18	0.18	28	0.83	0.04	0.88	0.79	0.88	0.84	0.23	0.18	0.88	0.28	0.28	0.21	0.88	0.21	0.88	0.21	
08125WF	07/18/93-08/12	87.15	32	0.8588	0.0188	0.8785	0.8871	0.8421	0.8488	0.84	0.83	0.84	0.84	0.84	0.85	0.84	0.85	0.84	0.85	
08013WF	08/12/93-08/28	0	34	15.88	2.48	18.28	15.14	18.22	11.88	7.88	0.18	18.22	7.88	8.78	7.88	1.28	7.88	1.28	7.88	
08063WF	08/28/93-09/28	2.19	38	2.18	0.28	3.88	2.78	2.84	2.25	1.28	0.83	2.84	1.85	1.18	1.73	1.28	1.73	1.28	1.73	

Sample	Date	Vol. (L)	Flu	Phen	Acd	Flu	Pyrene	Chrysene	BbF	BkF	BaP	Di(1,2,3-cd)P	BghiP
704071R	13/20/92-01/12	85.0	1.827	7.698	0.283	3.381	2.258	1.065	0.898	0.875	1.087	2.228	0.320
704072R	01/12/93-01/27	17.10	2.010	0.342	0.278	3.434	2.460	1.106	0.831	0.890	3.106	2.972	2.658
704074R	02/09/93-02/23	58.042	1.104	5.397	0.088	3.509	2.465	1.447	1.406	0.572	1.044	0.760	0.230
704075R	02/23/93-03/08	84.262	0.227	NO	0.032	1.398	0.559	0.197	0.155	0.120	0.311	0.200	0.154
704077R	03/09/93-03/23	40.12	0.821	2.814	0.082	1.138	1.004	0.488	0.473	0.259	1.008	0.455	0.241
704078R	04/05/93-04/26	20.5	0.316	2.011	0.082	0.938	2.209	0.351	0.213	0.173	4.360	0.238	0.250
704079R	04/20/93-04/24	164.8	0.507	1.812	0.102	1.051	0.369	0.878	0.213	0.107	1.807	0.160	0.175
704081R	05/04/93-05/18	21.8	0.084	NO	0.050	0.511	0.380	0.158	0.858	0.043	0.721	0.034	0.094
704082R	05/18/93-06/01	11	3.958	10.741	0.588	0.887	3.019	0.940	0.928	NO	NO	0.848	0.002
704083R	06/01/93-06/15	7.85	0.973	NO	0.152	3.415	1.018	1.287	0.451	1.190	7.382	1.338	0.074
704084R	06/21/93-06/29	11.2	0.193	NO	0.344	1.132	0.344	0.832	0.627	NO	0.787	0.088	1.205
704085R	07/14/93-07/14	19	0.998	3.858	0.344	0.000	0.680	0.208	0.119	NO	NO	0.000	0.125
704086R	07/14/93-07/28	12.9	0.236	NO	0.260	0.000	0.830	0.333	0.181	0.348	4.881	0.000	0.644
704088R	08/12/93-08/12	82.85	0.208	0.810	0.002	0.428	0.372	0.165	0.110	0.513	3.241	0.218	0.251
704089R	08/12/93-08/26	13.51	0.410	NO	0.000	0.433	0.285	0.187	0.134	0.125	3.021	0.169	0.111
704090R	08/28/93-09/06	7.223	0.848	0.212	0.000	0.948	2.337	0.301	0.177	0.148	0.508	0.215	0.252
704091R	09/06/93-09/20	18.8	0.598	0.907	0.090	0.383	0.283	0.090	0.090	0.043	0.508	0.133	0.338
704091R	09/21/93-09/24	18.8	0.244	NO	0.051	1.401	0.486	0.106	0.090	0.878	1.503	0.059	0.134

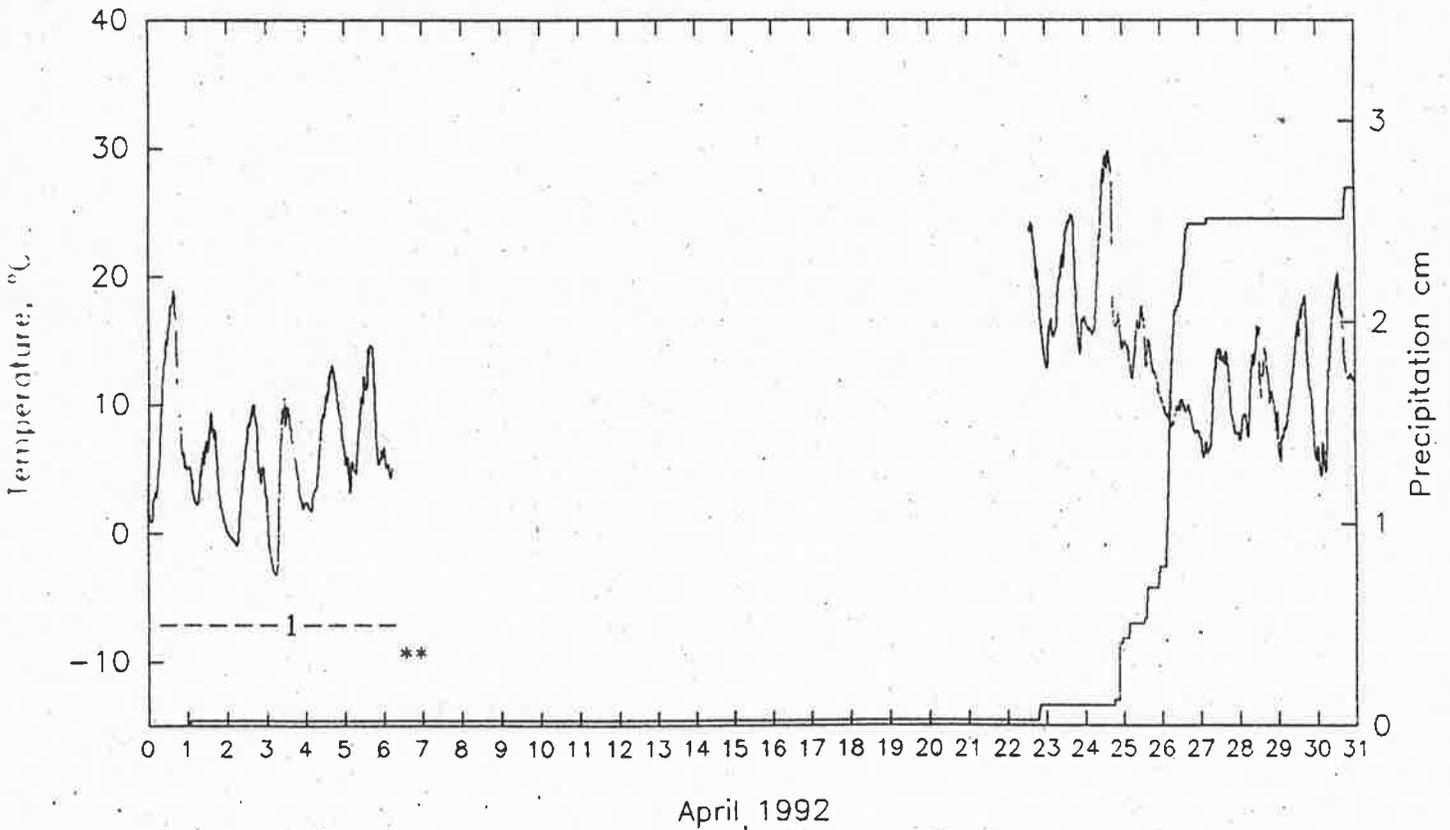
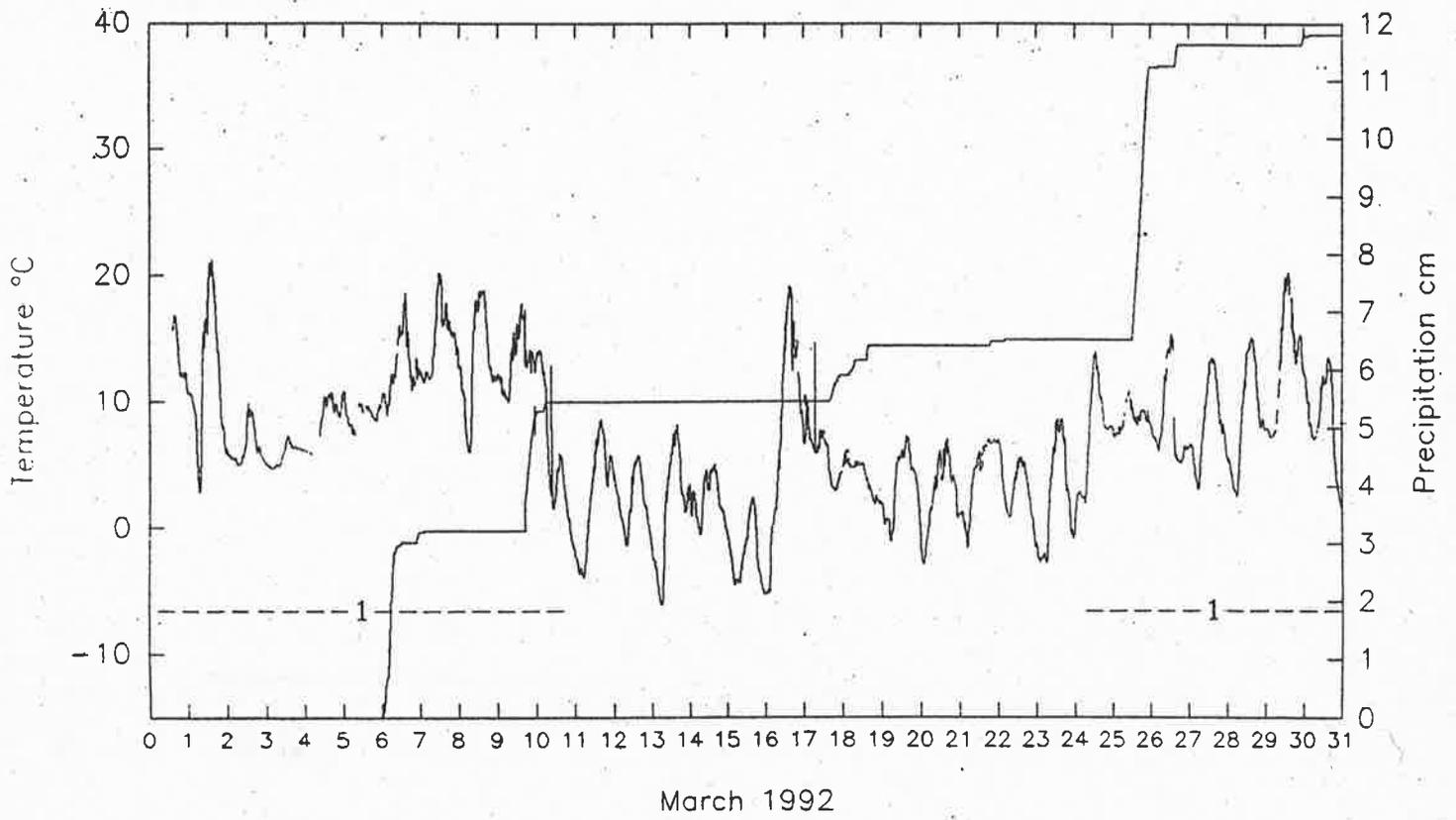
HAVEN 1992 DATA

End Date	ODU Type	grams	cm	Al	As	Cd	Cu	Fe	Pb	Mn	Ni	Se	Zn	Ca	K	Mg	Na	Cl	NO3	SO4
ID				ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug/L	ug Cl/L	ug N/L	ug S/L
10792	260	1232	1.74	2.52	0.026	0.053	0.39	1.43	0.18	0.11	0.28	0.04	2.25	0.0	173.8	473.5	4522.0	5511.41	42.56	335.62
11492	264	415	0.59	8.95	0.119	0.191	0.74	ISA	0.77	0.26	ISA	0.11	ISA	50.6	39.3	35.3	215.5	474.68	423.77	805.55
12192	NS			NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS
12892	265	1271	1.80	4.09	0.017	0.051	0.22	3.61	0.44	0.27	0.24	0.06	6.88	70.1	49.3	171.2	1302.0	2607.70	145.46	414.79
20492	300			FB	FB	FB	FB	FB	FB	FB	FB	FB	FB	FB	FB	FB	FB	36.51	0.00	0.00
21192	NS			NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	2209.95	2029.94	1429.02
21892	271	1957	2.77	5.21	0.060	0.443	0.15	3.51	0.45	0.23	0.31	0.11	8.32	36.9	25.2	46.1	319.6	621.79	282.93	636.62
22592	275	737	1.04	2.25	0.054	0.037	0.33	1.32	0.19	0.25	<0.12	0.07	1.70	25.9	12.2	20.8	136.7	196.04	105.84	406.78
30392	278	1826	2.58	2.67	0.021	0.011	0.14	1.57	0.22	0.15	0.07	0.07	1.09	18.8	10.8	17.8	90.4	252.76	126.70	300.04
31092	282	1572	2.22	63.45	0.107	0.089	0.89	48.91	1.18	4.75	ISA	0.10	ISA	419.3	96.7	201.5	1366.0	2189.75	207.47	431.78
31792	285	1132	1.60	17.83	0.069	0.102	0.28	12.01	0.46	1.03	0.57	0.11	4.83	156.1	73.2	62.6	474.5	961.05	166.88	522.50
32492	NS			NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	2039.44	1057.81	1594.43
33192	290	2920	4.13	5.66	0.027	0.082	0.95	4.93	0.29	0.40	<0.12	0.06	4.85	81.4	50.8	141.1	1116.0	2004.70	87.36	225.67
40792	NS	54	0.08	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	1266.98	2117.72	2213.41
41492	294	718	1.02	14.63	0.103	0.038	0.32	11.61	0.45	1.64	0.17	0.13	10.16	101.6	130.9	147.3	1061.0	1775.34	417.89	1116.80
42192	NS	46	0.07	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	NS	2540.70	1994.94	4148.27
42892	298	3459	4.89	7.10	0.055	0.096	0.65	8.55	0.28	1.48	0.12	0.06	2.76	75.8	88.9	101.1	724.1	1296.76	293.99	611.29
50592	303	472	0.67	48.24	0.193	0.078	0.73	56.29	0.91	3.95	0.23	0.15	5.99	217.3	42.4	62.8	259.6	494.53	910.53	1018.71
51292	306	3150	4.46	19.26	0.037	0.016	0.25	17.17	0.32	1.29	0.26	0.16	3.42	277.4	258.0	688.0	6824.0	*****	199.91	1479.35
51992	310	818	1.16	10.77	0.079	0.023	0.44	10.06	0.52	1.38	<0.12	0.21	2.31	67.3	22.1	26.9	124.9	508.71	526.52	1266.18
52692	312	145	0.21	30.20	ISA	0.087	1.06	18.75	0.88	1.31	ISA	0.19	17.68	159.2	178.8	205.0	2157.0	2967.52	696.34	1970.76
60292	315	2018	2.85	7.20	0.027	0.047	0.30	3.10	0.34	0.24	<0.12	0.05	5.76	59.0	47.7	116.6	871.8	1768.25	135.80	332.41
60992	318	1097	1.55	3.71	0.039	0.058	0.36	3.68	0.41	0.40	0.20	0.05	5.34	45.0	31.6	65.5	485.9	842.29	184.93	446.53
61692	320	396	0.56	4.87	0.045	0.052	0.44	4.26	0.43	0.28	<0.12	0.09	17.07	46.2	74.4	15.8	136.6	313.38	224.55	467.04
62392	323	2150	3.04	3.93	0.085	0.124	0.83	7.90	0.49	0.68	0.12	0.11	5.07	72.4	39.0	78.3	589.0	1244.30	397.45	1141.49
63092	326	3080	4.36	10.27	0.088	0.106	0.35	14.45	0.38	1.23	0.38	0.17	4.09	90.8	82.1	40.4	257.3	610.80	403.47	1229.00
July Comp			19.27	8.26	0.028	0.128	0.47	47.40	0.14	1.33	7.55	0.10	2.31	40.0	29.8	55.7	494.6	1284	184.7	607.7
Aug Comp			25.60	8.34	0.048	0.104	0.28	17.22	0.26	0.88	2.07	0.20	4.24	51.4	31.6	49.8	411.9	1575	232.9	807.4
Sep Comp			16.63	6.29	0.018	0.063	0.11	7.80	0.14	0.38	0.64	0.08	2.66	79.0	72.0	191.5	1304.0	1029	144.8	438.9
Oct Comp			2.50	13.23	0.093	0.144	0.65	10.20	0.25	1.01	0.10	0.13	8.14	136.4	80.3	197.5	1399.0	4021	133.1	484.6
Nov Comp			14.11	2.68	0.056	0.069	0.25	3.56	0.30	0.23	0.41	0.12	4.53	33.6	31.9	61.2	491.9	2609	537.4	1081.1
Dec Comp			12.20	1.86	0.026	0.114	0.16	2.35	0.10	0.27	0.10	0.08	3.33	70.5	67.5	191.5	1437.0	2428	170.7	467.3

Elms Climatographic Data



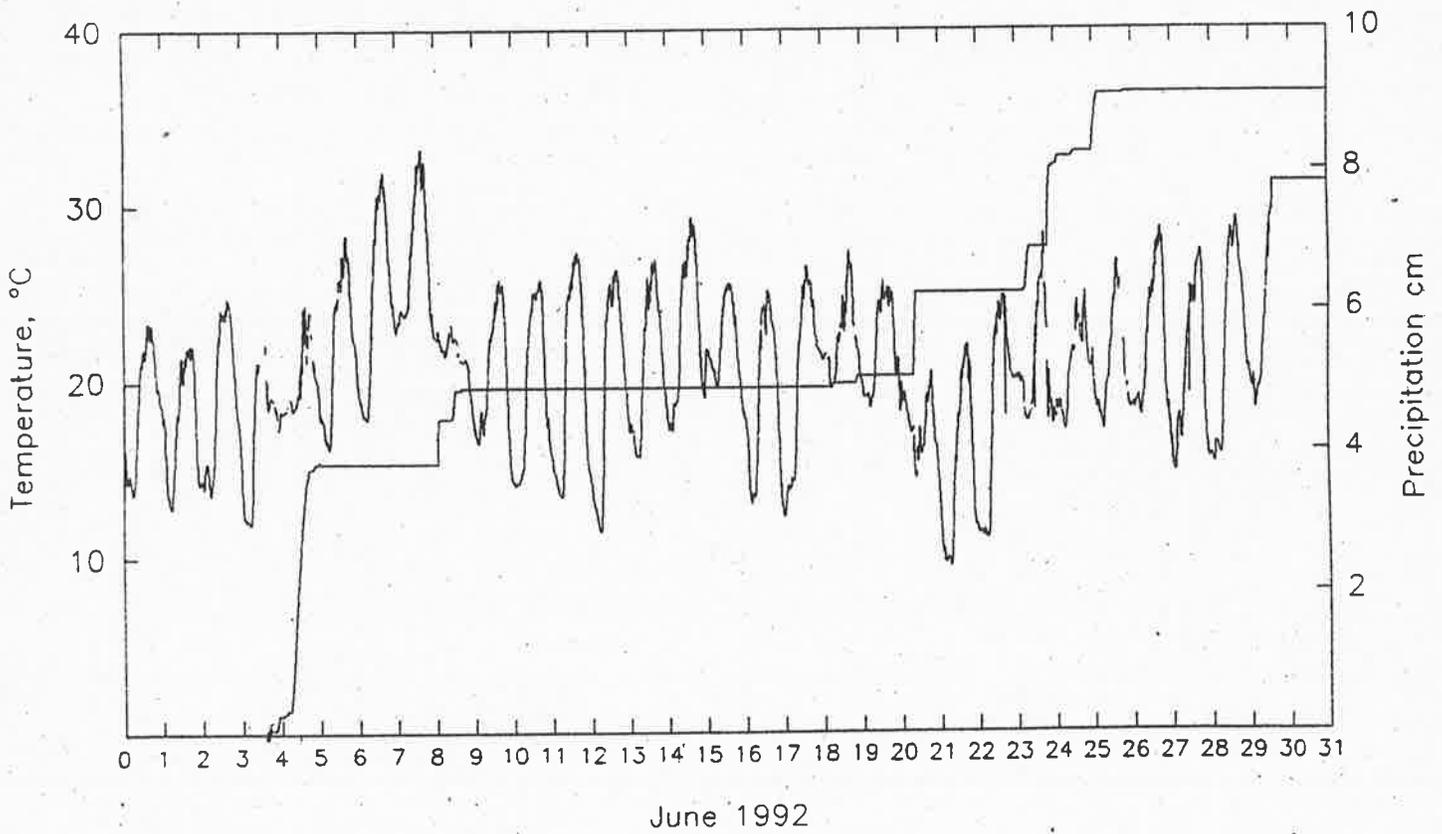
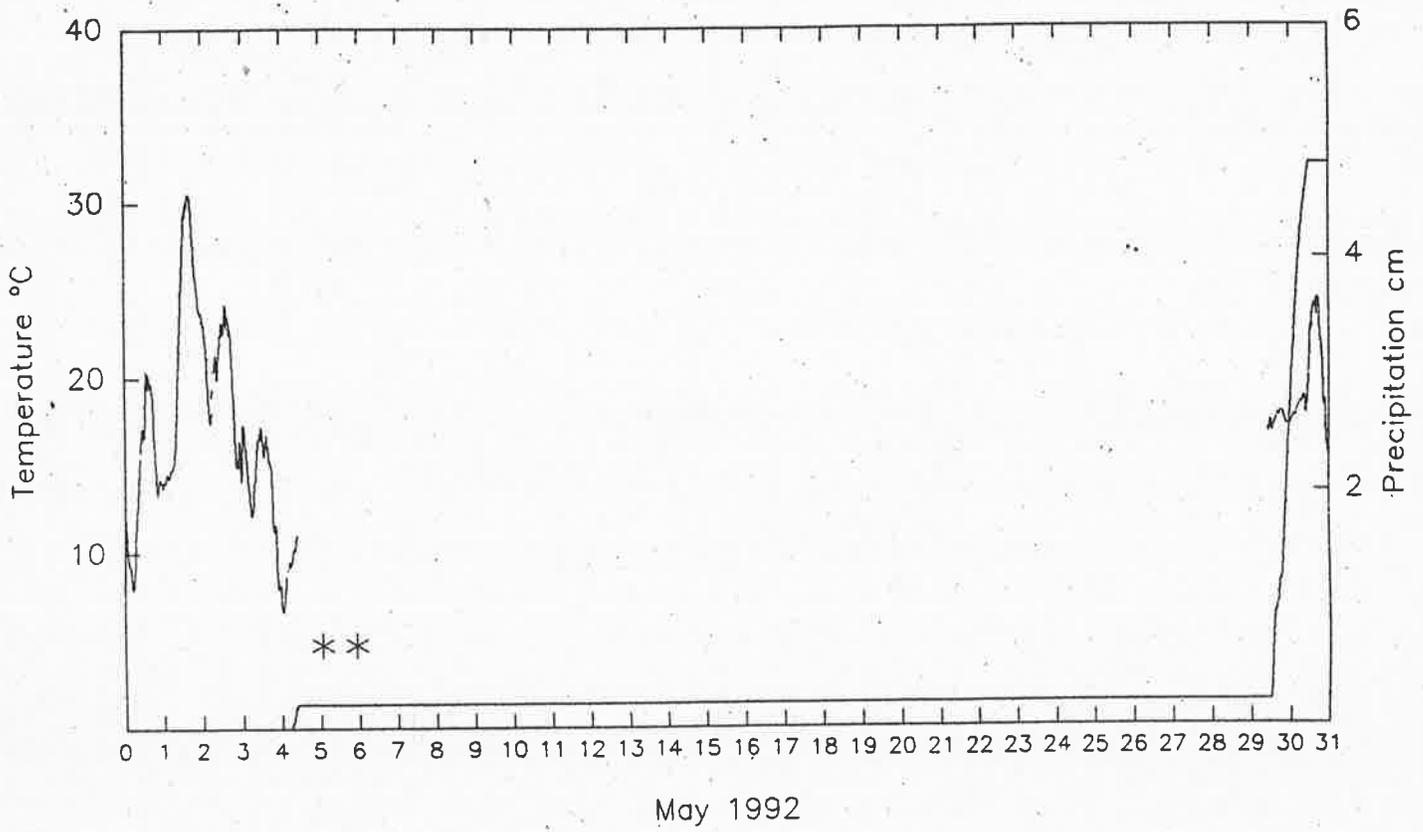
Elms Climatographic Data



1 = sampler down

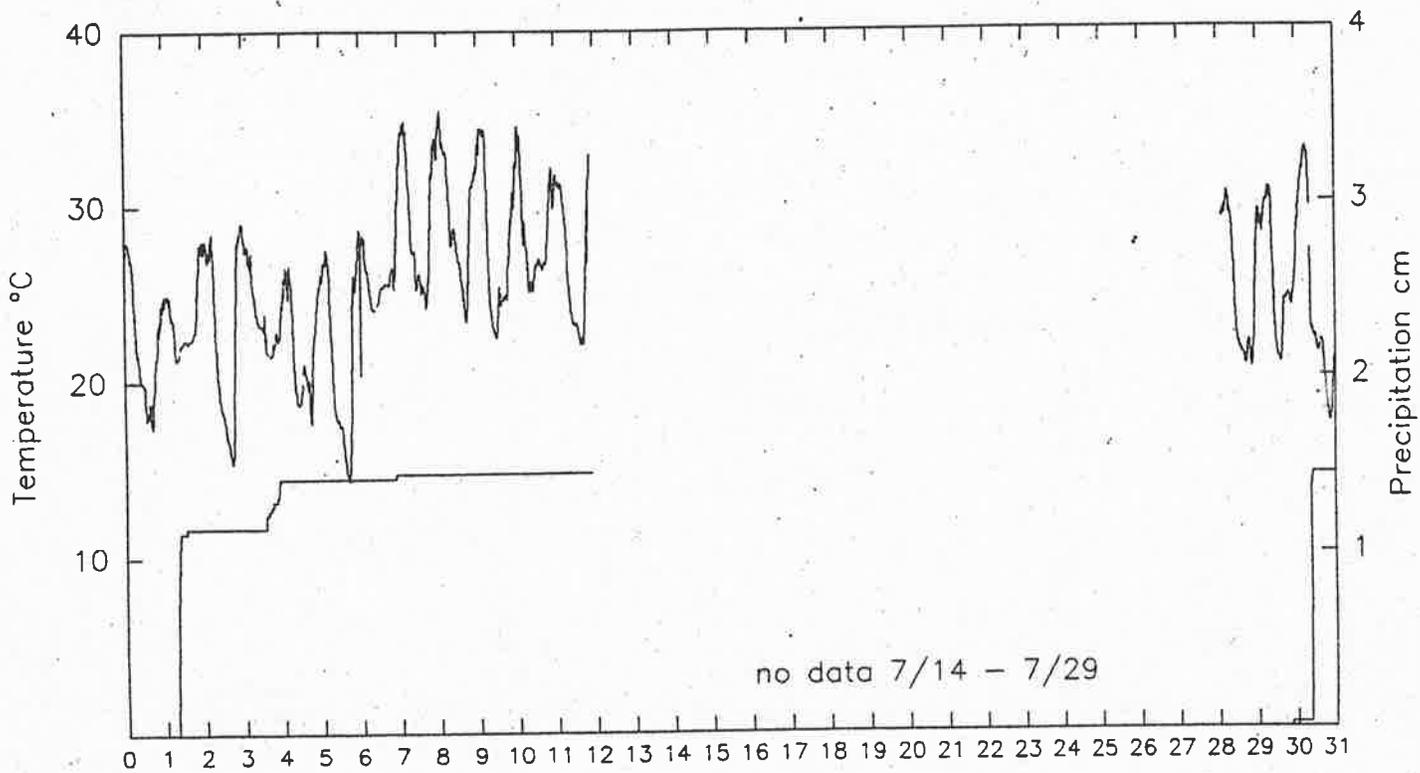
** lost data from 4/7 - 4/21

Elms Climatographic Data

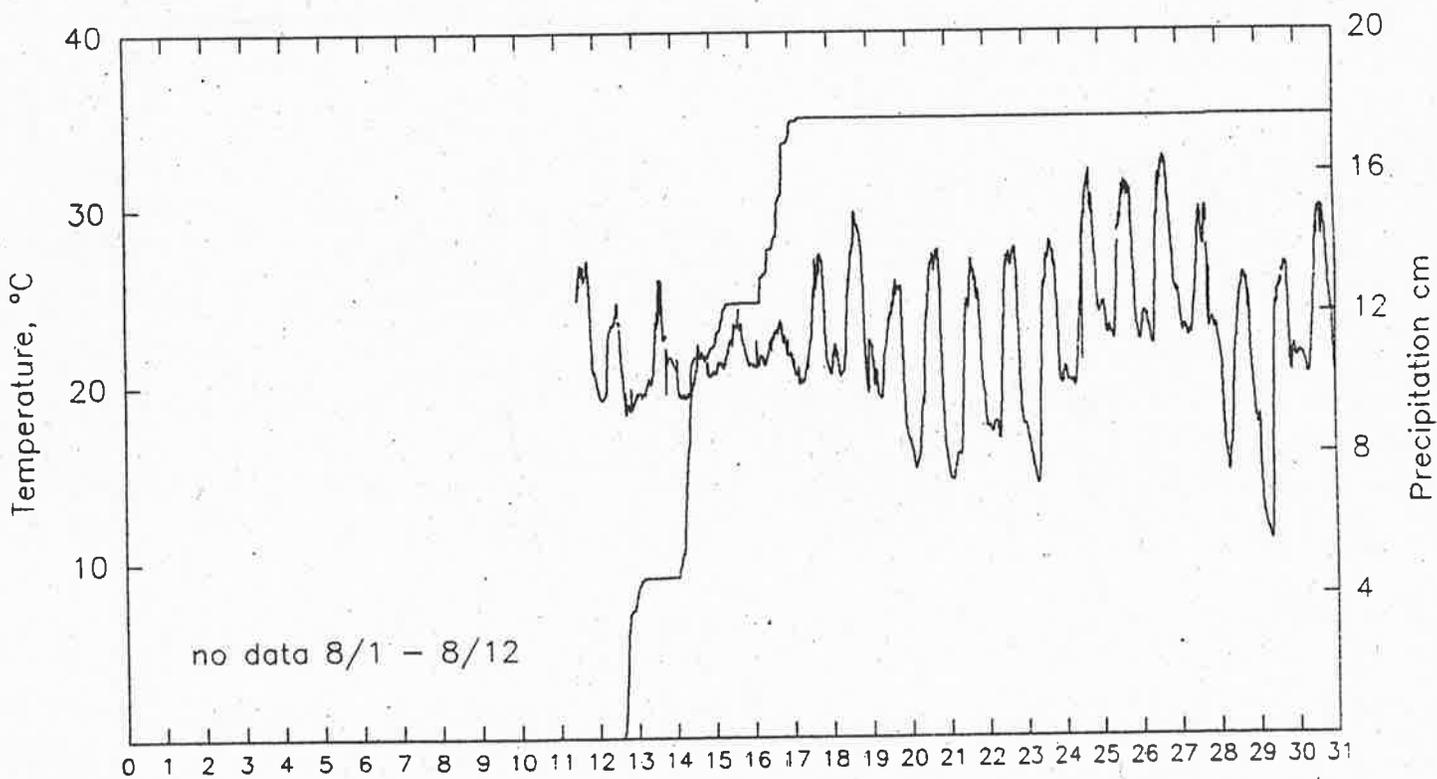


** lost data from 5/6 - 5/30

Elms Climatographic Data

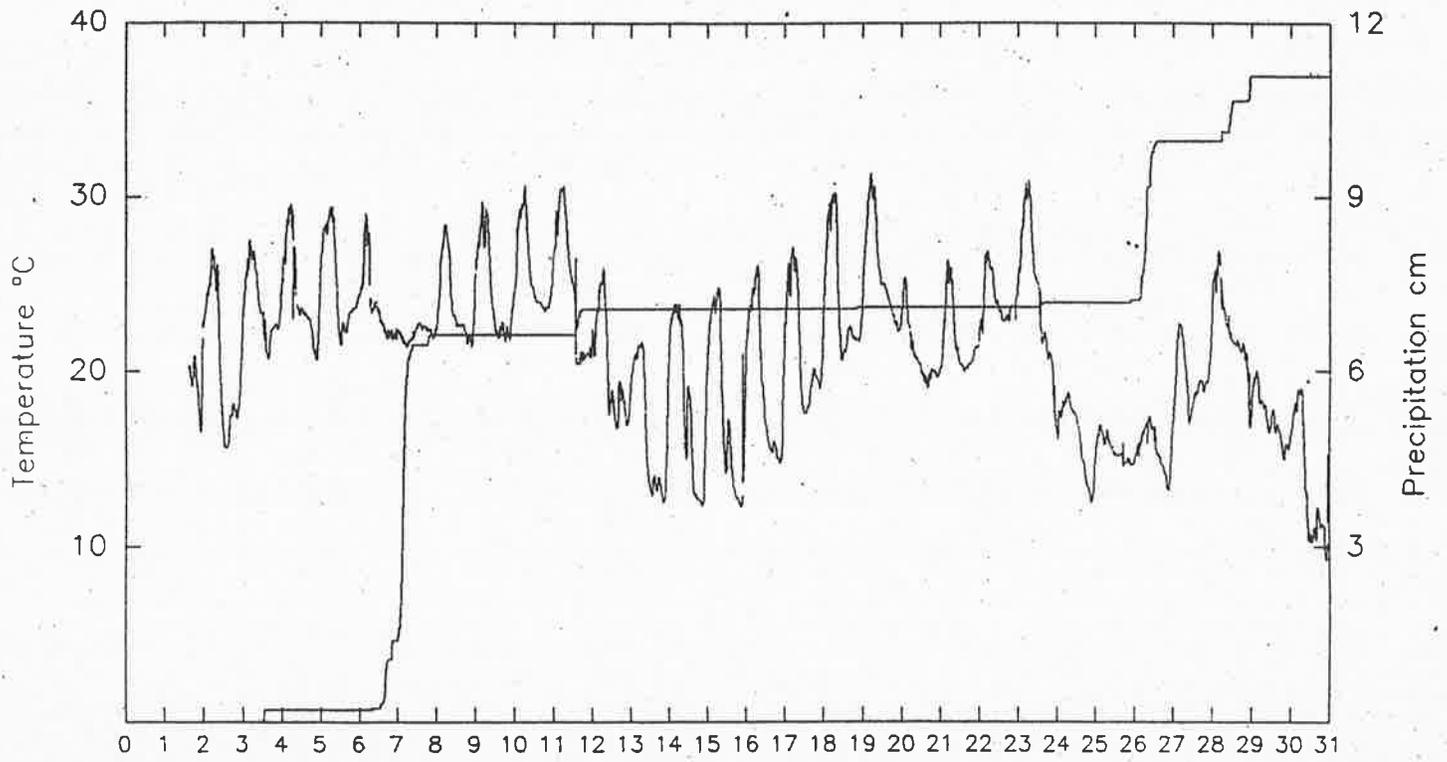


July 1992

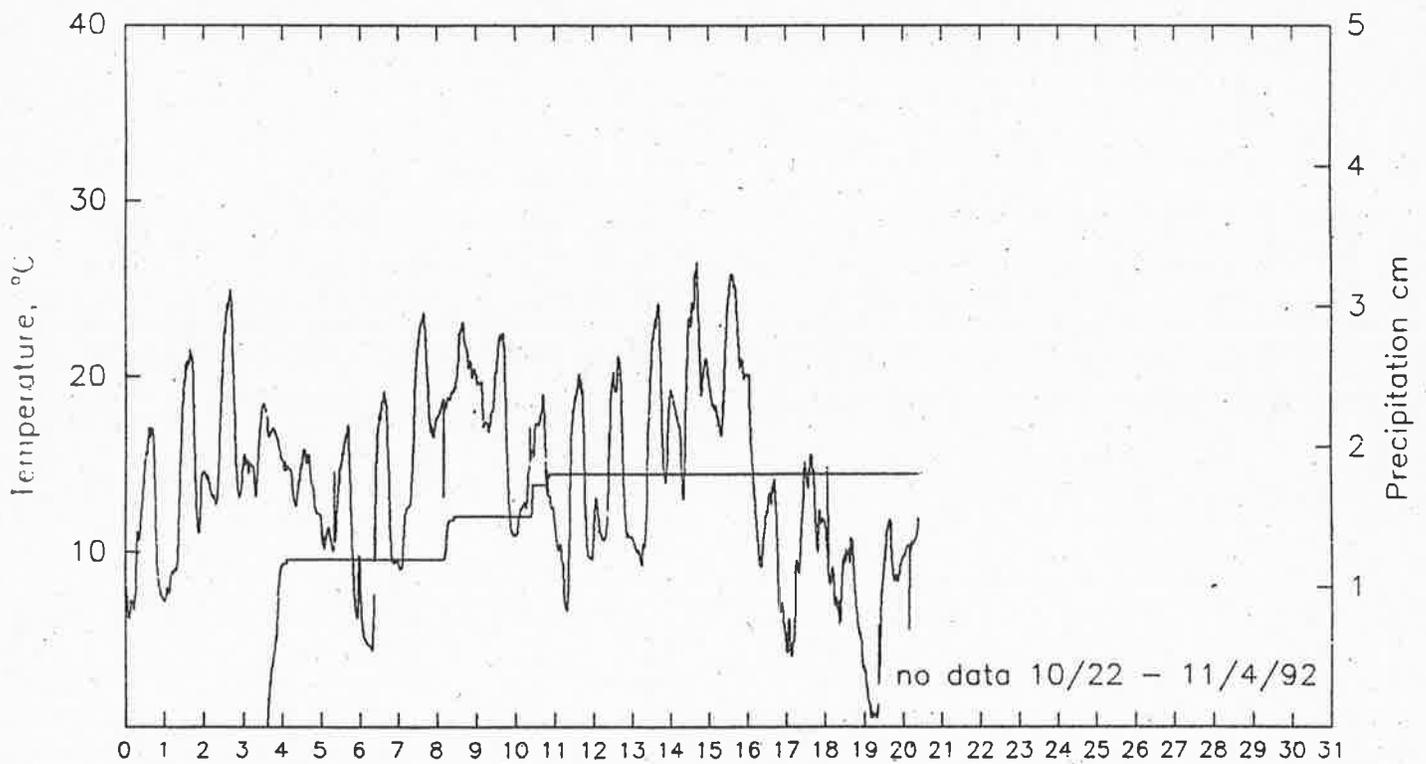


August 1992

Elms Climatographic Data

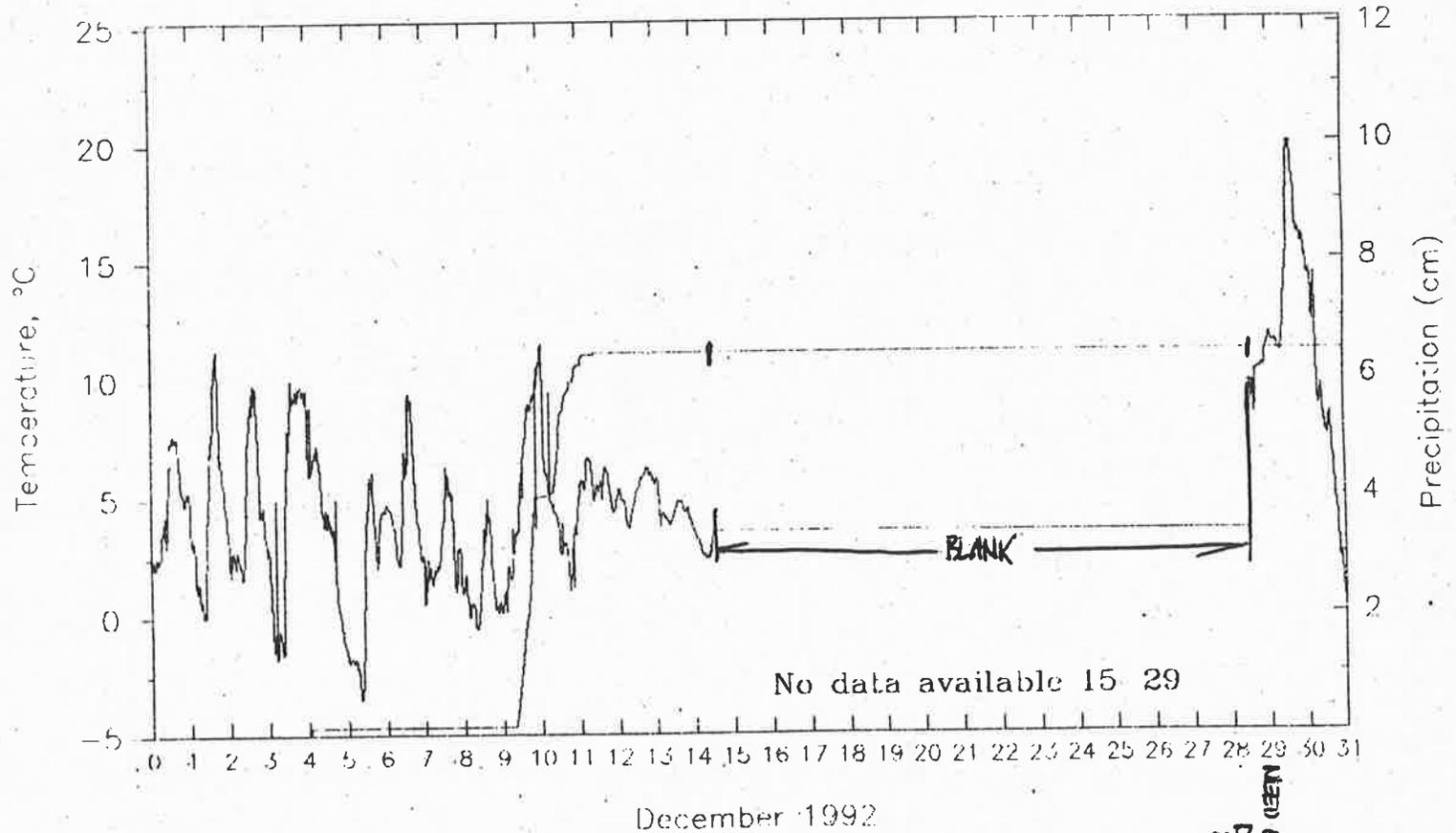
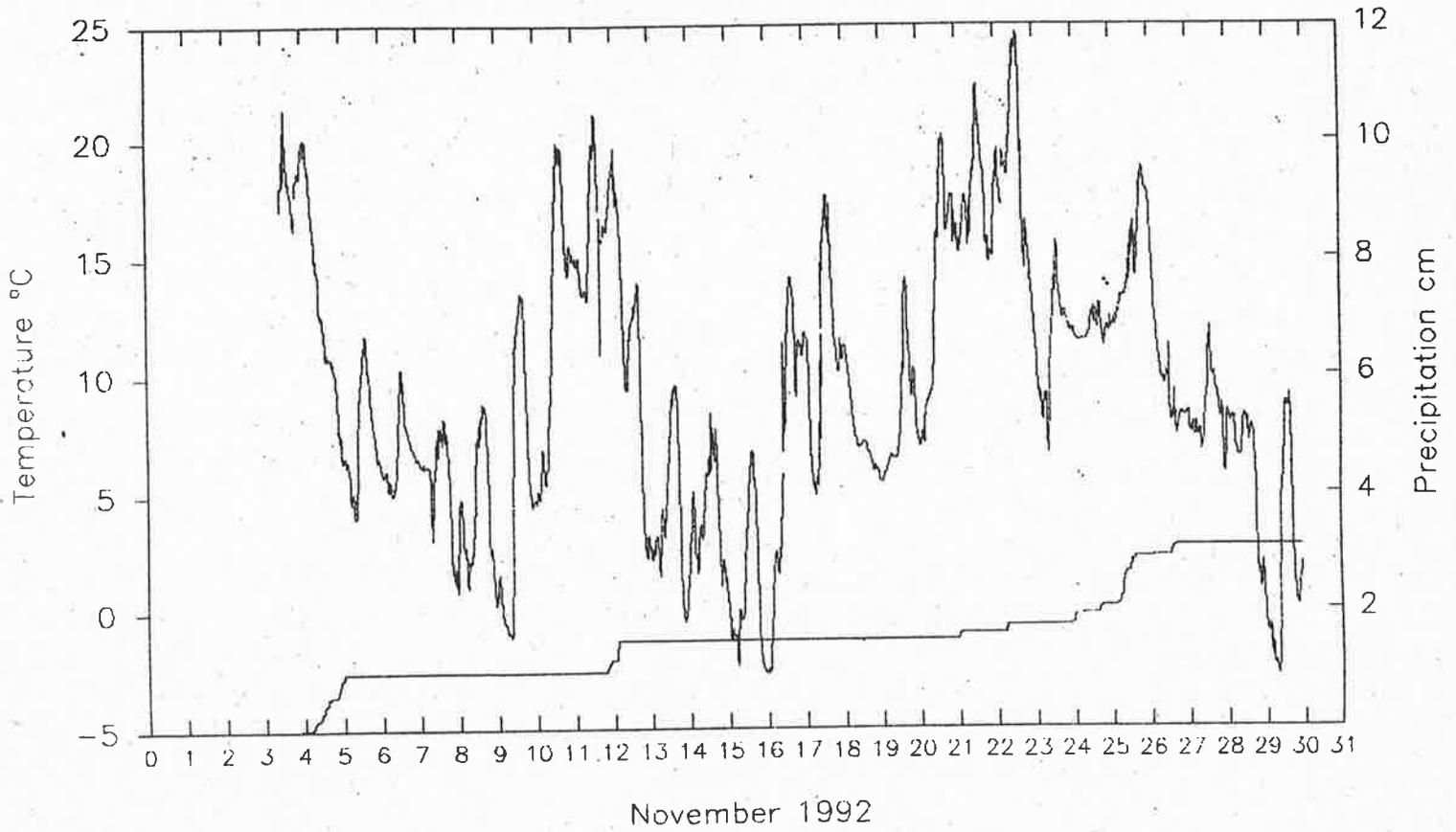


September 1992



October 1992

Elms Climatographic Data

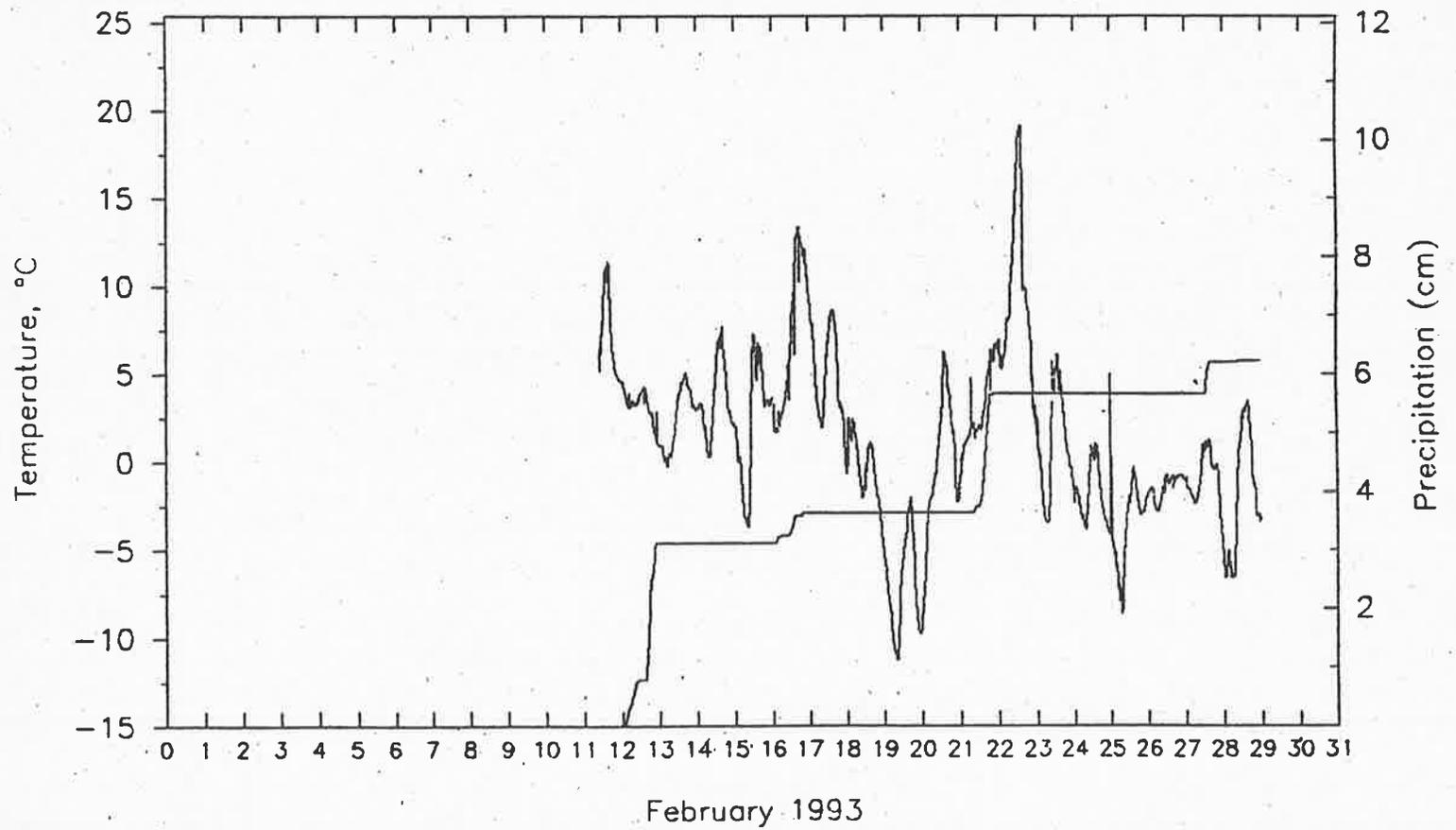
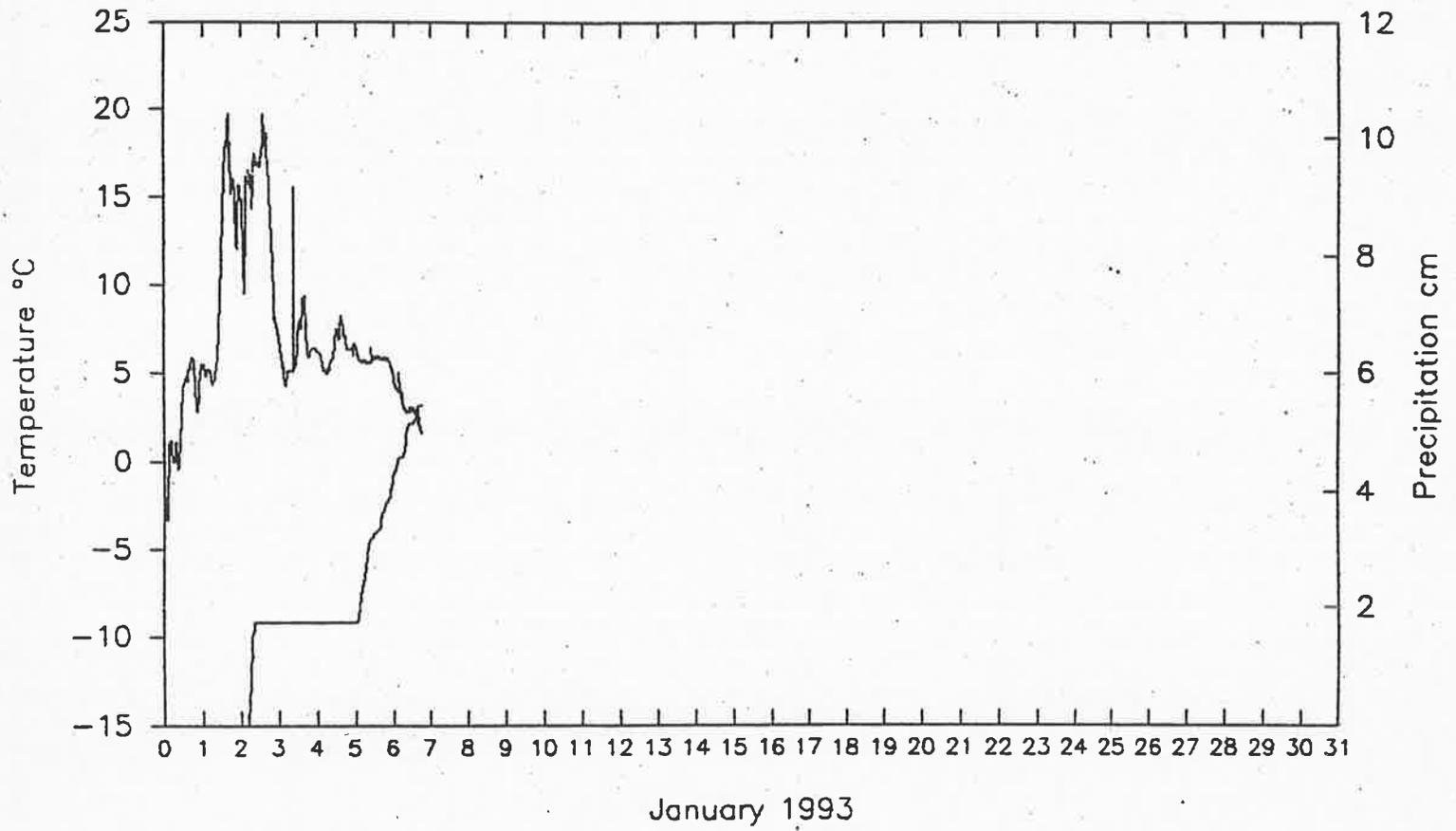


December 1992

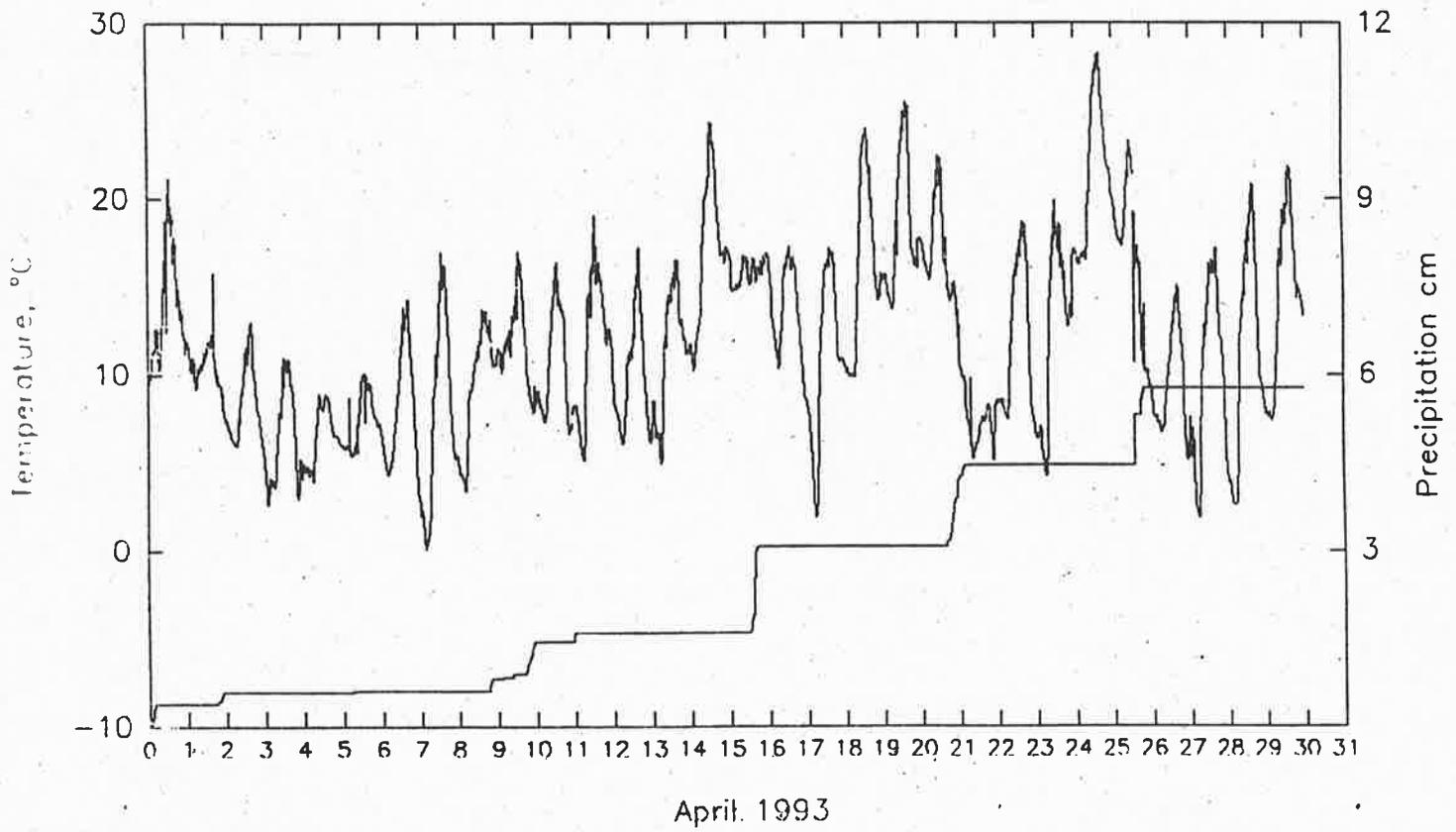
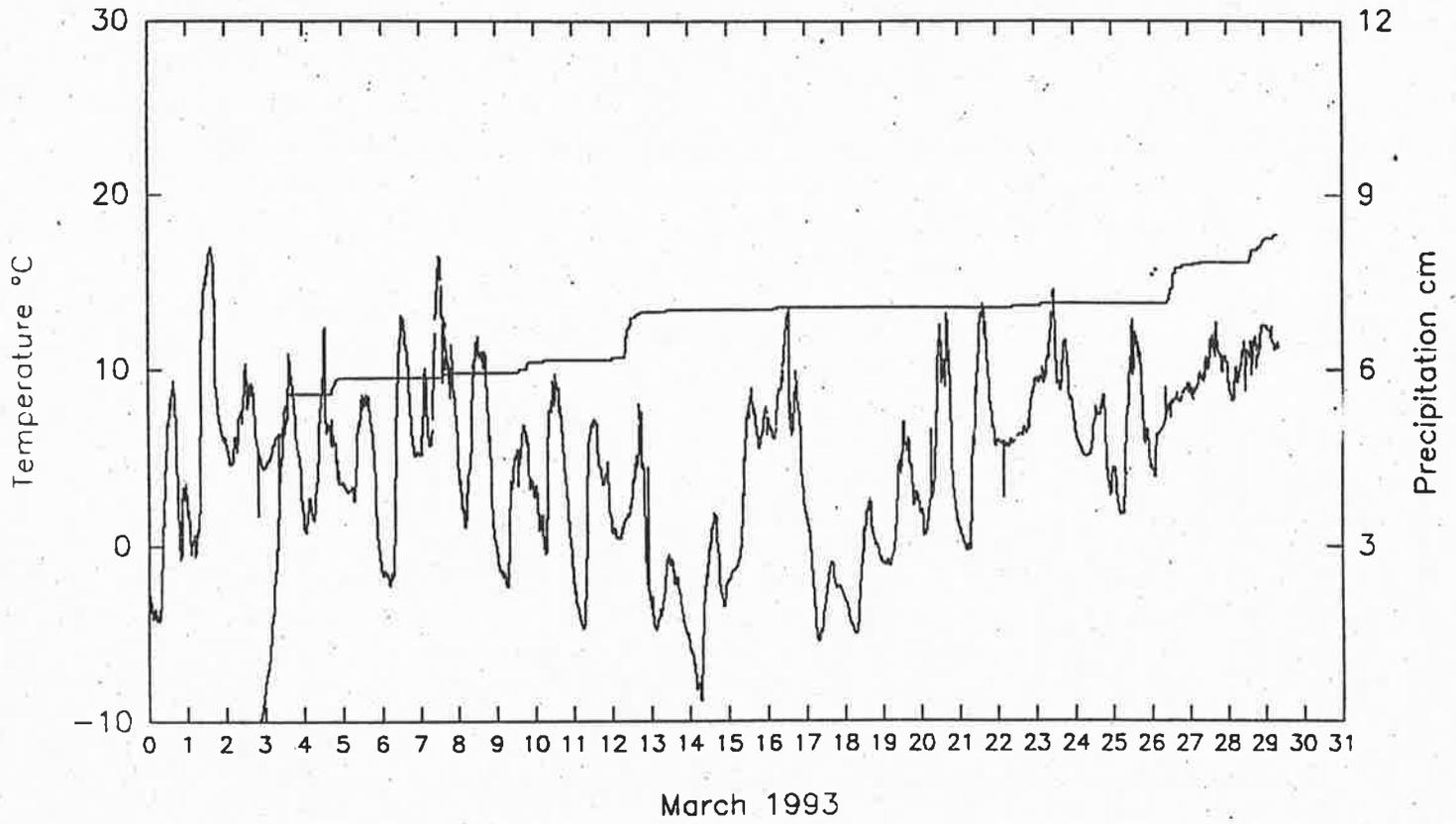
ELMS TEMP & PRECIP DATA

NEED DATA FOR
DATA FOR/UP TO
DEC '29

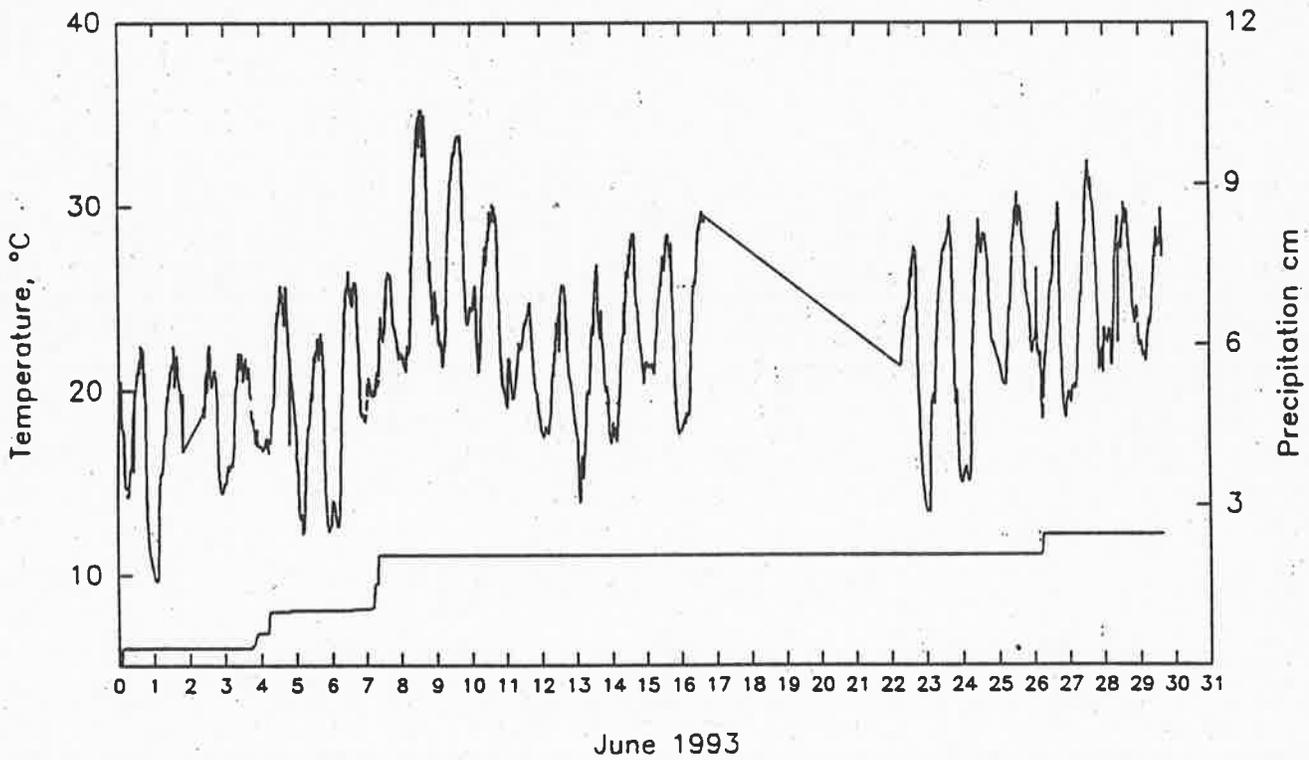
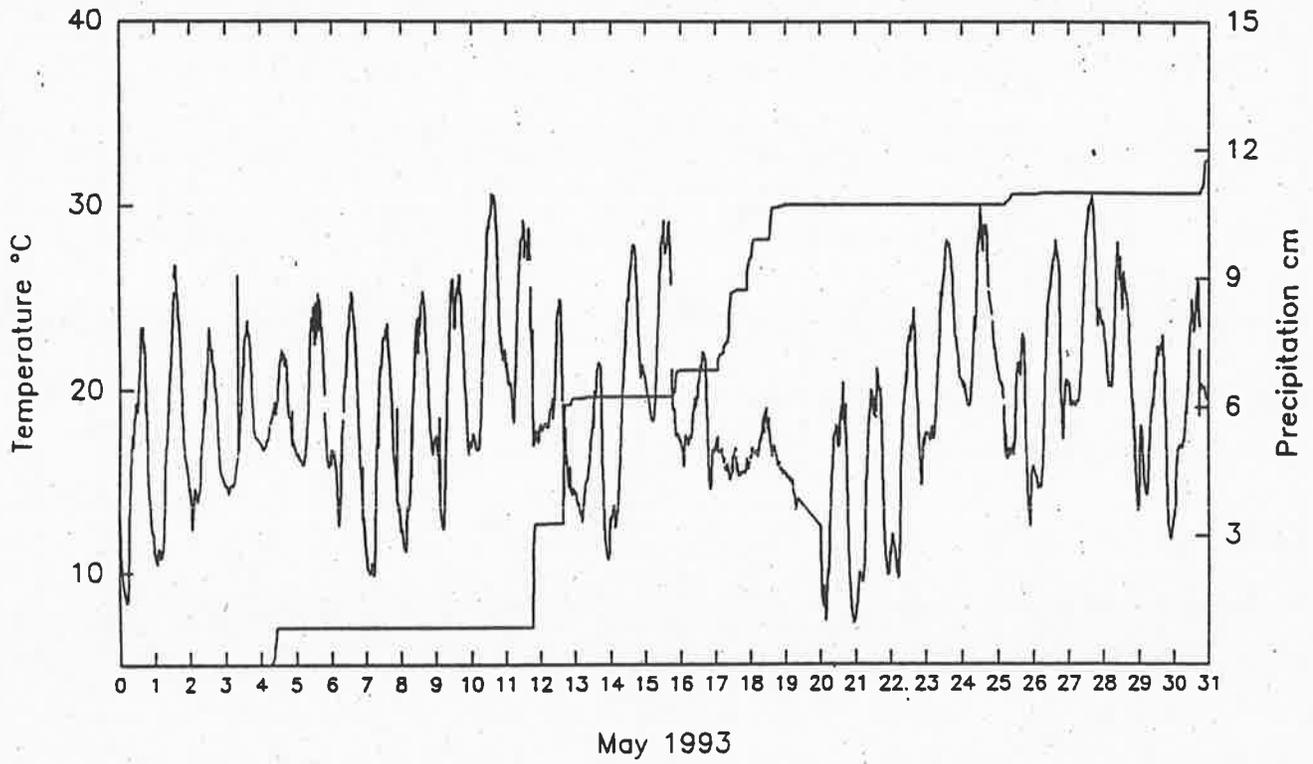
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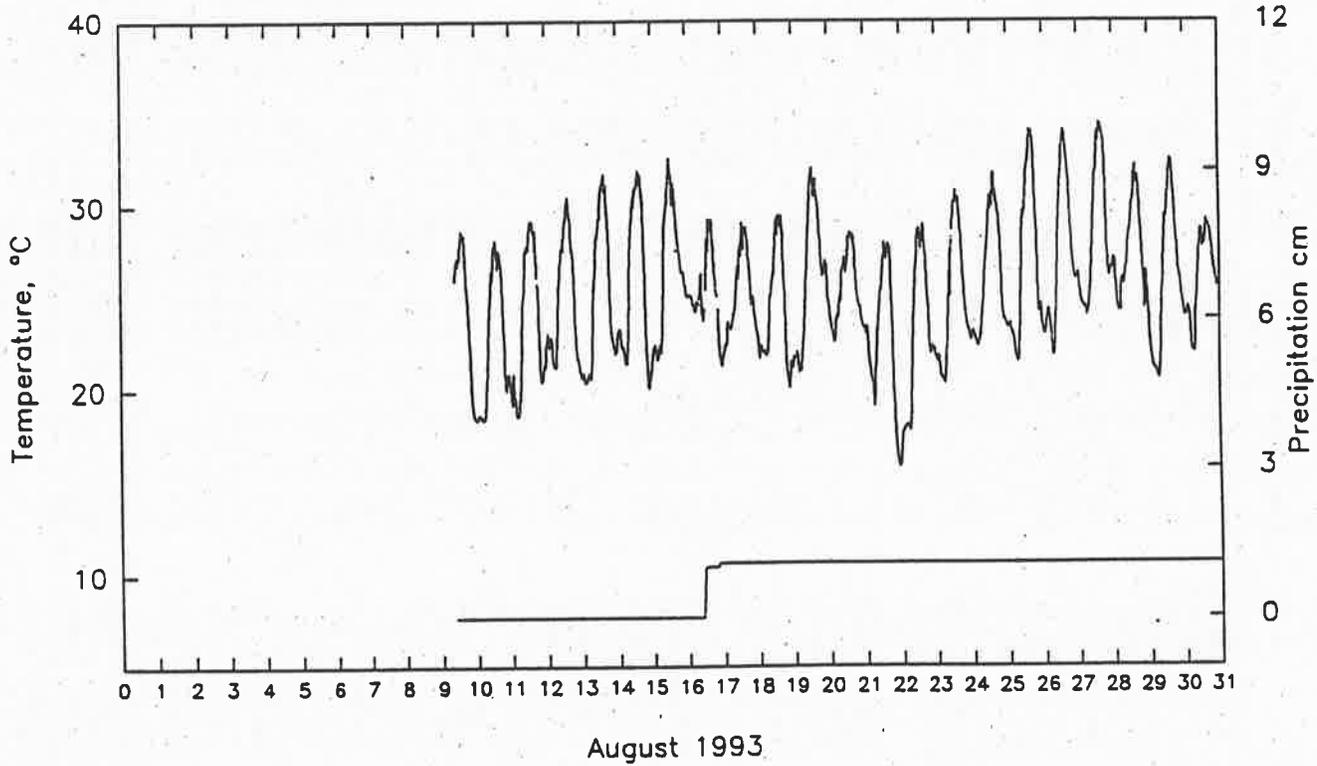
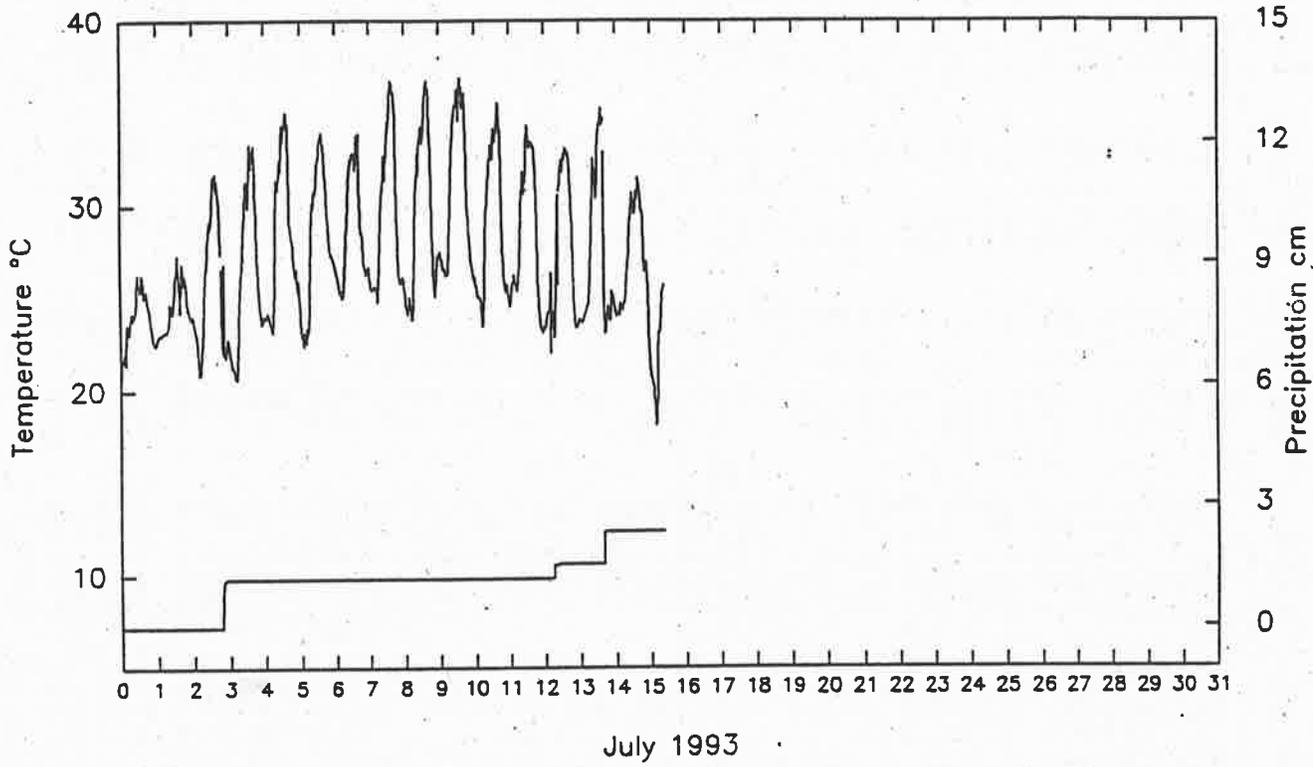
Elms Climatographic Data



Elms Climatographic Data



Elms Climatographic Data



Sample ID

Sample ID	AI ng/m ³	As ng/m ³	Cd ng/m ³	Cr ng/m ³	Cu ng/m ³	Ni ng/m ³	Pb ng/m ³	S ng/m ³	ng/m ³		
W123	180.4	56.9	0.312	0.13	2.9	0.4	0.20	1.3	1627	129	2.16
W124	165.2	50.2	0.180	0.12	1.2	0.3	0.17	0.6	772	61	1.27
W125	124.1	39.6	0.202	0.12	4.3	0.5	0.17	0.8	4.81	541	0.68
W126			0.109	0.12	1.7	0.3	0.20	0.5	1024	81	0.91
W127			0.259	0.12	2.2	0.3	0.19	0.9	1085	86	0.97
W128	160.1	50.9	0.204	0.12	1.5	0.3	0.16	0.9	1186	94	2.70
W129			0.247	0.11	0.9	0.2	0.16	0.5	781	62	0.81
W130			0.030	0.13	3.2	0.4	0.16	1.6	1741	138	3.39
W131			0.227	0.12	1.5	0.3	0.16	0.8	1297	103	1.55
W132			0.249	0.11	1.1	0.3	0.19	0.6	3233	255	1.10
W133			0.268	0.11	2.1	0.3	0.18	0.9	1017	81	4.77
W134			0.188	0.13	2	0.3	0.19	1.4	189	18	0.63
W135			0.062	0.12	1.1	0.3	0.15	0.5	736	58	1.76
W136	< 2.525	2.525	0.066	0.06	1.1	0.3	0.15	0.25	536	41	0.22
W137	55.9	13.6	0.117	0.13	2.2	0.4	0.15	0.7	478	36	1.22
W138	< 3.3	3.3	0.161	0.12	0.98	0.19	0.19	1	637	49	1.18
W139	37.6	13.9	0.447	0.10	1	0.3	0.17	0.8	565	42	1.65
W140	< 3.825	3.825	0.150	0.12	2.4	0.4	0.16	0.9	840	64	1.01
W141	17	3	0.167	0.12	< 0.25	0.25	0.17	0.7	1872	128	1.81
W142	223/93	2/24/93	0.209	0.11	1.3	0.3	0.15	0.8	1621	122	1.48
W143	302/93	3/03/93	0.167	0.12	2.5	0.5	0.18	1.1	2423	185	3.24
W144	309/93	3/10/93	0.176	0.11	1.4	0.4	0.16	0.9	1496	113	1.35
W145	316/93	3/17/93	0.055	0.11	< 0.15	0.15	0.21	0.6	957	73	0.69
W146	330/93	3/31/93	0.034	0.12	1.1	0.4	0.18	1.1	1095	86	0.98
W147	406/93	4/07/93	0.111	0.10	1.2	0.3	0.16	0.7	556	42	0.48
W148	413/93	4/15/93	0.139	0.11	1.9	0.4	0.17	1	918	70	2.30
W149	420/93	4/21/93	0.072	0.12	< 0.3	0.3	0.17	0.7	1151	88	0.72
W150	427/93	4/28/93	0.204	0.12	2.4	0.5	0.18	0.9	575	45	0.52
W151	504/93	5/05/93	0.006	0.11	1.1	0.4	0.08	0.6	999	78	0.49
W152	518/93	5/19/93	0.081	0.12	1.1	0.4	0.17	0.8	1535	120	1.04
W153	525/93	5/26/93	0.137	0.12	< 0.2	0.2	0.18	0.8	2565	200	1.36
W154	601/93	6/02/93	0.134	0.12	1.7	0.4	0.18	0.9	1627	126	2.81
W155	608/93	6/09/93	0.147	0.13	< 0.4	0.4	0.18	0.8	2708	217	2.42
W156	615/93	6/16/93	0.33	0.13	1	0.4	0.13	0.7	1667	132	1.16
W157	622/93	6/23/93	0.70	0.14	1.4	0.4	0.14	1.5	2627	210	2.30
W158	629/93	6/30/93	1.29	0.14	6	0.8	0.14	1.5	3932	311	3.54
W159	706/93	7/07/93	1.06	0.14	2.1	0.5	0.15	1.1	5369	425	2.29
W160	713/93	7/14/93	3.48	0.15	1.3	0.4	0.15	1.1	5780	466	4.24
W161	720/93	7/21/93	1.8	0.13	1.8	0.4	0.13	1.4	3552	280	2.97
W162	727/93	7/28/93	1.24	0.13	1.9	0.4	0.13	0.9	4418	351	2.60
W163	83/93	8/4/93	1.28	0.13	1.3	0.4	0.13	0.8	2937	231	2.21
W164	810/93	8/11/93	0.67	0.12	< 0.15	0.15	0.12	0.7	2396	186	1.74
W165	817/93	8/18/93	0.58	0.12	< 0.15	0.15	0.14	1	4338	337	3.45
166	824/93	8/25/93	0.99	0.14	2.6	0.5	0.14	1			
167	831/93	9/1/93									
168	97/93	9/8/93									
169	914/93	9/15/93									
170	912/93	9/22/93									
171	928/93	9/29/93									

Wye Year 1 First 12 months statistics

ave	STD	avg error	count	min	max	median	max/mean
119	0.713	0.14	0.744	3.023	2.31	4.0	2.7
59	0.342	0.08	0.549	2.830	1.80	3.3	2.7
12.2	0.071	0.02	0.232	0.308	0.27	0.5	0.2
47	48	48	48	48	48	48	48
25	0.337	0.01	0.168	0.182	0.15	0.2	0.6
303	1.963	0.43	2.279	11.391	9.24	13.6	3.5
107	0.611	0.12	0.595	1.939	1.74	4.1	1.4
2.5	2.8	3.2	3.1	3.8	4.0	4.1	2.3

Sample ID	Al		As		Cd		Cr		Cu		Ni		Pb		S	
	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma	ng/m ³	Sigma
1st 18 mos	First 18 months statistics															
ave	114	0.643	0.14	0.718	2.09	3.001	3.9	2586	1.6							
STD	59	0.328	0.07	0.485	1.54	2.491	2.3	1637	0.6							
avg error	10	0.050	0.02	0.171	0.24	0.290	0.5	379	0.1							
count	75	76	75	75	73	76	76	70	76							
min	25	0.105	0.01	0.020	0.15	0.080	0.1	14	0.4							
max	303	1.963	0.43	2.279	9.24	11.391	13.6	7386	3.5							
median	95	0.558	0.13	0.639	1.67	2.285	3.9	2299	1.5							
max/mean	2.7	3.1	3.1	3.2	4.4	3.8	3.5	2.9	2.2							
Year 2	Year 2 statistics															
ave	102	0.616	0.15	0.751	1.62	3.638	4.4	2147	1.8							
STD	63	0.308	0.05	0.373	0.69	2.106	2.3	1485	0.6							
avg error	6	0.013	0.01	0.061	0.19	0.216	0.5	232	0.1							
count	51	51	50	47	48	48	51	51	51							
min	33	0.105	0.05	0.020	0.35	0.080	0.1	14	0.4							
max	320	1.558	0.29	1.828	3.75	8.685	13.0	7386	4.2							
median	84	0.527	0.15	0.706	1.50	3.052	4.3	1699	1.8							
max/mean	3.1	2.5	1.9	2.4	2.3	2.4	3.0	3.4	2.3							
Year 3	Year 3 statistics															
ave	82	0.644	0.21	0.898	1.68	3.137	5.7	1828	1.7							
STD	62	0.287	0.17	0.734	0.98	2.834	2.9	1301	1.1							
avg error	18	0.026	0.01	0.121	0.34	0.196	0.8	144	0.1							
count	39	61	51	56	56	48	59	61	60							
min	3	0.095	0.01	0.065	0.15	0.075	0.3	54	0.0							
max	260	1.411	1.11	3.822	6.00	13.003	14.4	5780	4.8							
median	70	0.584	0.17	0.681	1.60	2.348	5.2	1535	1.6							
max/mean	3.2	2.2	5.4	4.3	3.6	4.1	2.5	3.2	2.8							

* Sampler malfunction, either no sample was collected or the air volume was inaccurate, values (when reported) are ng/filter, but not included in averages or flux estimates.

* Value not included in average.

ND = not detected.

Notes: December, 1994, pic

Analyze data for final submittal to J. Baker - Detection limits handled as follows:

If value/sigma <= 2, then value taken as value/2, and sigma = value/2

If value/sigma > 2, then left as is

If value > 6 times avg, then remove from average - place in parentheses

Se	Zn	V	Br	Fe	Mn
ng/m ³					
Sigma	Sigma	Sigma	Sigma	Sigma	Sigma
± 0.091	5.57 ± 0.76	0.564 ± 0.044	1.43 ± 0.13	91 ± 4.6	3.44 ± 0.03
± 0.10	17.5 ± 2.4	3.80 ± 0.23	2.68 ± 0.28	127 ± 12.4	1.53 ± 0.08
± 0.11	5.37 ± 0.79	1.74 ± 0.14	1.89 ± 0.22	91 ± 7.8	2.90 ± 0.15
± 0.19	10.0 ± 1.4	3.36 ± 0.26	3.03 ± 0.24	115 ± 6.7	4.09 ± 0.20
± 0.18	9.68 ± 1.4	4.51 ± 0.35	2.77 ± 0.29	167 ± 9.7	4.71 ± 0.24
± 0.20	7.10 ± 1.0	4.09 ± 0.35	1.54 ± 0.16	116 ± 7.3	3.25 ± 0.16
± 0.21	7.68 ± 1.1	5.08 ± 0.43	2.54 ± 0.23	129 ± 8.1	1.84 ± 0.09
± 0.11	48.2 ± 6.7	1.81 ± 0.14	1.84 ± 0.18	283 ± 16.0	3.55 ± 0.18
± 0.22	9.73 ± 1.2	(26.5 ± 2.25)	3.00 ± 0.29	162 ± 7.4	4.05 ± 0.20
± 0.12	9.97 ± 0.99	3.12 ± 0.24	1.89 ± 0.17	85 ± 3.1	2.10 ± 0.02
± 0.66)	(37.9 ± 4.6)	(14.0 ± 1.1)	ND	(292 ± 15)	<
± 0.23	13.0 ± 1.6	3.22 ± 0.25	1.11 ± 0.20	149 ± 7.6	4.24 ± 0.21
± 0.30	12.2 ± 1.5	4.43 ± 0.35	4.38 ± 0.37	183 ± 9.1	6.69 ± 0.33
± 0.18	10.7 ± 1.3	2.72 ± 0.21	2.09 ± 0.14	96 ± 11.2	3.15 ± 0.16
± 0.17)	(92 ± 114)	(1.3 ± 0.61)	(1.69 ± 0.13)	(10100 ± 660)	(2.94 ± 0.16)
± 0.38	20.0 ± 2.5	3.47 ± 0.22	4.97 ± 0.34	160 ± 9.5	5.81 ± 0.29
± 0.17	6.79 ± 0.89	1.51 ± 0.10	2.48 ± 0.18	109 ± 8.1	3.67 ± 0.18
± 0.082	5.94 ± 0.75	1.95 ± 0.15	6.85 ± 0.57	101 ± 5.1	3.21 ± 0.18
± 0.056	8.88 ± 1.1	1.97 ± 0.15	2.45 ± 0.19	70 ± 4.4	3.15 ± 0.16
± 0.078	18.4 ± 2.2	1.59 ± 0.10	1.90 ± 0.16	100 ± 4.7	4.44 ± 0.22
± 0.22	23.9 ± 2.9	3.49 ± 0.27	5.36 ± 0.36	277 ± 12.5	8.20 ± 0.41
± 0.11	9.75 ± 1.2	1.61 ± 0.13	2.09 ± 0.18	95 ± 4.8	4.14 ± 0.21
± 0.16	15.0 ± 1.9	3.17 ± 0.25	3.23 ± 0.27	148 ± 6.2	3.84 ± 0.19
± 0.081	9.16 ± 1.1	1.43 ± 0.11	2.21 ± 0.19	63 ± 3.5	2.40 ± 0.12
± 0.089	14.9 ± 1.4	1.89 ± 0.12	1.58 ± 0.15	65 ± 2.0	1.03 ± 0.08
± 0.18	23.9 ± 2.2	3.62 ± 0.23	4.52 ± 0.38	154 ± 4.3	1.75 ± 0.13
± 0.089	10.2 ± 1.0	2.46 ± 0.16	2.67 ± 0.24	33 ± 2.7	0.81 ± 0.08
± 0.18	18.0 ± 1.7	3.82 ± 0.24	ND	89 ± 3.3	1.40 ± 0.11
± 0.34	35.7 ± 4.5	7.03 ± 0.55	0.064 ± 0.008	160 ± 7.9	<
± 0.20	16.1 ± 1.5	4.02 ± 0.34	4.36 ± 0.35	93 ± 3.8	2.86 ± 0.11
± 0.21	12.7 ± 1.3	4.40 ± 0.37	3.50 ± 0.28	73 ± 3.3	2.38 ± 0.09
± 0.24	25.1 ± 2.4	5.44 ± 0.44	3.43 ± 0.23	169 ± 6.0	4.98 ± 0.10
± 0.20	15.1 ± 1.4	3.83 ± 0.32	4.10 ± 0.33	103 ± 3.2	3.01 ± 0.12
± 0.29	27.1 ± 2.5	6.68 ± 0.56	8.43 ± 0.68	118 ± 4.0	5.32 ± 0.21
± 0.13	9.16 ± 0.9	3.26 ± 0.26	2.75 ± 0.18	50 ± 2.3	1.18 ± 0.02
± 0.14	12.3 ± 1.2	2.95 ± 0.25	2.98 ± 0.24	78 ± 3.1	2.43 ± 0.09
± 0.15	13.8 ± 1.3	2.58 ± 0.22	3.71 ± 0.30	76 ± 3.3	2.24 ± 0.09
± 0.17	14.0 ± 1.4	2.39 ± 0.19	2.53 ± 0.17	115 ± 4.4	3.06 ± 0.06
± 0.12	13.5 ± 1.3	3.87 ± 0.33	2.78 ± 0.23	72 ± 2.4	2.14 ± 0.08
± 0.11	12.6 ± 1.2	3.20 ± 0.27	3.16 ± 0.26	86 ± 3.1	2.73 ± 0.11
± 0.089	8.67 ± 0.86	3.81 ± 0.31	2.80 ± 0.19	83 ± 2.8	2.44 ± 0.05
± 0.15	12.2 ± 1.2	3.67 ± 0.31	4.06 ± 0.34	105 ± 4.0	3.05 ± 0.12
± 0.10	12.0 ± 1.2	2.72 ± 0.23	4.60 ± 0.40	138 ± 5.1	3.40 ± 0.13
± 0.10	7.21 ± 0.77	2.41 ± 0.19	2.39 ± 0.17	60 ± 2.9	1.22 ± 0.03
± 0.11	12.3 ± 1.2	3.04 ± 0.26	4.67 ± 0.41	95 ± 3.0	2.30 ± 0.09
± 0.079	6.32 ± 0.69	1.42 ± 0.12	1.68 ± 0.14	88 ± 3.3	3.26 ± 0.13
± 0.20	21.4 ± 2.0	3.39 ± 0.27	3.20 ± 0.22	177 ± 5.1	4.43 ± 0.09
± 0.095	9.20 ± 1.0	2.95 ± 0.25	2.83 ± 0.28	120 ± 4.2	2.99 ± 0.12
± 0.12	5.91 ± 0.73	3.47 ± 0.29	3.45 ± 0.31	171 ± 5.9	3.31 ± 0.13
± 0.18	13.6 ± 1.3	4.53 ± 0.36	2.60 ± 0.18	142 ± 4.4	4.53 ± 0.09
± 0.12	13.8 ± 1.3	2.73 ± 0.22	2.66 ± 0.18	194 ± 5.5	3.76 ± 0.08
± 0.14	10.9 ± 1.0	2.50 ± 0.21	2.81 ± 0.24	168 ± 4.3	3.68 ± 0.14
± 0.11	12.8 ± 1.2	4.73 ± 0.40	2.81 ± 0.24	72 ± 2.6	2.02 ± 0.08
± 0.12	7.20 ± 0.74	3.66 ± 0.30	2.99 ± 0.21	92 ± 3.0	2.52 ± 0.06
± 0.13	4.70 ± 0.66	3.55 ± 0.30	2.13 ± 0.19	69 ± 3.0	1.74 ± 0.07

(Wye Site)

Se	Zn	V	Br	Fe	Mn
Sigma	ng/m ³				
Sigma	Sigma	Sigma	Sigma	Sigma	Sigma
± 0.15	14.49 ± 1.06	4.909 ± 0.265	3.146 ± 0.113	225 ± 5.4	3.821 ± 0.079
± 0.11	13.18 ± 0.97	3.734 ± 0.222	1.897 ± 0.120	164 ± 4.2	2.884 ± 0.113
± 0.10	8.33 ± 0.66	2.195 ± 0.119	2.350 ± 0.084	86 ± 3.0	2.924 ± 0.060
± 0.10	12.75 ± 0.94	2.478 ± 0.136	1.975 ± 0.083	216 ± 5.1	3.231 ± 0.067
± 0.10	19.40 ± 1.36	2.795 ± 0.166	2.381 ± 0.143	118 ± 3.4	2.421 ± 0.095
± 0.11	9.14 ± 0.71	4.622 ± 0.248	2.987 ± 0.115	103 ± 3.3	2.487 ± 0.052
± 0.10	12.01 ± 0.89	2.625 ± 0.156	2.715 ± 0.158	151 ± 4.0	2.306 ± 0.090
± 0.09	13.81 ± 0.99	4.101 ± 0.219	2.121 ± 0.078	68 ± 2.5	3.034 ± 0.061
± 0.08	10.02 ± 0.75	3.284 ± 0.195	1.867 ± 0.105	55 ± 2.2	1.822 ± 0.071
± 0.12	15.68 ± 1.15	2.137 ± 0.114	2.978 ± 0.111	96 ± 3.9	3.044 ± 0.062
± 0.09	7.69 ± 0.64	2.077 ± 0.123	2.120 ± 0.118	80 ± 3.2	1.865 ± 0.073
± 0.09	27.34 ± 1.89	2.286 ± 0.122	3.883 ± 0.126	116 ± 3.6	3.206 ± 0.065
± 0.11	21.50 ± 1.51	1.683 ± 0.100	2.116 ± 0.118	89 ± 3.1	2.388 ± 0.093
± 0.05	31.75 ± 2.18	5.692 ± 0.305	3.421 ± 0.112	103 ± 3.4	2.116 ± 0.044
± 0.03	6.47 ± 0.55	0.845 ± 0.051	1.336 ± 0.078	75 ± 2.8	2.103 ± 0.082
± 0.10	12.77 ± 0.93	2.613 ± 0.140	1.035 ± 0.035	65 ± 2.8	1.269 ± 0.026
± 0.10	9.07 ± 0.74	3.453 ± 0.205	1.870 ± 0.102	71 ± 2.7	1.572 ± 0.061
± 0.14	12.63 ± 1.01	4.886 ± 0.262	2.727 ± 0.116	58 ± 3.3	1.368 ± 0.029
± 0.11	22.06 ± 1.63	4.104 ± 0.244	4.193 ± 0.236	85 ± 4.3	2.480 ± 0.097
± 0.11	15.66 ± 1.11	2.799 ± 0.160	2.276 ± 0.103	102 ± 5.1	2.336 ± 0.048
± 0.15	20.31 ± 1.42	4.742 ± 0.281	3.101 ± 0.168	89.08 ± 2.96	1.636 ± 0.066
± 0.14	19.87 ± 1.39	4.806 ± 0.258	4.109 ± 0.138	137.07 ± 3.85	3.282 ± 0.069
± 0.10	(51.31 ± 46.95)	4.465 ± 0.267	4.050 ± 0.225	93.19 ± 3.28	2.489 ± 0.101
± 0.06	8.6 ± 8.6	5.777 ± 0.314	4.174 ± 0.179	67.63 ± 2.51	1.613 ± 0.040
± 0.08	11.5 ± 11.5	3.485 ± 0.209	3.930 ± 0.219	114.54 ± 3.17	2.878 ± 0.117
± 0.08	5.7 ± 5.7	2.545 ± 0.138	2.667 ± 0.092	55.90 ± 2.20	2.694 ± 0.060
± 0.237	29.275 ± 1.975	4.220 ± 0.228	3.439 ± 0.113	75.57 ± 2.80	2.300 ± 0.052
± 0.108	14.273 ± 0.978	11.21 ± 0.21	0.010 ± 0.010	128.30 ± 4.40	2.5 ± 0.3
± 0.135	17.640 ± 1.198	3.64 ± 0.07	2.151 ± 0.016	74.99 ± 3.66	1.3 ± 0.2
± 0.118	19.218 ± 1.308	5.86 ± 0.11	2.685 ± 0.015	62.29 ± 3.41	2 ± 0.2
± 0.103	19.298 ± 1.313	5.43 ± 0.09	1.841 ± 0.019	79.54 ± 3.77	1.8 ± 0.2
± 0.083	15.211 ± 1.041	3.69 ± 0.08	0.006 ± 0.006	74.36 ± 3.70	1.7 ± 0.2
± 0.121	13.917 ± 0.956	5.03 ± 0.09	2.698 ± 0.029	63.76 ± 3.68	0.9 ± 0.2
± 0.138	19.199 ± 1.329	6.73 ± 0.12	3.305 ± 0.020	103.22 ± 3.91	2.9 ± 0.3
± 0.109	11.599 ± 0.844	4.51 ± 0.09	2.102 ± 0.016	111.12 ± 4.49	3.3 ± 0.3
± 0.133	21.119 ± 1.438	2.74 ± 0.05	2.573 ± 0.011	106.02 ± 4.39	2.7 ± 0.3
± 0.062	14.978 ± 1.035	6.37 ± 0.13	3.587 ± 0.028	145.43 ± 4.65	3.9 ± 0.4
± 0.091	11.663 ± 0.817	2.52 ± 0.05	2.648 ± 0.024	182.68 ± 5.27	3.7 ± 0.4
± 0.131	15.728 ± 1.096	3.33 ± 0.07	1.395 ± 0.013	101.61 ± 4.23	1.1 ± 0.2
± 0.057	5.930 ± 0.463	2.92 ± 0.06	1.956 ± 0.014	104.99 ± 4.16	2.7 ± 0.3
± 0.091	17.285 ± 1.180	5.76 ± 0.10	1.055 ± 0.020	260.37 ± 6.41	6.1 ± 0.5
± 0.105	17.817 ± 1.222	4.06 ± 0.09	1.623 ± 0.014	57.71 ± 3.84	1.4 ± 0.2
± 0.090	13.653 ± 0.988	3.47 ± 0.06	0.009 ± 0.009	137.88 ± 4.79	2.6 ± 0.3
± 0.083	6.739 ± 0.491	3.39 ± 0.06	1.589 ± 0.014	384.81 ± 8.72	8.2 ± 0.7
± 0.088	15.806 ± 1.131	5.01 ± 0.09	1.593 ± 0.010	195.24 ± 6.24	4.4 ± 0.4
± 0.094	14.239 ± 0.779	2.14 ± 0.05	2.138 ± 0.021	54.91 ± 3.87	0.9 ± 0.2
± 0.135	15.591 ± 1.076	3.20 ± 0.06	0.007 ± 0.007	156.51 ± 5.58	3.8 ± 0.4
± 0.077	0.008 ± 0.008	2.88 ± 0.06	2.408 ± 0.017	99.89 ± 3.67	2.2 ± 0.2
± 0.143	12.327 ± 0.934	4.46 ± 0.11	0.003 ± 0.003	104.05 ± 4.60	2.3 ± 0.3
± 0.15	12.1 ± 1.4	7.2 ± 1.2	3.92 ± 0.11	103.50 ± 4.57	2.2 ± 0.3
± 0.24	15.1 ± 1.2	8.6 ± 1	4.56 ± 0.09	551.32 ± 11.62	8.5 ± 0.7
± 0.10	9.6 ± 0.7	1.6 ± 0.5	2.40 ± 0.07	(1786.6 419)	(26.4 22)
± 0.11	10.2 ± 0.8	2.91 ± 0.08	106 ± 5	171 ± 7	4.4 ± 0.5
± 0.18	15.2 ± 1.1	3.7 ± 0.6	2.91 ± 0.08	106 ± 5	3.1 ± 0.4
± 0.14	9.5 ± 0.7	3.3 ± 0.6	4.20 ± 0.09	111 ± 5	2.5 ± 0.3
± 0.13	18.4 ± 1.3	2.2 ± 0.5	3.57 ± 0.08	101 ± 5	1.4 ± 0.3
± 0.02	20 ± 0.5	1.9 ± 0.3	0.73 ± 0.07	45 ± 4	4.1 ± 0.5
± 0.04	4.5 ± 0.5	0.95 ± 0.07	4.08 ± 0.09	154 ± 6	4.1 ± 0.5
± 0.05	8.9 ± 0.6	1.64 ± 0.04	0.49 ± 0.11	27 ± 6	4.1 ± 0.5
± 0.05	6.9 ± 0.6	1.30 ± 0.06	0.95 ± 0.07	25 ± 4	1.8 ± 0.2
± 0.06	12.2 ± 0.9	6.8 ± 0.8	2.25 ± 0.07	66 ± 4	2.4 ± 0.3
				95 ± 5	2 ± 0.3

Se	Zn		V		Br		Fe		Mn	
	ng/m ³	Sigma								
0.13	269	1.9	2.4	0.6	4.38	0.08	175	6	7.3	0.7
0.09	89	0.7	3.6	0.6	2.76	0.06	97	5	2.3	0.3
0.06	100	0.8	5	0.7	1.55	0.06	80	5	2.1	0.3
0.06	90	0.7	3.2	0.5	2.29	0.07	43	4	0.9	0.2
0.07	179	1.3	7.6	0.9	4.52	0.09	68	5	1.4	0.3
0.17	140	1.1	4.9	0.7	4.73	0.08	134	6	2.7	0.3
0.06	95	0.7	3.7	0.5	2.97	0.09	35	4	0.6	0.2
0.21	269	1.9	4.8	0.7	12.11	0.18	111	6	9.7	0.9
0.10	180	1.3	3.1	0.5	3.09	0.07	113	5	3.3	0.4
0.08	62	0.5	14.8	1.6	4.58	0.11	44	4	0.8	0.2
0.28	130	1.0	4.4	0.7	5.77	0.11	89	5	1.8	0.3
0.06	86	0.7	3.7	0.6	1.70	0.07	69	5		
0.11	88	0.7	1.5	0.4	1.63	0.07	57	4	2.2	0.3
0.03	10	0.2	1.3	0.5	0.27	0.06	7	3		
0.08	9.1	0.7	1.4	0.5	2.83	0.06	74	5	2.5	0.3
0.08	14.4	1.0	6.7	0.9	4.31	0.07	68	4	1.9	0.3
0.10	99	0.7	2	0.5	3.62	0.06	85	4	1.4	0.3
0.07	12.1	0.9	8.7	1	6.19	0.08	53	4	1.5	0.3
0.11	9.6	0.7	1.2	0.5	3.32	0.09	35	4	1.3	0.02
0.09	9.1	0.7	1.4	0.5	2.88	0.06	61	4	2.5	0.3
0.19	17.4	1.3	3.5	0.6	5.50	0.09	103	5	3	0.4
0.09	12.6	0.9	1.5	0.6	2.96	0.06	74	4	3.2	0.4
0.05	3.6	0.4	5	0.7	2.36	0.07	41	4	1.6	0.3
0.07	9.5	0.7	5.6	0.8	1.78	0.06	60	4	2.2	0.3
0.04	8.7	0.7	5.6	0.7	3.48	0.08	100	4	2.2	0.3
0.14	11.7	0.9	5.3	0.7	3.83	0.07	99	5	3.9	0.4
0.05	4.5	0.5	2.8	0.6	4.38	0.12	99	4	2.4	0.4
0.06	11.9	0.9	4.1	0.7	3.10	0.07	84	5	2.5	0.4
0.04	4.1	0.4	5.2	0.7	2.87	0.07	63	4	1.1	0.3
0.08	6.0	0.5	3.8	0.6	3.29	0.07	40	4	1.5	0.3
0.09	9.7	0.8	0.6	0.6	3.07	0.08	116	5	3.4	0.4
0.16	14.2	1.0	1.2	0.5	2.54	0.06	112	5	3.1	0.4
0.15	8.7	0.7	2.4	0.6	5.33	0.07	78	6	1.7	0.3
0.08	4.9	0.5	3.2	0.6	3.41	0.08	68	4	2.2	0.3
0.15	16.6	1.2	1.9	0.8	3.51	0.09	121	6	4.5	0.5
0.21	22.4	1.6	2.7	0.7	5.07	0.07	147	6	5.8	0.8
0.15	9.0	0.7	6.3	0.9	4.42	0.10	224	8	3.2	0.4
0.25	13.1	1.1	8.3	1.1	5.40	0.08	278	9	6.8	0.7
0.18	26.4	1.9	4.1	0.7	4.51	0.14	236	8	6.7	0.7
0.13	11.5	1.0	7.5	1	4.26	0.14	233	9	4.4	0.5
0.13	18.4	1.3	4.5	0.7	4.25	0.07	118	5	2.8	0.4
0.11	6.2	0.5	6.6	0.9	2.29	0.07	52	4	1.5	0.3
0.21	17.3	1.3	11.5	1.3	4.86	0.08	150	6	2.4	0.4

139	3.2	3.076	117	3.127
80	1.3	1.446	50	1.535
1.6	0.3	0.256	6	0.135
48	47	47	48	49
54	0.6	0.064	33	0.027
48.2	7.0	8.429	283	8.202
12.2	3.2	2.778	102	3.054
3.5	2.2	2.7	2.4	2.6

Se	Zn	V	Br	Fe	Mn
Sigma	ng/m ³				
Sigma	Sigma	Sigma	Sigma	Sigma	Sigma
14.1	3.3	2.915	114	2.897	1.331
7.4	1.3	1.260	48	0.112	77
1.4	0.2	0.212	5	0.027	8.202
76	75	75	76	2.871	2.8
4.7	0.6	0.084	33		
48.2	7.0	8.429	293		
12.6	3.2	2.738	98		
3.4	2.1	2.9	2.5		
44.7	3.9	2.412	113	2.616	1.241
5.9	1.7	1.000	61	0.159	51
1.6	0.2	0.093	4	0.900	8.200
50	51	51	55	2.487	3.1
4.7	0.8	0.006	385		
31.8	11.2	4.183	96		
13.8	3.7	2.381	3.4		
2.2	2.9	1.7	104	2.929	1.868
11.4	4.3	3.328	79	0.369	55
5.8	2.7	1.834	5	0.600	9.700
0.9	0.6	0.075	60	2.350	3.3
61	56	61	7		
0.0	0.8	0.003	551		
26.9	14.8	12.108	89		
10.0	3.7	3.291	5.3		
2.4	3.4	3.6			

Appendix E-2. Concentrations of Elements in Chesapeake Bay Aerosol.

(Elms Site)

(Elms Site)

Sample ID	Year	Al ng/m ³ Sigma	As ng/m ³ Sigma	Cd ng/m ³ Sigma	Cr ng/m ³ Sigma	Cu ng/m ³ Sigma	Ni ng/m ³ Sigma	Pb ng/m ³ Sigma	S ng/m ³ Sigma
E1*	06/05/90								
E2	06/12/90	29.4 ± 2.8	0.323 ± 0.045	< 0.017	0.017	< 0.160	0.160	< 0.062 ± 0.052	< 0.76
E3	06/20/90	354 ± 30	0.379 ± 0.053	0.0968 ± 0.012	0.706 ± 0.14	(9.65 ± 0.55)	(11.1 ± 0.68)	3.08 ± 0.32	1726 ± 531
E5	06/26/90	293 ± 25	0.410 ± 0.026	0.0894 ± 0.0087	0.733 ± 0.17	(7.99 ± 0.51)	(11.2 ± 0.67)	3.31 ± 0.31	4026 ± 477
E6	07/03/90	(10380 ± 902)*	(2.95 ± 0.31)*	(2.59 ± 1.1)*	ND	(30.3 ± 18)*	(104 ± 44)*	(90.4 ± 45)*	(8224 ± 16426)*
E7	07/10/90	184 ± 16	0.294 ± 0.044	0.0654 ± 0.0078	0.185 ± 0.082	7.41 ± 0.47	7.91 ± 0.50	2.28 ± 0.26	2490 ± 444
E8	07/17/90	286 ± 25	0.674 ± 0.089	0.139 ± 0.012	1.05 ± 0.13	2.13 ± 0.20	3.81 ± 0.42	5.77 ± 0.52	4384 ± 588
E9	07/24/90	87.9 ± 7.7	0.472 ± 0.069	0.145 ± 0.017	0.852 ± 0.18	2.17 ± 0.28	1.65 ± 0.23	4.96 ± 0.80	3262 ± 563
E10	07/31/90	160 ± 14	0.482 ± 0.072	0.0902 ± 0.020	0.730 ± 0.18	1.21 ± 0.20	1.65 ± 0.23	2.10 ± 0.58	2468 ± 378
E11	08/07/90	152 ± 13	1.04 ± 0.13	0.223 ± 0.020	0.913 ± 0.18	3.09 ± 0.31	(6.05 ± 0.55)	9.24 ± 0.94	5202 ± 642
E12	08/14/90	147 ± 13	0.540 ± 0.084	0.130 ± 0.015	0.520 ± 0.17	2.59 ± 0.27	(7.44 ± 0.50)	5.42 ± 0.89	4050 ± 703
E13	08/21/90	85.2 ± 7.4	0.832 ± 0.12	0.158 ± 0.015	0.623 ± 0.18	2.18 ± 0.24	7.27 ± 0.46	5.85 ± 0.53	6324 ± 734
E14	08/28/90	309 ± 27	0.688 ± 0.10	0.168 ± 0.016	0.889 ± 0.19	1.94 ± 0.23	0.862 ± 0.12	6.40 ± 0.52	5001 ± 864
E15	09/04/90	435 ± 37	0.805 ± 0.13	0.142 ± 0.015	0.972 ± 0.23	1.94 ± 0.24	0.677 ± 0.14	5.22 ± 0.88	5182 ± 983
E16	09/11/90	246 ± 21	0.857 ± 0.12	0.151 ± 0.021	0.946 ± 0.20	2.23 ± 0.23	0.665 ± 0.16	5.65 ± 1.1	3904 ± 502
E17	09/18/90	118 ± 10	0.888 ± 0.12	0.178 ± 0.019	1.407 ± 0.23	1.96 ± 0.22	3.75 ± 0.26	5.87 ± 0.51	2303 ± 297
E18	09/25/90	274 ± 24	1.56 ± 0.18	0.435 ± 0.031	1.407 ± 0.23	3.91 ± 0.33	5.14 ± 0.39	14.2 ± 1.1	4439 ± 653
E19	10/02/90	173 ± 15	0.697 ± 0.10	0.120 ± 0.014	0.611 ± 0.19	1.36 ± 0.22	0.564 ± 0.15	3.84 ± 0.56	1302 ± 407
E20	10/09/90	82.1 ± 7.2	< 0.085 ± 0.085	0.0630 ± 0.013	0.501 ± 0.18	0.555 ± 0.20	0.714 ± 0.14	2.23 ± 0.44	1035 ± 416
E21	10/16/90	78.7 ± 6.8	0.601 ± 0.085	0.120 ± 0.012	0.440 ± 0.12	3.69 ± 0.28	5.38 ± 0.48	4.84 ± 0.61	736 ± 228
E22	10/24/90	(3438 ± 302)*	(45.5 ± 5.5)*	(13.4 ± 1.4)*	(58 ± 4.0)*	(178 ± 21)*	(281 ± 72)*	(526 ± 79)*	(8206 ± 14657)*
E23	11/3/90	141 ± 12	1.43 ± 0.16	0.230 ± 0.020	0.857 ± 0.12	2.69 ± 0.22	3.20 ± 0.27	11.5 ± 0.75	1535 ± 291
E24	11/20/90	266 ± 23	1.42 ± 0.16	0.394 ± 0.032	1.42 ± 0.16	3.00 ± 0.27	(61.2 ± 4.3)*	11.9 ± 0.97	2682 ± 529
E25	11/27/90	89.4 ± 7.7	0.924 ± 0.060	0.189 ± 0.015	0.825 ± 0.23	2.42 ± 0.21	1.09 ± 0.19	7.21 ± 0.55	2380 ± 605
E26	12/04/90	97.1 ± 8.4	0.811 ± 0.053	0.235 ± 0.017	1.070 ± 0.23	8.28 ± 0.52	5.86 ± 0.43	9.32 ± 0.61	2183 ± 373
E27	12/11/90	93.0 ± 8.0	1.009 ± 0.063	0.181 ± 0.023	0.553 ± 0.22	2.49 ± 0.21	1.00 ± 0.14	6.53 ± 0.58	1627 ± 247
E28	12/18/90	23.4 ± 2.1	0.378 ± 0.025	0.091 ± 0.009	0.102 ± 0.102	2.94 ± 0.22	1.45 ± 0.15	2.54 ± 0.29	1730 ± 280
E29	12/26/90	42.7 ± 3.8	0.710 ± 0.047	0.150 ± 0.013	0.423 ± 0.20	1.82 ± 0.18	4.07 ± 0.34	5.85 ± 0.41	2944 ± 427
E30	01/09/91	52.5 ± 4.7	0.765 ± 0.052	0.200 ± 0.017	1.34 ± 0.31	2.23 ± 0.25	5.48 ± 0.42	7.27 ± 0.93	1789 ± 288
E31	01/15/91	101 ± 8.8	0.721 ± 0.046	0.139 ± 0.015	1.32 ± 0.33	1.97 ± 0.25	1.30 ± 0.19	4.27 ± 0.78	1923 ± 412
E32	01/22/91	64.4 ± 5.6	0.730 ± 0.045	0.149 ± 0.016	1.13 ± 0.32	2.16 ± 0.25	2.19 ± 0.20	4.12 ± 0.39	ND
E33	01/29/91	102 ± 8.9	0.449 ± 0.032	0.072 ± 0.013	0.700 ± 0.23	6.19 ± 0.45	4.72 ± 0.36	3.13 ± 0.37	ND
E34	02/05/91	143 ± 12	0.592 ± 0.037	0.148 ± 0.017	1.30 ± 0.35	1.81 ± 0.26	2.44 ± 0.28	2.26 ± 0.44	ND
E35	02/12/91	106 ± 9.3	0.916 ± 0.057	0.101 ± 0.016	1.27 ± 0.34	3.04 ± 0.29	1.61 ± 0.21	2.26 ± 0.41	1884 ± 520
E36	02/19/91	78.7 ± 6.8	0.449 ± 0.032	0.072 ± 0.013	0.302 ± 0.302	1.08 ± 0.24	2.14 ± 0.26	2.01 ± 0.36	1504 ± 287
E37	02/26/91	58.6 ± 5.1	0.495 ± 0.032	0.069 ± 0.012	0.688 ± 0.29	1.95 ± 0.20	0.84 ± 0.26	2.26 ± 0.41	ND
E38	03/05/91	102 ± 8.8	0.571 ± 0.036	0.122 ± 0.015	0.275 ± 0.275	1.31 ± 0.22	4.22 ± 0.28	4.12 ± 0.43	1314 ± 336
E39	03/12/91	59.1 ± 5.2	0.446 ± 0.029	0.072 ± 0.015	0.070 ± 0.070	1.24 ± 0.25	2.43 ± 0.23	2.96 ± 0.43	2438 ± 329
E40	03/19/91	125 ± 11	0.674 ± 0.043	0.113 ± 0.017	0.178 ± 0.178	2.57 ± 0.30	4.19 ± 0.31	5.52 ± 0.72	2603 ± 403
E41	03/26/91	144 ± 12	0.455 ± 0.031	0.110 ± 0.014	0.144 ± 0.144	2.50 ± 0.26	1.65 ± 0.22	3.59 ± 0.39	2059 ± 321
E42	04/02/91	142 ± 12	0.473 ± 0.032	0.062 ± 0.010	0.724 ± 0.24	0.160 ± 0.160	< 0.082 ± 0.082	1.58 ± 0.25	ND
E43	04/09/91	161 ± 14	0.542 ± 0.039	< 0.010 ± 0.010	0.194 ± 0.194	1.59 ± 0.23	2.21 ± 0.18	1.59 ± 0.38	ND
E44	04/16/91	64.8 ± 5.7	0.327 ± 0.023	0.187 ± 0.026	0.077 ± 0.077	0.74 ± 0.21	2.89 ± 0.40	3.61 ± 0.37	2647 ± 439
E45	04/23/91	207 ± 18	0.600 ± 0.039	0.0816 ± 0.019	0.165 ± 0.165	1.94 ± 0.24	1.67 ± 0.19	3.87 ± 0.68	3988 ± 689
E46	04/30/91	11.8 ± 2.2	0.108 ± 0.007	< 0.012 ± 0.012	ND	0.200 ± 0.200	0.616 ± 0.14	< 0.262 ± 0.262	989 ± 96
E47	05/07/91	30.7 ± 3.8	0.399 ± 0.027	0.052 ± 0.018	0.035 ± 0.035	< 0.200 ± 0.200	0.148 ± 0.148	0.19 ± 0.19	4127 ± 414
E48	05/14/91	180 ± 16	0.285 ± 0.024	< 0.005 ± 0.005	0.206 ± 0.206	0.280 ± 0.280	0.513 ± 0.14	0.19 ± 0.19	5099 ± 1280
E49	05/21/91	566 ± 49	0.671 ± 0.059	0.066 ± 0.012	1.16 ± 0.29	1.14 ± 0.21	1.77 ± 0.19	4.14 ± 0.53	2343 ± 1111
E50	05/29/91	369 ± 32	0.802 ± 0.050	0.266 ± 0.032	0.885 ± 0.38	0.932 ± 0.27	1.77 ± 0.21	6.12 ± 0.76	4550 ± 721
E51	6/04/91	295 ± 25	0.340 ± 0.021	< 0.012 ± 0.012	0.238 ± 0.238	< 0.150 ± 0.150	< 0.06 ± 0.06	< 0.18 ± 0.18	3146 ± 498
E52	6/11/91	350 ± 30	0.665 ± 0.042	0.0957 ± 0.019	0.060 ± 0.060	< 0.330 ± 0.330	< 0.139 ± 0.139	< 1.77 ± 0.48	3205 ± 569
E53	6/18/91	55.8 ± 4.9	0.309 ± 0.021	0.172 ± 0.024	0.295 ± 0.295	1.31 ± 0.26	3.27 ± 0.26	4.75 ± 0.62	3943 ± 489
E54	6/25/91	395 ± 34	0.316 ± 0.025	0.146 ± 0.017	0.909 ± 0.37	< 0.37 ± 0.37	0.820 ± 0.16	2.86 ± 0.64	6040 ± 930
E55	7/02/91	128 ± 11	0.424 ± 0.027	0.219 ± 0.054	0.108 ± 0.108	3.910 ± 0.340	2.24 ± 0.21	4.55 ± 0.67	4806 ± 534
E56	7/09/91	117 ± 10	0.483 ± 0.030	0.112 ± 0.016	0.797 ± 0.35	1.59 ± 0.26	3.11 ± 0.33	4.33 ± 0.47	4644 ± 546
E57	7/16/91	531.9 ± 32.8	0.43 ± 0.01	0.1849 ± 0.0721	1.15 ± 0.06	0.90 ± 0.19	4.31 ± 0.23	1.27 ± 0.30	6638 ± 690
E58	7/23/91	178.1 ± 11.1	0.52 ± 0.02	ND	0.52 ± 0.04	< 0.09 ± 0.09	ND	< 1.27 ± 0.30	3308 ± 355

Sample ID

Sample ID	Year	Al ng/m ³ Sigma	As ng/m ³ Sigma	Cd ng/m ³ Sigma	Cr ng/m ³ Sigma	Cu ng/m ³ Sigma	Ni ng/m ³ Sigma	Pb ng/m ³ Sigma	S ng/m ³ Sigma
E59	7/30/91	108.5 ± 6.9	0.34 ± 0.01	0.1420 ± 0.0414	0.43 ± 0.03	1.05 ± 0.16	0.79 ± 0.19	1.44 ± 0.15	3885 ± 330
E60	8/06/91	232.3 ± 14.4	0.39 ± 0.01	0.1816 ± 0.0427	1.03 ± 0.05	0.91 ± 0.15	0.77 ± 0.17	1.85 ± 0.21	3043 ± 463
E61	8/13/91	189.1 ± 11.8	0.66 ± 0.03	0.0560 ± 0.0244	0.73 ± 0.05	0.11 ± 0.11	0.09 ± 0.09	0.21 ± 0.21	5534 ± 506
E62	8/20/91	162.5 ± 10.3	0.70 ± 0.05	0.0950 ± 0.0031	0.70 ± 0.05	0.90 ± 0.18	3.96 ± 0.28	1.48 ± 0.36	6249 ± 631
E63	8/27/91	216.3 ± 13.5	0.36 ± 0.03	0.1342 ± 0.1342	0.69 ± 0.04	3.32 ± 0.18	1.51 ± 0.30	1.19 ± 0.33	2878 ± 699
E64	9/03/91	132.5 ± 8.4	0.77 ± 0.02	0.2664 ± 0.0785	0.91 ± 0.04	2.74 ± 0.18	2.70 ± 0.24	6.34 ± 0.44	3594 ± 463
E65	9/10/91	195.0 ± 12.2	0.57 ± 0.04	(0.649) ± (0.100)	0.80 ± 0.05	2.12 ± 0.19	1.41 ± 0.21	0.49 ± 0.49	4530 ± 171
E66	9/17/91	82.1 ± 5.3	0.50 ± 0.02	0.0397 ± 0.0031	0.67 ± 0.04	0.44 ± 0.15	0.14 ± 0.14	0.55 ± 0.15	1897 ± 171
E67	9/24/91	93.6 ± 6.0	0.49 ± 0.02	0.0749 ± 0.0023	0.63 ± 0.03	0.66 ± 0.15	0.62 ± 0.18	1.21 ± 0.40	2403 ± 394
E68	10/02/91	51.2 ± 3.6	0.46 ± 0.02	0.0777 ± 0.0036	0.30 ± 0.04	0.78 ± 0.21	ND	0.55 ± 0.18	1708 ± 237
E69	10/08/91	47.3 ± 3.2	0.34 ± 0.01	0.0702 ± 0.0025	0.29 ± 0.03	0.37 ± 0.15	0.78 ± 0.19	2.08 ± 0.46	1001 ± 177
E70	10/15/91	57.3 ± 3.8	0.56 ± 0.01	0.0910 ± 0.0034	0.42 ± 0.03	0.73 ± 0.18	0.52 ± 0.52	3.04 ± 0.79	4684 ± 506
E71	10/22/91	142.8 ± 9.7	0.75 ± 0.02	0.2213 ± 0.0085	0.59 ± 0.08	1.30 ± 0.42	6.32 ± 0.23	2.41 ± 0.17	1488 ± 281
E72	10/29/91	153.2 ± 9.6	0.86 ± 0.01	0.1024 ± 0.0172	1.24 ± 0.04	1.26 ± 0.13	(8.78) ± (0.201)	1.76 ± 0.49	2766 ± 375
E73	11/05/91	179.9 ± 11.3	0.71 ± 0.01	0.0731 ± 0.0731	1.04 ± 0.05	1.28 ± 0.14	(8.21) ± (0.166)	5.30 ± 0.26	1862 ± 265
E74	11/12/91	114.1 ± 7.2	0.27 ± 0.01	0.1739 ± 0.0020	0.52 ± 0.04	2.45 ± 0.14	(12.2) ± (0.279)	0.26 ± 0.26	1691 ± 365
E75	11/19/91	45.7 ± 3.1	0.70 ± 0.01	0.2268 ± 0.1268	0.30 ± 0.03	1.75 ± 0.13	(10.36) ± (0.179)	1.44 ± 0.13	1178 ± 221
E76	11/26/91	72.4 ± 4.7	0.50 ± 0.01	0.1749 ± 0.0749	0.42 ± 0.03	3.09 ± 0.15	7.20 ± 0.21	2.92 ± 0.62	1571 ± 218
E77	12/04/91	84.9 ± 5.5	0.73 ± 0.01	0.2659 ± 0.0092	0.59 ± 0.04	2.64 ± 0.14	4.93 ± 0.23	1.93 ± 0.22	1724 ± 272
E78	12/10/91	69.2 ± 4.5	1.05 ± 0.01	0.3539 ± 0.0813	0.54 ± 0.03	2.75 ± 0.14	7.90 ± 0.21	ND	1459 ± 241
E79	12/17/91	119.2 ± 7.5	1.01 ± 0.01	0.0560 ± 0.0560	0.88 ± 0.03	0.76 ± 0.13	5.95 ± 0.21	6.36 ± 0.23	1507 ± 292
E80	12/24/91	34.9 ± 2.4	0.75 ± 0.01	0.0317 ± 0.0317	0.40 ± 0.02	2.03 ± 0.12	9.10 ± 0.21	5.61 ± 0.28	2478 ± 353
E81	1/01/92	33.5 ± 2.4	0.78 ± 0.02	0.1144 ± 0.0541	0.49 ± 0.04	2.49 ± 0.13	3.74 ± 0.15	1.25 ± 0.27	1611 ± 165
E82	1/09/92	35.5 ± 2.7	0.66 ± 0.01	0.0652 ± 0.0028	0.46 ± 0.05	1.19 ± 0.18	0.67 ± 0.22	3.42 ± 0.37	975 ± 191
E83	1/14/92	108.6 ± 6.9	0.54 ± 0.01	0.0818 ± 0.0818	0.53 ± 0.03	1.38 ± 0.12	0.95 ± 0.19	3.45 ± 0.39	1924 ± 287
E84	1/21/92	56.0 ± 3.7	0.78 ± 0.01	0.1153 ± 0.0172	0.65 ± 0.03	1.14 ± 0.14	1.43 ± 0.21	11.6 ± 1.2	1864 ± 144
E85	1/28/92	51.20 ± 2.44	1.90 ± 0.01	0.135 ± 0.006	1.46 ± 0.06	4.1 ± 0.5	3.69 ± 0.15	11.6 ± 1.2	847 ± 65
E86	2/04/92	62.09 ± 2.56	0.19 ± 0.00	0.140 ± 0.009	0.79 ± 0.06	1.7 ± 0.2	3.32 ± 0.12	5.3 ± 0.6	1478 ± 114
E87	2/11/92	36.98 ± 1.84	1.02 ± 0.01	0.257 ± 0.008	1.22 ± 0.06	1.3 ± 0.2	3.39 ± 0.04	5.8 ± 0.7	1411 ± 109
E88	2/18/92	55.08 ± 2.68	0.30 ± 0.00	0.187 ± 0.003	0.39 ± 0.06	1.3 ± 0.2	2.59 ± 0.12	4.9 ± 0.6	1324 ± 103
E89	2/25/92	27.77 ± 1.43	0.26 ± 0.01	0.438 ± 0.005	0.55 ± 0.05	1.6 ± 0.2	2.81 ± 0.05	4.6 ± 0.5	2100 ± 166
E90	3/03/92	111.10 ± 5.24	0.43 ± 0.01	0.129 ± 0.002	0.33 ± 0.06	0.8 ± 0.2	2.89 ± 0.05	4.7 ± 0.6	1110 ± 88
E91	3/10/92	97.29 ± 4.58	0.41 ± 0.00	0.078 ± 0.004	0.63 ± 0.06	1.1 ± 0.2	3.11 ± 0.04	4.2 ± 0.5	1532 ± 121
E92	3/17/92	134.31 ± 6.32	0.27 ± 0.00	0.210 ± 0.002	1.40 ± 0.07	1.8 ± 0.3	2.31 ± 0.03	8.1 ± 0.9	1169 ± 92
E93	3/24/92	148.36 ± 6.89	0.72 ± 0.01	0.152 ± 0.004	1.00 ± 0.06	1.3 ± 0.2	2.68 ± 0.07	5 ± 0.6	1394 ± 110
E94	3/31/92	268.33 ± 12.34	0.87 ± 0.01	0.227 ± 0.003	1.58 ± 0.06	1.5 ± 0.2	2.34 ± 0.04	6.2 ± 0.7	2018 ± 159
E95	4/07/92	84.66 ± 4.51	0.50 ± 0.01	0.106 ± 0.007	1.05 ± 0.06	2.3 ± 0.3	3.14 ± 0.04	5.2 ± 0.7	1388 ± 110
E96	4/14/92	300.64 ± 13.87	0.21 ± 0.01	0.117 ± 0.006	0.41 ± 0.06	0.9 ± 0.2	2.13 ± 0.04	2.5 ± 0.4	1374 ± 108
E97	4/21/92	169.30 ± 7.95	0.95 ± 0.01	0.213 ± 0.003	0.35 ± 0.06	1.2 ± 0.2	2.18 ± 0.05	5.6 ± 0.7	1725 ± 136
E98	4/28/92	312.96 ± 14.42	0.87 ± 0.01	0.207 ± 0.007	0.83 ± 0.06	1.6 ± 0.2	7.59 ± 0.08	5.3 ± 0.7	2349 ± 185
E99	5/05/92	62.07 ± 2.97	0.43 ± 0.01	0.143 ± 0.009	0.84 ± 0.08	2.6 ± 0.3	5.17 ± 0.05	5.5 ± 0.7	2218 ± 175
E100	5/12/92	65.51 ± 3.11	0.37 ± 0.01	0.058 ± 0.003	0.99 ± 0.08	1.4 ± 0.2	0.98 ± 0.05	3.5 ± 0.5	1491 ± 118
E101	5/19/92	180.52 ± 8.44	0.63 ± 0.01	0.045 ± 0.003	0.78 ± 0.06	0.7 ± 0.2	4.15 ± 0.17	2.6 ± 0.4	1734 ± 137
E102	6/02/92	71.33 ± 3.38	0.08 ± 0.00	0.43 ± 0.013	0.89 ± 0.06	(6.2) ± (0.7)	15.55 ± 0.38	5.30 ± 0.60	1733 ± 137
E103	6/09/92	74.98 ± 3.59	0.73 ± 0.01	0.23 ± 0.003	0.31 ± 0.06	1.50 ± 0.20	(32.48 ± 0.74)	1.70 ± 0.30	814 ± 64
E104	6/16/92	105.41 ± 4.92	0.34 ± 0.01	0.51 ± 0.030	0.70 ± 0.06	0.70 ± 0.20	10.13 ± 0.19	4.80 ± 0.60	3073 ± 242
E105	6/24/92	(1378.158) ± (63.978)	0.33 ± 0.01	0.34 ± 0.006	0.57 ± 0.06	1.90 ± 0.30	6.63 ± 0.03	3.70 ± 0.50	2118 ± 167
E106	7/01/92	3623.964.7)	0.79 ± 0.04	0.144 ± 0.005	1.98 ± 0.10	1.50 ± 0.30	17.16 ± 0.15	4.80 ± 0.60	3170 ± 250
E107	7/07/92	ND	0.66 ± 0.03	0.174 ± 0.003	3.98 ± 0.21	4.9 ± 0.6	1.82 ± 0.23	3.4 ± 0.7	3039 ± 240
E108	7/14/92	ND	ND	0.173 ± 0.002	0.64 ± 0.13	1 ± 0.3	2.84 ± 0.36	6.2 ± 0.8	5662 ± 462
E109	7/29/92	198.6	0.48 ± 0.02	0.173 ± 0.002	0.70 ± 0.12	1.4 ± 0.3	1.89 ± 0.17	5.3 ± 0.7	2032 ± 160
E110	8/11/92	ND	0.85 ± 0.03	0.120 ± 0.003	0.06 ± 0.06	0.6 ± 0.2	2.21 ± 0.16	ND	2907 ± 229
E111	8/18/92	60.3	0.60 ± 0.04	0.126 ± 0.009	0.40 ± 0.11	7.2 ± 0.8	1.32 ± 0.20	3.2 ± 0.5	3341 ± 263
E112	8/25/92	ND	0.21 ± 0.03	0.256 ± 0.030	0.40 ± 0.10	ND	1.86 ± 0.66	8.3 ± 1.8	2579 ± 204
E113	9/08/92	ND	0.26 ± 0.01	0.129 ± 0.018	ND	ND	7.40 ± 0.84	ND	1037 ± 83
E114	9/09/92	152	0.74 ± 0.03	0.037 ± 0.013	0.97 ± 0.15	ND	0.22 ± 0.22	5.8 ± 1.6	870 ± 70
E115	9/15/92	233.6	0.35 ± 0.02	0.184 ± 0.004	0.05 ± 0.05	4.1 ± 0.5	1.05 ± 0.21	6.8 ± 0.9	485 ± 39
E116	9/29/92	8	0.12 ± 0.01	0.396 ± 0.014	0.51 ± 0.14	1.7 ± 0.3	0.98 ± 0.18	ND	205 ± 16
E117	10/06/92	68.8	0.74 ± 0.03	0.104 ± 0.019	0.45 ± 0.13	2 ± 0.4	1.26 ± 0.27	5.5 ± 0.8	1867 ± 147
E118	10/13/92	ND	0.35 ± 0.02	0.156 ± 0.006	0.64 ± 0.16	2.1 ± 0.4	1.54 ± 0.22	3.4 ± 0.6	799 ± 63
E119	10/20/92	ND	0.70 ± 0.03	0.176 ± 0.008	0.59 ± 0.14	2.5 ± 0.4	0.98 ± 0.23	2.9 ± 0.7	1084 ± 86
E120	11/03/92	ND	0.94 ± 0.03	0.147 ± 0.003	0.62 ± 0.15	3.3 ± 0.5	3.63 ± 0.24	6.9 ± 0.9	830 ± 66
E121	11/10/92	212.2	0.32 ± 0.04	0.147 ± 0.003	0.38 ± 0.14	1.5 ± 0.3	0.47 ± 0.20	4.9 ± 0.7	1138 ± 90
E122	11/17/92	65.3	0.32 ± 0.04	0.147 ± 0.003	0.38 ± 0.14	1.5 ± 0.3	2.57 ± 0.21	2.9 ± 0.6	966 ± 76
E123	11/24/92	ND	0.32 ± 0.04	0.147 ± 0.003	0.38 ± 0.14	1.5 ± 0.3	2.57 ± 0.21	2.9 ± 0.6	966 ± 76

Year 3

Sample ID	Year 1	Al	As	Cd	Cr	Cu	Ni	Pb	S			
		ng/m ³ Sigma										
E124	12/01/92	ND	1.29	0.021	0.72	0.14	0.44	0.6	7.4	0.9	1779	141
E125	12/08/92	ND	0.76	0.004	0.97	0.08	1.7	0.3	6.8	0.8	1528	121
E126	12/15/92	ND	0.50	0.005	<	0.10	2.5	0.4	3.3	0.7	3071	242
E127	12/22/92	ND	<	0.02	ND	ND	ND	ND	ND	ND	ND	ND
E128	12/29/92	ND	<	0.04	ND	ND	ND	ND	ND	ND	ND	ND
E129	1/05/93	ND	<	0.04	0.04	0.849	0.068	0.139	0.024	0.139	0.024	0.139
E130	1/19/93	3.4	ND	(2.73	0.066)	0.068	0.005	0.129	0.007	0.068	0.005	0.129
E131	1/26/93	25.9	ND	0.136	0.003	0.090	0.003	0.059	0.010	0.136	0.003	0.059
E132	2/02/93	88.6	0.75	0.068	0.005	0.068	0.005	0.068	0.005	0.068	0.005	0.068
E133	2/09/93	17.7	0.50	0.129	0.007	0.129	0.007	0.129	0.007	0.129	0.007	0.129
E134	2/16/93	6.8	0.29	0.090	0.003	0.090	0.003	0.090	0.003	0.090	0.003	0.090
E135	3/02/93	13.6	0.54	0.068	0.003	0.068	0.003	0.068	0.003	0.068	0.003	0.068
E136	3/09/93	78.1	0.40	0.068	0.003	0.068	0.003	0.068	0.003	0.068	0.003	0.068
E137	3/16/93	12.75	0.40	0.068	0.003	0.068	0.003	0.068	0.003	0.068	0.003	0.068
E138	3/16/93	14.6	0.38	0.136	0.003	0.136	0.003	0.136	0.003	0.136	0.003	0.136
E139	3/23/93	7	0.35	<	0.005	0.005	0.005	0.005	0.005	0.005	0.005	0.005
E140	3/30/93	38.7	0.40	0.092	0.018	0.092	0.018	0.092	0.018	0.092	0.018	0.092
E141	4/06/93	64.9	0.30	0.073	0.007	0.073	0.007	0.073	0.007	0.073	0.007	0.073
E142	4/13/93	44.1	0.36	0.036	0.005	0.036	0.005	0.036	0.005	0.036	0.005	0.036
E143	4/20/93	132.8	0.26	0.036	0.005	0.036	0.005	0.036	0.005	0.036	0.005	0.036
E144	5/04/93	48.6	1.07	0.102	0.005	0.102	0.005	0.102	0.005	0.102	0.005	0.102
E145	5/11/93	294.5	0.62	0.089	0.008	0.089	0.008	0.089	0.008	0.089	0.008	0.089
E146	5/18/93	12.35	0.57	0.055	0.004	0.055	0.004	0.055	0.004	0.055	0.004	0.055
E147	5/25/93	245.3	0.62	0.089	0.008	0.089	0.008	0.089	0.008	0.089	0.008	0.089
E148	6/01/93	49.8	0.44	0.044	0.002	0.044	0.002	0.044	0.002	0.044	0.002	0.044
E149	6/08/93	190.9	0.44	0.044	0.002	0.044	0.002	0.044	0.002	0.044	0.002	0.044
E150	6/15/93	68.3	0.44	0.044	0.002	0.044	0.002	0.044	0.002	0.044	0.002	0.044
E151	6/22/93	70.9	0.65	0.062	0.002	0.062	0.002	0.062	0.002	0.062	0.002	0.062
E152	6/29/93	59.3	1.21	0.044	0.002	0.044	0.002	0.044	0.002	0.044	0.002	0.044
E153	7/06/93	239	0.55	0.057	0.002	0.057	0.002	0.057	0.002	0.057	0.002	0.057
E154	7/13/93	164.7	1.05	0.003	0.003	0.003	0.003	0.003	0.003	0.003	0.003	0.003
E155	7/20/93	96.5	0.646	0.13	0.638	0.13	0.638	0.13	0.638	0.13	0.638	0.13
E156	7/27/93	155	0.308	0.08	0.403	0.08	0.403	0.08	0.403	0.08	0.403	0.08
E157	8/3/93	115	0.065	0.02	0.197	0.02	0.197	0.02	0.197	0.02	0.197	0.02
E158	8/10/93	13	46	46	45	46	45	46	45	46	45	46
E159	8/17/93	46	0.085	0.01	0.035	0.01	0.035	0.01	0.035	0.01	0.035	0.01
E160	8/24/93	12	0.566	0.43	1.407	0.43	1.407	0.43	1.407	0.43	1.407	0.43
E161	8/31/93	566	0.595	0.12	0.645	0.12	0.645	0.12	0.645	0.12	0.645	0.12
E162	9/7/93	112	2.4	3.2	2.2	3.2	2.2	3.2	2.2	3.2	2.2	3.2
E163	9/14/93	3.7	0.614	0.13	0.626	0.13	0.626	0.13	0.626	0.13	0.626	0.13
E164	9/21/93	155	0.275	0.08	0.365	0.08	0.365	0.08	0.365	0.08	0.365	0.08
E165	9/28/93	114	0.047	0.02	0.150	0.02	0.150	0.02	0.150	0.02	0.150	0.02
E166		76	76	74	75	74	75	74	75	74	75	74
E167		12	0.085	0.01	0.035	0.01	0.035	0.01	0.035	0.01	0.035	0.01
E168		566	1.555	0.43	1.407	0.43	1.407	0.43	1.407	0.43	1.407	0.43
E169		118	0.565	0.12	0.617	0.12	0.617	0.12	0.617	0.12	0.617	0.12
E170		3.7	2.5	3.2	2.2	3.2	2.2	3.2	2.2	3.2	2.2	3.2

First 12 months statistics

ave	155
STD	115
avg error	13
count	46
min	12
max	566
median	112
max/mean	3.7

First 18 months statistics

ave	155
STD	114
avg error	12
count	76
min	12
max	566
median	118
max/mean	3.7

18 mos

Elms
Year 1

18 mos

Sample ID	Al	As	Cd	Cr	Cu	Ni	Pb	S
	ng/m ³ Sigma							
Year 2								
2nd year statistics								
ave	137	0.563	0.14	0.689	1.48	2.81	3.40	2484
STD	104	0.289	0.08	0.340	0.93	2.11	2.33	1469
avg error	9	0.015	0.03	0.070	0.20	0.16	0.46	300
count	51	52	50	52	51	45	50	52
min	28	0.187	0.01	0.060	0.09	0.06	0.18	847
max	532	1.903	0.44	1.576	4.10	9.10	11.60	6638
median	111	0.515	0.13	0.637	1.30	2.34	3.02	1863
max/mean	3.9	3.3	3.1	2.3	2.8	3.2	3.4	2.7
Year 3								
3rd year statistics								
ave	101	0.526	0.18	0.750	1.93	2.93	4.55	2103
STD	81	0.286	0.16	0.647	1.62	3.77	2.12	1392
avg error	20	0.025	0.01	0.127	0.38	0.28	0.81	167
count	33	47	45	45	43	41	45	47
min	3	0.016	0.01	0.050	0.15	0.22	0.50	205
max	295	1.291	0.85	3.977	7.20	17.16	9.30	5862
median	71	0.497	0.13	0.620	1.60	1.49	4.60	1733
max/mean	2.9	2.5	4.7	5.3	3.7	5.9	2.0	2.8

¹ Sampler malfunction, either no sample was collected or the air volume was inaccurate, values (when reported) are ng/filter, but not included in averages or flux estimates.

² Value not included in average.

ND = not detected.

So	Zn	V	Br	Fe	Mn
ng/m ³ Sigma	ng/m ³ Sigma	ng/m ³ Sigma	ng/m ³ Sigma	ng/m ³ Sigma	ng/m ³ Sigma
0.310 ± 0.039	4.48 ± 0.58	1.20 ± 0.10	0.699 ± 0.067	42.2 ± 3.4	0.65 ± 0.03
0.767 ± 0.093	4.04 ± 0.51	1.96 ± 0.17	1.07 ± 0.10	325 ± 14	2.36 ± 0.09
1.08 ± 0.091	35.9 ± 3.3	3.26 ± 0.28	1.45 ± 0.12	416 ± 10	2.99 ± 0.12
(15.4 ± 1.6) ¹	(335 ± 40) ¹	(66 ± 4.3) ¹	(0.409 ± 0.048) ¹	(105 ± 3.6) ¹	(1.15 ± 0.048) ¹
0.693 ± 0.075	3.59 ± 0.43	1.85 ± 0.16	1.17 ± 0.12	108 ± 4.9	1.76 ± 0.07
2.16 ± 0.21	15.6 ± 1.8	4.74 ± 0.40	2.26 ± 0.20	282 ± 12	3.06 ± 0.12
0.770 ± 0.092	34.0 ± 4.0	4.52 ± 0.38	1.53 ± 0.16	127 ± 7.2	2.67 ± 0.11
0.852 ± 0.096	9.13 ± 1.2	2.44 ± 0.21	1.88 ± 0.17	177 ± 8.3	2.40 ± 0.09
1.87 ± 0.20	18.2 ± 2.2	5.16 ± 0.44	3.15 ± 0.28	175 ± 9.1	3.20 ± 0.13
1.02 ± 0.11	9.72 ± 1.1	4.39 ± 0.37	1.99 ± 0.20	152 ± 7.6	2.25 ± 0.09
1.67 ± 0.25	9.02 ± 1.2	3.93 ± 0.33	3.53 ± 0.30	90.2 ± 11	2.45 ± 0.10
1.43 ± 0.28	1.63 ± 0.25	3.89 ± 0.33	2.81 ± 0.26	223 ± 16	4.30 ± 0.17
1.95 ± 0.30	7.87 ± 1.1	4.18 ± 0.38	3.80 ± 0.32	369 ± 21	4.15 ± 0.16
1.26 ± 0.22	6.94 ± 0.94	3.03 ± 0.28	2.26 ± 0.20	225 ± 16	3.22 ± 0.13
1.45 ± 0.22	6.35 ± 0.88	1.36 ± 0.12	2.40 ± 0.20	79.4 ± 8.1	2.41 ± 0.09
2.23 ± 0.36	16.9 ± 2.3	7.77 ± 0.66	4.52 ± 0.38	12.0 ± 12.0	4.93 ± 0.19
1.25 ± 0.22	7.47 ± 1.01	2.41 ± 0.21	2.38 ± 0.21	131 ± 11	2.78 ± 0.11
1.20 ± 0.23	3.81 ± 0.69	2.77 ± 0.24	2.02 ± 0.26	76.3 ± 12	1.67 ± 0.07
0.86 ± 0.09	7.89 ± 0.87	3.12 ± 0.26	2.69 ± 0.23	80.7 ± 4.0	2.50 ± 0.10
(6 ± 7.2) ¹	(780 ± 93) ¹	(154 ± 13) ¹	(165 ± 15) ¹	(5800 ± 370) ¹	(193 ± 7.6) ¹
2.60 ± 0.23	17.0 ± 2.0	4.81 ± 0.41	4.82 ± 0.39	122 ± 5.7	3.14 ± 0.12
1.91 ± 0.18	15.9 ± 1.8	5.81 ± 0.49	4.67 ± 0.38	189 ± 8.6	4.21 ± 0.16
2.36 ± 0.20	19.9 ± 1.9	2.67 ± 0.23	5.39 ± 0.44	109 ± 3.8	3.14 ± 0.12
3.22 ± 0.28	24.0 ± 2.3	6.29 ± 0.53	4.92 ± 0.40	115 ± 4.2	3.67 ± 0.14
1.98 ± 0.17	13.5 ± 1.3	3.31 ± 0.28	4.11 ± 0.33	72.9 ± 2.2	2.14 ± 0.08
0.876 ± 0.076	5.80 ± 0.57	3.44 ± 0.29	2.47 ± 0.20	30.2 ± 1.3	1.10 ± 0.04
2.01 ± 0.17	20.3 ± 1.9	5.44 ± 0.48	4.24 ± 0.34	81.7 ± 3.1	1.90 ± 0.07
3.71 ± 0.32	17.3 ± 1.7	6.75 ± 0.57	6.17 ± 0.50	89.7 ± 4.1	2.37 ± 0.09
2.61 ± 0.22	18.3 ± 1.7	4.69 ± 0.38	3.45 ± 0.23	82.3 ± 2.8	2.00 ± 0.04
2.74 ± 0.23	22.9 ± 2.1	6.34 ± 0.54	5.56 ± 0.46	103 ± 3.2	3.28 ± 0.13
1.82 ± 0.15	12.0 ± 1.2	3.33 ± 0.28	4.64 ± 0.39	72.9 ± 2.6	2.34 ± 0.09
3.80 ± 0.32	20.5 ± 1.9	8.35 ± 0.68	4.85 ± 0.33	142 ± 4.5	4.58 ± 0.10
1.70 ± 0.15	8.99 ± 0.95	2.80 ± 0.24	3.54 ± 0.29	94.0 ± 3.9	1.93 ± 0.08
1.76 ± 0.16	13.1 ± 1.3	4.36 ± 0.37	3.14 ± 0.27	114 ± 4.8	2.01 ± 0.08
1.69 ± 0.14	14.6 ± 1.4	2.65 ± 0.21	2.49 ± 0.17	63.1 ± 2.4	1.76 ± 0.04
1.74 ± 0.15	12.6 ± 1.2	4.78 ± 0.40	3.78 ± 0.31	110 ± 3.3	3.02 ± 0.12
2.11 ± 0.18	7.40 ± 0.77	5.10 ± 0.43	3.28 ± 0.28	52.7 ± 2.2	1.56 ± 0.06
1.50 ± 0.13	17.4 ± 1.7	4.67 ± 0.38	2.57 ± 0.18	115 ± 3.7	2.82 ± 0.06
1.20 ± 0.11	13.2 ± 1.3	3.48 ± 0.29	3.40 ± 0.28	102 ± 3.8	2.49 ± 0.10
1.17 ± 0.10	9.31 ± 0.94	2.11 ± 0.18	2.85 ± 0.24	146 ± 5.5	2.12 ± 0.08
0.700 ± 0.063	5.17 ± 0.58	2.71 ± 0.22	2.67 ± 0.19	114 ± 3.6	2.39 ± 0.05
0.907 ± 0.077	8.53 ± 0.84	2.42 ± 0.21	2.63 ± 0.23	63.8 ± 1.8	1.30 ± 0.05
1.11 ± 0.094	9.08 ± 0.90	3.96 ± 0.34	3.63 ± 0.31	111 ± 3.3	2.51 ± 0.10
0.562 ± 0.050	3.25 ± 0.41	1.10 ± 0.09	1.05 ± 0.07	13.9 ± 1.3	0.55 ± 0.01
1.94 ± 0.17	7.69 ± 0.84	4.34 ± 0.37	4.39 ± 0.35	27.9 ± 2.4	1.08 ± 0.04
0.887 ± 0.085	5.78 ± 0.66	4.17 ± 0.35	4.10 ± 0.37	141 ± 5.6	3.04 ± 0.12
1.60 ± 0.14	7.96 ± 0.85	4.41 ± 0.36	6.07 ± 0.47	465 ± 12	12.65 ± 0.26
2.30 ± 0.19	15.7 ± 1.5	4.17 ± 0.35	3.02 ± 0.26	332 ± 8.6	4.05 ± 0.16
1.17 ± 0.10	7.13 ± 0.74	3.19 ± 0.27	2.42 ± 0.23	183 ± 5.1	3.34 ± 0.13
0.708 ± 0.066	4.67 ± 0.57	3.22 ± 0.26	3.25 ± 0.23	174 ± 5.0	4.48 ± 0.09
1.45 ± 0.13	9.82 ± 1.0	5.30 ± 0.45	2.86 ± 0.24	83.1 ± 3.7	1.84 ± 0.07
1.63 ± 0.15	5.92 ± 0.81	4.45 ± 0.38	3.60 ± 0.33	400 ± 15	2.82 ± 0.11
1.58 ± 0.13	6.65 ± 0.88	4.17 ± 0.35	2.58 ± 0.22	172 ± 4.8	2.51 ± 0.10
1.95 ± 0.16	10.6 ± 1.0	6.75 ± 0.57	3.80 ± 0.33	102 ± 3.4	2.92 ± 0.12
2.32 ± 0.14	8.45 ± 0.127	3.682 ± 0.89	2.177 ± 0.231	447 ± 10.2	4.265 ± 0.170
1.53 ± 0.09	9.21 ± 0.694	3.443 ± 0.76	2.358 ± 0.186	138 ± 4.1	2.713 ± 0.058

(Elms site)

Se	Zn	V	Br	Fe	Mn
ng/m ³ Sigma					
1.82 ± 0.09	7.93 ± 0.102	2.126 ± 0.64	1.828 ± 0.127	117 ± 3.4	1.608 ± 0.066
1.46 ± 0.09	11.99 ± 0.094	3.193 ± 0.90	2.539 ± 0.171	283 ± 6.3	2.735 ± 0.057
2.20 ± 0.13	22.27 ± 0.137	3.560 ± 1.58	2.182 ± 0.212	506.7 ± 11.1	4.436 ± 0.176
2.15 ± 0.13	11.90 ± 0.143	4.176 ± 0.95	3.086 ± 0.226	180 ± 5.1	2.364 ± 0.052
1.31 ± 0.08	8.26 ± 0.155	4.328 ± 0.70	2.238 ± 0.258	249 ± 5.9	2.312 ± 0.094
1.81 ± 0.11	27.10 ± 0.118	4.721 ± 1.88	3.176 ± 0.253	158 ± 4.3	2.876 ± 0.061
1.98 ± 0.12	13.55 ± 0.147	4.681 ± 1.04	2.449 ± 0.278	183 ± 5.0	2.686 ± 0.108
1.60 ± 0.10	11.98 ± 0.086	2.242 ± 0.91	2.103 ± 0.120	151 ± 4.1	2.698 ± 0.057
1.84 ± 0.11	13.26 ± 0.03	3.271 ± 0.96	2.3 ± 0.194	118 ± 3.4	(26.23) ± (1.025)
1.26 ± 0.08	8.64 ± 0.02	3.601 ± 0.70	2.3 ± 0.182	64 ± 3.0	(14.76) ± (0.298)
1.28 ± 0.08	7.81 ± 0.113	3.327 ± 0.63	2.061 ± 0.198	68 ± 2.9	1.362 ± 0.057
1.29 ± 0.08	10.92 ± 0.081	2.072 ± 0.83	2.301 ± 0.112	75 ± 2.9	2.274 ± 0.051
2.82 ± 0.17	17.30 ± 0.250	13.100 ± 1.41	4.593 ± 0.781	158 ± 6.4	2.244 ± 0.089
2.24 ± 0.13	21.95 ± 0.151	5.678 ± 1.52	4.765 ± 0.305	195 ± 4.6	4.246 ± 0.091
2.34 ± 0.14	21.52 ± 0.286	10.886 ± 1.54	5.085 ± 0.651	222 ± 5.8	4.134 ± 0.166
3.01 ± 0.18	16.97 ± 0.139	9.897 ± 1.23	4.432 ± 0.530	113 ± 3.8	2.695 ± 0.058
1.42 ± 0.08	10.38 ± 0.156	5.474 ± 0.77	2.627 ± 0.327	59 ± 2.5	1.187 ± 0.050
1.48 ± 0.09	11.76 ± 0.092	6.058 ± 0.85	2.844 ± 0.326	67 ± 2.4	1.558 ± 0.035
2.68 ± 0.15	14.52 ± 0.205	4.298 ± 1.05	3.684 ± 0.258	72 ± 2.7	2.009 ± 0.083
2.23 ± 0.13	13.56 ± 0.115	7.631 ± 0.97	3.670 ± 0.410	70 ± 2.5	2.075 ± 0.046
3.34 ± 0.19	20.96 ± 0.262	5.475 ± 1.46	4.813 ± 0.328	156 ± 3.8	2.791 ± 0.113
1.94 ± 0.11	17.83 ± 0.096	6.700 ± 1.23	3.152 ± 0.361	68 ± 2.1	1.356 ± 0.031
2.39 ± 0.14	19.39 ± 1.36	6.485 ± 0.386	4.207 ± 0.235	87 ± 3	2.329 ± 0.094
2.45 ± 0.15	22.37 ± 1.59	4.134 ± 0.221	4.062 ± 0.130	267 ± 7	2.837 ± 0.061
1.77 ± 0.10	11.61 ± 0.85	2.772 ± 0.168	2.331 ± 0.128	123 ± 3	1.951 ± 0.080
2.27 ± 0.13	18.51 ± 1.29	6.804 ± 0.370	4.423 ± 0.140	76 ± 3	2.012 ± 0.048
3.96 ± 0.22	28.95 ± 1.95	8.12 ± 0.14	5.680 ± 0.028	119.68 ± 4.13	2.6 ± 0.3
1.31 ± 0.07	13.73 ± 0.93	6.51 ± 0.12	0.553 ± 0.004	81.60 ± 3.18	1.6 ± 0.2
2.06 ± 0.12	17.91 ± 1.21	8.95 ± 0.16	3.379 ± 0.014	73.03 ± 3.39	2.1 ± 0.2
1.97 ± 0.11	15.17 ± 1.04	8.95 ± 0.16	1.805 ± 0.008	42.74 ± 3.25	1.2 ± 0.2
1.61 ± 0.09	13.11 ± 0.89	6.20 ± 0.12	2.479 ± 0.013	61.88 ± 3.16	1.5 ± 0.2
1.18 ± 0.07	15.05 ± 1.03	5.84 ± 0.11	2.525 ± 0.021	35.54 ± 3.35	0.6 ± 0.2
1.57 ± 0.09	11.73 ± 0.81	4.62 ± 0.09	3.151 ± 0.013	105.07 ± 3.85	2.3 ± 0.3
2.14 ± 0.12	21.35 ± 1.48	7.74 ± 0.14	2.486 ± 0.009	111.49 ± 4.44	3.3 ± 0.3
1.39 ± 0.08	16.97 ± 1.15	6.29 ± 0.12	0.779 ± 0.004	113.21 ± 3.99	2.3 ± 0.3
2.17 ± 0.12	17.28 ± 1.17	5.34 ± 0.10	4.066 ± 0.020	149.91 ± 4.46	3.1 ± 0.3
1.87 ± 0.11	16.16 ± 1.11	6.44 ± 0.12	2.125 ± 0.018	235.97 ± 5.99	2.6 ± 0.3
0.89 ± 0.05	9.43 ± 0.66	5.54 ± 0.11	1.878 ± 0.017	103.24 ± 3.99	1 ± 0.2
1.15 ± 0.07	7.43 ± 0.55	2.96 ± 0.08	1.297 ± 0.010	69.53 ± 3.91	1.6 ± 0.2
1.63 ± 0.09	13.98 ± 0.96	3.91 ± 0.08	< 0.004 ± 0.004	240.22 ± 6.00	4.6 ± 0.4
1.61 ± 0.09	17.31 ± 1.18	9.76 ± 0.18	0.757 ± 0.005	187.89 ± 5.54	2.4 ± 0.3
1.47 ± 0.08	14.60 ± 1.02	3.33 ± 0.08	2.752 ± 0.017	252.32 ± 7.02	3.2 ± 0.3
1.08 ± 0.06	8.16 ± 0.57	2.52 ± 0.05	1.980 ± 0.014	67.84 ± 3.53	1.4 ± 0.2
1.11 ± 0.06	7.76 ± 0.55	4.40 ± 0.08	< 0.001 ± 0.001	62.99 ± 3.57	0.6 ± 0.2
1.28 ± 0.07	14.59 ± 1.00	7.40 ± 0.14	2.905 ± 0.019	183.60 ± 5.13	2.000 ± 0.200
0.49 ± 0.03	5.91 ± 0.42	1.36 ± 0.03	0.263 ± 0.004	67.32 ± 3.27	0.900 ± 0.200
2.12 ± 0.12	17.01 ± 1.16	4.80 ± 0.09	< 0.005 ± 0.005	91.46 ± 3.96	2.100 ± 0.200
1.71 ± 0.10	9.92 ± 0.70	2.74 ± 0.05	1.475 ± 0.011	107.85 ± 4.08	1.600 ± 0.200
2.20 ± 0.13	8.37 ± 0.65	5.11 ± 0.24	1.351 ± 0.014	(826.1 ± 17.85)	11.800 ± 1.000
2.49 ± 0.18	11.4 ± 1.6	5 ± 1.2	3.82 ± 0.12	(2018 ± 51.6)	(27 ± 2.2)
2.42 ± 0.15	10.2 ± 0.8	2.2 ± 0.5	4.19 ± 0.09	130 ± 6	3.1 ± 0.4
1.15 ± 0.08	14.5 ± 1.1	1.7 ± 0.4	2.26 ± 0.07	62 ± 4	2.1 ± 0.3
1.07 ± 0.07	3.3 ± 0.4	3.4 ± 0.5	4.02 ± 0.08	130 ± 3	0.8 ± 0.2
2.18 ± 0.13	6.4 ± 0.6	2.8 ± 0.5	3.27 ± 0.23	305 ± 16	1.5 ± 0.3
1.08 ± 0.14	14.7 ± 1.7	ND	1.16 ± 0.27	15 ± 15	4.9 ± 1
ND	< 0.8 ± 0.8	ND	2.21 ± 0.25	67 ± 13	ND
0.43 ± 0.09	3.1 ± 0.9	ND	1.62 ± 0.07	172 ± 6	3.6 ± 0.4
0.99 ± 0.09	(153.3 ± 10.32)	ND	0.78 ± 0.07	35 ± 4	1 ± 0.3
2.12 ± 0.13	4.2 ± 0.4	2.1 ± 0.5	3.94 ± 0.10	145 ± 6	3.1 ± 0.4
0.91 ± 0.06	12.7 ± 1.0	5.61 ± 0.11	2.54 ± 0.09	100 ± 5	1.8 ± 0.3
1.41 ± 0.10	8.6 ± 0.7	3.9 ± 0.6	3.72 ± 0.13	75 ± 6	2.5 ± 0.4
0.71 ± 0.06	21.5 ± 1.6	5.5 ± 0.8	2.54 ± 0.09	131 ± 7	1.7 ± 0.3
2.46 ± 0.10	34.2 ± 2.5	6.7 ± 0.9	3.89 ± 0.10	37 ± 5	2.4 ± 0.4
0.86 ± 0.06	12.2 ± 1.0	4.1 ± 0.7	3.21 ± 0.13	131 ± 7	ND
	7.3 ± 0.6				

Se ng/m³ Sigma	Zn ng/m³ Sigma	V ng/m³ Sigma	Br ng/m³ Sigma	Fe ng/m³ Sigma	Mn ng/m³ Sigma
3.29	19.8	1.9	0.6	119	6
0.75	1.2	0.15	0.15	108	8
0.92	7.4	0.6	0.07	41	4
1.31	0.6	0.9	0.08	21	3
1.42	0.5	0.6	0.07	80	5
0.91	4.3	0.8	0.07	32	4
0.75	0.4	0.5	0.07	52	5
1.64	0.4	1.1	0.08	70	5
2.20	3.8	0.7	0.09	83	5
0.33	2.6	0.7	0.07	93	5
0.69	2.3	0.7	0.09	23	3
0.54	4.4	0.8	0.06	137	5
0.56	4.4	0.8	0.07	51	4
2.49	2.9	0.8	0.08	277	8
0.80	2.4	0.8	0.08	28	4
1.76	4.9	0.6	0.08	266	8
1.40	0.9	0.6	0.07	73	5
1.09	8.4	0.7	0.08	164	6
0.88	10.3	0.8	0.07	76	4
4.52	5.2	0.5	0.08	116	5
2.28	17.4	1.3	0.09	216	7
1.93	21.5	1.5	0.11	232	8
2.92	9.3	0.9	0.07	161	6
	5.2	0.6	0.08	202	8
	17.8	1.3	0.07		
		6.5	0.9		
		2.1			
		3.4			
		1.9			
		3.064			
		465			
		12			
		465			
		110			
		2.430			
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		3.250			
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		7			
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		0.551			
		12.651			
		2.430			
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		2.802			
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		12.651			
		2.430			
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		138			
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		46			
		0.551			
		12.651			
		2.430			
		4.5			
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		12.651			
		2.430			
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		46			
		0.551			
		12.651			
		2.430			
		4.5			
		3.4			
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		0.551			</

Se ng/m ³ Sigma	Zn ng/m ³ Sigma	V ng/m ³ Sigma	Br ng/m ³ Sigma	Fe ng/m ³ Sigma	Mn ng/m ³ Sigma
18	13.9	5.3	2.769	147	2.433
0.6	5.4	2.3	1.227	99	0.965
0.1	0.6	0.6	0.182	5	0.146
52	52	51	52	-52	50
0.7	4.7	2.1	0.001	36	0.600
4.0	28.9	13.1	5.680	507	4.600
1.6	13.4	4.7	2.532	117	2.320
2.2	2.1	2.5	2.1	3.4	1.9
1.5	14.3	4.1	2.847	104	2.611
0.9	12.3	2.6	1.690	71	2.006
0.1	1.2	0.6	0.111	8	0.385
46	47	45	48	47	42
0.2	0.8	0.2	0.005	13	0.250
4.5	55.8	13.2	8.603	305	11.800
1.3	10.0	3.8	2.854	83	2.100
3.1	3.9	3.2	3.0	2.9	4.5

Sample ID

Appendix E-3 Concentrations of Elements in Chesapeake Bay Aerosol
(Haven Beach Site)

Sample ID

Sample ID	Year	AJ	As	Cd	Cr	Cu	Ni	Pb	S
ng/m ³ Sigma									
V2	11/13/90	172 ± 15	1.28 ± 0.062	0.187 ± 0.019	1.17 ± 0.36	3.20 ± 0.31	(12.95 ± 0.94)	5.86 ± 0.49	ND
V3	11/20/90	158 ± 14	1.45 ± 0.089	0.424 ± 0.033	1.58 ± 0.35	3.35 ± 0.31	(13.35 ± 0.94)	11.0 ± 1.3	2584 ± 590
V4	11/27/90	(223 ± 19)	(3.97 ± 0.24)	(0.203 ± 0.046)	(9.77 ± 1.25)	(17.23 ± 1.30)	(39.7 ± 2.4)	(12.7 ± 2.0)	ND
V5	12/11/90	80.9 ± 7.2	0.827 ± 0.052	0.149 ± 0.020	1.22 ± 0.44	(7.21 ± 0.535)	(8.13 ± 0.52)	5.10 ± 0.87	2513 ± 483
V6	12/18/90	37.8 ± 5.7	0.285 ± 0.026	< 0.017 ± 0.017	< 0.143 ± 0.143	(5.98 ± 0.61)	(91.1 ± 5.6)	ND	1765 ± 509
V7	12/25/90	77.3 ± 8.9	0.341 ± 0.035	< 0.179 ± 0.039	< 0.88 ± 0.88	2.74 ± 0.65	< 0.377 ± 0.377	5.22 ± 0.95	1446 ± 240
V8	1/01/91	67.1 ± 6.1	0.891 ± 0.057	0.261 ± 0.032	1.90 ± 0.69	2.45 ± 0.49	1.61 ± 0.98	(12.06 ± 1.12)	ND
V9	1/08/91								
V10	1/15/91								
V11	1/22/91								
V12	2/05/91	10.8 ± 2.0	0.040 ± 0.003	< 0.007 ± 0.007	< 0.57 ± 0.57	< 0.37	< 0.136	< 0.3	ND
V13	2/12/91	(215 ± 19)	(1.43 ± 0.094)	(0.424 ± 0.052)	(2.61 ± 1.1)	(6.74 ± 0.86)	(1.42 ± 0.68)	(7.65 ± 1.26)	ND
V14	2/19/91	102 ± 12	0.526 ± 0.036	0.257 ± 0.051	< 0.29 ± 0.29	(7.09 ± 0.92)	3.81 ± 0.79	< 1.16 ± 1.16	ND
V15	2/19/91								
V16	3/19/91	83.3 ± 7.4	0.645 ± 0.040	0.071 ± 0.029	< 0.16 ± 0.16	(11.59 ± 0.83)	(25.4 ± 2.1)	5.05 ± 0.76	ND
V17	3/26/91	(164 ± 14)	(0.86 ± 0.06)	(0.383 ± 0.05)	(0.957 ± 1.06)	(12.8 ± 1.1)	(7.01 ± 0.62)	(6.72 ± 1.1)	(3440 ± 621)
V18	4/02/91	347 ± 30	0.548 ± 0.036	< 0.036	< 0.23 ± 0.23	< 0.74	< 0.6	< 0.1	3160 ± 882
V19	4/09/91	127 ± 11	0.692 ± 0.044	0.114 ± 0.021	0.261 ± 0.261	4.08 ± 0.31	(8.84 ± 0.567)	4.08 ± 0.47	2348 ± 331
V20	4/16/91	193 ± 12	0.543 ± 0.037	0.214 ± 0.051	0.164 ± 0.164	3.05 ± 0.36	2.21 ± 0.37	2.97 ± 1.1	ND
V21	4/23/91	136 ± 12	0.581 ± 0.042	0.128 ± 0.021	< 0.020 ± 0.020	(8.68 ± 0.97)	1.26 ± 1.26	ND	ND
V22	4/30/91	93.6 ± 8.2	0.560 ± 0.039	0.098 ± 0.021	0.101 ± 0.101	2.04 ± 0.36	3.88 ± 0.40	4.26 ± 0.59	4564 ± 760
V23	5/07/91	157 ± 14	0.630 ± 0.039	0.098 ± 0.021	0.211 ± 0.211	(8.07 ± 0.59)	(3.33 ± 0.29)	1.22 ± 0.59	3723 ± 575
V24	5/14/91	136 ± 12	0.548 ± 0.035	0.130 ± 0.021	< 0.268 ± 0.268	4.64 ± 0.43	(8.79 ± 0.769)	4.72 ± 0.54	2085 ± 486
V25	5/21/91	67.9 ± 5.9	0.371 ± 0.034	0.087 ± 0.012	0.161 ± 0.161	2.68 ± 0.24	2.76 ± 0.28	2.44 ± 0.51	ND
V26	5/28/91	540 ± 46	0.355 ± 0.028	0.131 ± 0.016	0.827 ± 0.26	0.836 ± 0.18	2.67 ± 0.30	0.85 ± 0.33	(7369 ± 961)
V27	6/04/91	325 ± 28	0.840 ± 0.052	0.074 ± 0.017	1.01 ± 0.40	(10.6 ± 0.69)	3.35 ± 0.50	1.81 ± 0.52	ND
V28	6/11/91	(435 ± 37)	(0.926 ± 0.059)	(0.241 ± 0.046)	(1.39 ± 0.96)	(8.67 ± 0.85)	(9.07 ± 1.02)	(3.96 ± 1.15)	(7352 ± 1096)
V29	6/18/91								
V30	6/25/91								
V31	8/07/91	270.0 ± 16.8	0.442 ± 0.008	< 0.093 ± 0.023	0.92 ± 0.05	1.723 ± 0.217	0.910 ± 0.268	0.982 ± 0.184	4097 ± 510
V32	8/13/91	59.2 ± 9.0	0.27 ± 0.01	< 0.0792 ± 0.0792	0.27 ± 0.03	0.82 ± 0.82	3.86 ± 1.37	5.82 ± 1.64	2568 ± 343
V33	8/20/91	55.5 ± 3.9	0.03 ± 0.01	< 0.0192 ± 0.0192	0.20 ± 0.04	2.01 ± 0.24	6.23 ± 0.44	0.98 ± 0.31	4681 ± 529
V34	8/27/91	64.4 ± 4.1	0.42 ± 0.02	< 0.0078 ± 0.0078	0.78 ± 0.04	0.74 ± 0.10	2.374 ± 0.24	1.29 ± 0.20	3080 ± 556
V35	9/10/91	67.0 ± 4.2	0.19 ± 0.00	0.0881 ± 0.0112	0.31 ± 0.02	0.99 ± 0.10	1.38 ± 0.12	1.58 ± 0.28	1798 ± 243
V36	9/17/91	90.3 ± 6.3	0.20 ± 0.02	< 0.014 ± 0.014	0.35 ± 0.06	< 0.216 ± 0.216	2.246 ± 0.390	< 0.372 ± 0.372	27 ± 27
V37	9/24/91								
V38	10/01/91	43.0 ± 3.7	0.15 ± 0.01	(1.05 ± 0.238)	0.178 ± 0.036	0.87 ± 0.87	(10.8) ± (2.94)	0.68 ± 0.68	3072 ± 519
V39	10/08/91	48.3 ± 3.4	0.28 ± 0.01	0.1904 ± 0.0306	0.28 ± 0.04	1.24 ± 0.18	1.65 ± 0.29	2.85 ± 0.22	1448 ± 209
V40	10/15/91	118.9 ± 7.6	0.43 ± 0.02	< 0.0594 ± 0.0594	0.39 ± 0.06	2.95 ± 0.23	5.83 ± 0.27	4.03 ± 1.35	1365 ± 199
V41	10/22/91	98.4 ± 6.4	0.58 ± 0.04	< 0.1145 ± 0.1145	0.78 ± 0.04	1.27 ± 0.20	2.374 ± 0.24	0.12 ± 0.12	17 ± 17
V42	10/29/91	94.7 ± 6.6	0.95 ± 0.04	< 0.1310 ± 0.1310	0.88 ± 0.06	1.01 ± 0.23	2.087 ± 0.27	< 1.29 ± 1.29	43 ± 43
V43	11/05/91	119.5 ± 7.8	0.78 ± 0.01	< 0.0012 ± 0.0012	0.47 ± 0.06	0.17 ± 0.17	0.27 ± 0.27	1.10 ± 0.36	2336 ± 365
V44	11/12/91	168.5 ± 10.8	1.02 ± 0.04	0.1732 ± 0.0199	0.69 ± 0.06	3.56 ± 0.26	(8.464) ± (0.3)	1.71 ± 0.72	33 ± 33
V45	11/19/91	65.7 ± 4.4	< 0.0001 ± 0.0001	0.0685 ± 0.0665	0.32 ± 0.04	4.04 ± 0.19	(8.467) ± (0.34)	0.28 ± 0.28	781 ± 241
V46	12/03/91	114.2 ± 7.5	0.02 ± 0.001	(0.605) ± (0.0429)	(17.1) ± (0.249)	(23.4) ± (0.387)	(5.728 ± 0.23)	5.17 ± 0.51	1578 ± 484
V47	12/10/91	179 ± 1.8	0.72 ± 0.00	< 0.0972 ± 0.0972	0.65 ± 0.04	4.25 ± 0.20	5.39 ± 0.23	2.53 ± 0.27	7 ± 7
V48	1/07/92	41.6 ± 3.0	0.82 ± 0.01	< 0.1478 ± 0.1478	0.72 ± 0.05	0.64 ± 0.17	0.72 ± 0.20	0.63 ± 0.63	3063 ± 440
V49	1/14/92	67.9 ± 4.5	ND	0.0822 ± 0.0277	0.30 ± 0.03	1.06 ± 0.17	0.95 ± 0.21	1.95 ± 0.62	1132 ± 175
V50	1/21/92	39.5 ± 3.0	ND	(0.741) ± (0.602)	0.49 ± 0.04	ND	0.63 ± 0.63	1.69 ± 0.28	1567 ± 217
V51	1/28/92	33.7 ± 2.4	0.48 ± 0.01	0.1071 ± 0.0440	0.25 ± 0.03	1.73 ± 0.14	5.63 ± 0.17	1.99 ± 0.28	1799 ± 288
V52	2/04/92	11.13 ± 0.62	0.577 ± 0.014	0.121 ± 0.006	0.481 ± 0.062	1.2 ± 0.2	5.16 ± 0.11	5 ± 0.6	1351 ± 98
V53	2/11/92	60.53 ± 2.95	0.259 ± 0.005	0.219 ± 0.006	0.317 ± 0.079	1.8 ± 0.3	3.35 ± 0.09	4.1 ± 0.6	1381 ± 107
V54	2/18/92	21.81 ± 1.18	0.057 ± 0.005	0.132 ± 0.006	0.246 ± 0.076	0.6 ± 0.2	3.46 ± 0.20	3.6 ± 0.5	2114 ± 165
V55	3/10/92	77.95 ± 3.75	0.546 ± 0.012	0.146 ± 0.008	0.426 ± 0.078	0.9 ± 0.2	4.16 ± 0.08	4.4 ± 0.6	1240 ± 98
V56	3/17/92	74.78 ± 3.50	0.850 ± 0.008	0.158 ± 0.002	0.718 ± 0.040	1.7 ± 0.2	2.73 ± 0.03	6.4 ± 0.7	1519 ± 120
V57	3/24/92	93.86 ± 4.48	0.672 ± 0.010	0.197 ± 0.003	0.543 ± 0.056	4.7 ± 0.5	4.88 ± 0.08	5.9 ± 0.7	1581 ± 125
V58	4/07/92	144.60 ± 6.74	0.680 ± 0.028	0.228 ± 0.004	0.624 ± 0.059	1.8 ± 0.2	3.38 ± 0.04	7.1 ± 0.8	2240 ± 177

Sample ID

Sample ID	Year	AI	As	Cd	Cr	Cu	Ni	Pb	S
		ng/m ³ Sigma							
V62	4/14/92	90.59 ± 4.28	0.409 ± 0.026	0.106 ± 0.007	0.608 ± 0.060	0.8 ± 0.2	4.18 ± 0.07	3 ± 0.4	1498 ± 118
V63	4/21/92	64.58 ± 3.14	0.335 ± 0.028	0.086 ± 0.005	0.261 ± 0.061	2.3 ± 0.3	1.96 ± 0.21	2.7 ± 0.4	1453 ± 115
V64	4/28/92	218.22 ± 10.11	0.694 ± 0.016	0.206 ± 0.009	0.570 ± 0.063	1.5 ± 0.2	0.570 ± 0.063	4.7 ± 0.6	1683 ± 133
V66	5/12/92	63.59 ± 3.04	0.438 ± 0.011	0.115 ± 0.007	0.452 ± 0.061	1.9 ± 0.3	3.97 ± 0.07	3.8 ± 0.5	2199 ± 173
V67	5/19/92	161.27 ± 7.57	0.546 ± 0.018	0.111 ± 0.005	0.569 ± 0.062	0.9 ± 0.2	1.93 ± 0.08	4.8 ± 0.6	1816 ± 143
V68	5/26/92	40.73 ± 1.97	0.351 ± 0.019	0.092 ± 0.011	0.226 ± 0.061	0.4 ± 0.1	3.7 ± 0.05	2.6 ± 0.4	1625 ± 128
V69	6/02/92	61.56 ± 3.01	0.108 ± 0.009	0.107 ± 0.004	0.434 ± 0.060	0.7 ± 0.2	2.5 ± 0.07	2.5 ± 0.4	1910 ± 151
V70	6/09/92	121.49 ± 5.71	0.579 ± 0.009	ND ±	0.435 ± 0.060	1.10 ± 0.20	ND ±	3.70 ± 0.50	1805 ± 142
V71	6/16/92	74.61 ± 3.56	0.402 ± 0.014	0.44 ± 0.006	0.440 ± 0.064	1.40 ± 0.20	13.34 ± 0.27	2.90 ± 0.40	1511 ± 119
V72	6/24/92	69.72 ± 3.33	0.490 ± 0.022	0.40 ± 0.011	0.312 ± 0.061	1.20 ± 0.20	10.65 ± 0.12	3.30 ± 0.40	2577 ± 203
V73	6/30/92	85.38 ± 4.02	0.685 ± 0.018	0.22 ± 0.004	0.416 ± 0.062	1.20 ± 0.20	4.64 ± 0.12	3.60 ± 0.50	2088 ± 165
V74	7/07/92	(1679.4 ± 78.0)	0.882 ± 0.015	0.42 ± 0.013	(2.59 ± 0.085)	2.50 ± 0.30	8.91 ± 0.11	4.60 ± 0.60	3440 ± 271
V75	7/14/92	(3568.2 ± 950.1)	0.96 ± 0.05	0.144 ± 0.004	(3.863 ± 0.185)	2.7 ± 0.4	3.58 ± 0.32	4.5 ± 0.7	3527 ± 278
V76	7/21/92		0.54 ± 0.03	0.100 ± 0.005	0.70 ± 0.14	3.1 ± 0.4	2.21 ± 0.31	2.5 ± 0.5	2590 ± 200
V77	7/28/92		0.52 ± 0.02	0.139 ± 0.003	0.59 ± 0.13	5.5 ± 1.3	1.81 ± 0.19	4.6 ± 0.7	1692 ± 134
V78	8/04/92		0.60 ± 0.05	0.096 ± 0.029	(2.925 ± 0.607)	3.1 ± 0.6	(24.13 ± 1.08)	6.5 ± 1.2	3497 ± 277
V79	8/11/92		0.79 ± 0.04	0.128 ± 0.024	0.63 ± 0.28	2.3 ± 0.4	5.60 ± 0.77	6.5 ± 1.2	5174 ± 408
V80	8/18/92		0.19 ± 0.01	0.024 ± 0.004	<	2.1 ± 0.4	1.71 ± 0.19	2.2 ± 0.5	1045 ± 83
V81	8/25/92		0.88 ± 0.03	0.126 ± 0.004	0.13 ± 0.13	1.5 ± 0.2	2.61 ± 0.20	3.4 ± 0.6	2310 ± 182
V82	9/01/92		0.31 ± 0.01	0.169 ± 0.003	0.81 ± 0.15	2.3 ± 0.4	3.66 ± 0.25	5.6 ± 0.8	3127 ± 247
V83	9/15/92		0.53 ± 0.03	0.072 ± 0.002	<	2.1 ± 0.4	1.60 ± 0.15	2.2 ± 0.4	918 ± 73
V84	9/22/92		0.36 ± 0.05	0.047 ± 0.009	0.10 ± 0.10	2.1 ± 0.4	0.08 ± 0.08	1.7 ± 0.5	751 ± 60
V85	9/29/92		0.32 ± 0.01	0.087 ± 0.008	0.29 ± 0.14	1.3 ± 0.3	0.59 ± 0.21	3.6 ± 0.6	626 ± 50
V86	10/06/92		0.27 ± 0.04	0.808 ± 0.023	0.08 ± 0.08	1.5 ± 0.3	1.78 ± 0.20	2.8 ± 0.6	542 ± 43
V87	10/13/92		1.15 ± 0.02	0.247 ± 0.002	0.24 ± 0.07	1.5 ± 0.2	1.75 ± 0.10	6.9 ± 0.8	1928 ± 152
V88	10/20/92		0.62 ± 0.02	0.161 ± 0.001	0.33 ± 0.06	0.8 ± 0.1	0.91 ± 0.09	3.2 ± 0.4	844 ± 67
V89	10/27/92		0.52 ± 0.03	0.486 ± 0.016	0.47 ± 0.15	1.9 ± 0.4	1.24 ± 0.22	5.1 ± 0.7	1048 ± 83
V90	11/03/92		0.30 ± 0.02	0.140 ± 0.007	0.32 ± 0.14	2.2 ± 0.4	3.27 ± 0.23	3 ± 0.6	1039 ± 82
V91	11/10/92		0.94 ± 0.04	0.268 ± 0.005	1.16 ± 0.11	2 ± 0.3	4.68 ± 0.15	5 ± 0.6	1104 ± 87
V92	11/24/92		0.26 ± 0.03	0.187 ± 0.003	0.25 ± 0.12	0.8 ± 0.2	1.48 ± 0.21	1.7 ± 0.5	810 ± 64
V93	12/01/92		1.83 ± 0.05	0.048 ± 0.004	0.37 ± 0.14	3 ± 0.4	0.14 ± 0.14	7.9 ± 1	1508 ± 119
V94	12/08/92		0.74 ± 0.05	0.107 ± 0.004	1.08 ± 0.17	2.4 ± 0.4	0.60 ± 0.19	11.6 ± 1.4	2182 ± 172
V95	12/15/92		0.93 ± 0.03	0.350 ± 0.008	0.09 ± 0.09	1.6 ± 0.3	0.29 ± 0.19	3.8 ± 0.6	2177 ± 172
V96	12/22/92		0.59 ± 0.02	0.208 ± 0.010	0.84 ± 0.14	2.8 ± 0.4	2.91 ± 0.19	6.9 ± 0.9	972 ± 77
V97	12/29/92		0.45 ± 0.02	0.119 ± 0.005	0.08 ± 0.08	2.7 ± 0.4	1.69 ± 0.21	4.5 ± 0.7	1328 ± 105
V98	1/05/93		0.04 ± 0.00	0.045 ± 0.004	0.41 ± 0.13	1.4 ± 0.3	1.89 ± 0.20	4.5 ± 0.7	707 ± 56
V99	1/12/93		0.00 ± 0.00	0.045 ± 0.004	0.07 ± 0.07	3.4 ± 0.4	1.71 ± 0.12	1.0 ± 0.3	124 ± 10
V100	1/19/93		0.57 ± 0.03	0.192 ± 0.006	0.78 ± 0.10	1.5 ± 0.3	1.64 ± 0.14	4.8 ± 0.6	752 ± 60
V101	1/26/93		0.46 ± 0.02	0.089 ± 0.003	0.47 ± 0.09	2 ± 0.4	1.39 ± 0.15	3.7 ± 0.6	456 ± 33.1
V102	2/02/93		0.71 ± 0.02	0.167 ± 0.005	0.66 ± 0.10	1.3 ± 0.3	3.37 ± 0.16	3.7 ± 0.6	578.2 ± 41.4
V103	2/09/93		0.75 ± 0.04	0.146 ± 0.003	0.35 ± 0.11	1.6 ± 0.4	3.09 ± 0.15	4.9 ± 0.8	695.2 ± 50.6
V104	2/16/93		0.37 ± 0.02	0.128 ± 0.005	0.20 ± 0.09	0.9 ± 0.3	1.25 ± 0.13	3 ± 0.6	1573.4 ± 113
V105	2/23/93		0.64 ± 0.03	0.111 ± 0.007	0.28 ± 0.10	1.3 ± 0.3	0.52 ± 0.14	4.1 ± 0.7	1682.8 ± 123
V106	3/02/93		0.85 ± 0.02	0.133 ± 0.003	0.43 ± 0.12	2.2 ± 0.4	2.63 ± 0.17	2.8 ± 0.7	864.3 ± 62.2
V108	3/09/93		1.81 ± 0.05	0.045 ± 0.006	0.24 ± 0.10	1.5 ± 0.3	0.09 ± 0.09	3.5 ± 0.7	1352.1 ± 97.7
V110	3/16/93		0.44 ± 0.02	0.386 ± 0.002	0.29 ± 0.10	1.2 ± 0.3	3.51 ± 0.14	4.2 ± 0.7	1036 ± 75.3
V111	3/30/93		0.43 ± 0.02	0.122 ± 0.006	0.33 ± 0.09	0.8 ± 0.3	1.92 ± 0.13	6.4 ± 0.9	1033.2 ± 74
V112	4/06/93		0.46 ± 0.01	0.094 ± 0.002	0.47 ± 0.10	2.7 ± 0.4	3.26 ± 0.16	5 ± 0.7	561.4 ± 40.6
V113	4/13/93		0.36 ± 0.05	0.071 ± 0.004	0.77 ± 0.10	3 ± 0.5	1.90 ± 0.16	4 ± 0.7	886.9 ± 64.1
V114	4/20/93		0.33 ± 0.01	0.116 ± 0.005	0.55 ± 0.11	1.2 ± 0.4	3.07 ± 0.16	2.1 ± 0.7	1001 ± 73.4
V115	4/27/93		0.33 ± 0.01	0.007 ± 0.007	0.50 ± 0.10	1.4 ± 0.3	1.80 ± 0.15	3.8 ± 0.6	647.4 ± 47.1
V116	5/04/93		0.62 ± 0.02	0.153 ± 0.004	1.33 ± 0.10	1.8 ± 0.4	1.92 ± 0.14	1.1 ± 0.5	593 ± 43.4
V117	5/11/93		0.65 ± 0.02	0.065 ± 0.002	0.07 ± 0.07	0.4 ± 0.4	1.07 ± 0.18	5.4 ± 0.8	4695.8 ± 340
V119	5/18/93		0.47 ± 0.02	0.107 ± 0.002	0.52 ± 0.11	<	1.10 ± 0.18	3.1 ± 0.7	2478.4 ± 198
V120a	5/25/93		0.50 ± 0.03	0.081 ± 0.002	0.36 ± 0.12	0.25 ± 0.25	0.50 ± 0.19	4.7 ± 0.8	2256.6 ± 169
V121	6/01/93		0.80 ± 0.02	0.159 ± 0.008	0.52 ± 0.10	<	2.13 ± 0.15	3.4 ± 0.7	2415 ± 181
V122	6/08/93		0.19 ± 0.02	0.44 ± 0.12	0.44 ± 0.12	2 ± 0.4	0.52 ± 0.10	3.3 ± 0.7	3289 ± 246
V123	6/15/93		0.48 ± 0.02	0.47 ± 0.11	0.47 ± 0.11	1.6 ± 0.4	0.44 ± 0.12	0.45 ± 0.45	564.1 ± 42.5
V124	6/22/93		0.77 ± 0.02	0.77 ± 0.02	0.49 ± 0.12	0.8 ± 0.3	0.49 ± 0.12	3.9 ± 0.7	2745.4 ± 205
V125	6/29/93		0.81 ± 0.03	0.78 ± 0.12	0.78 ± 0.12	0.9 ± 0.3	0.78 ± 0.12	2.4 ± 0.7	2583.5 ± 194
V126	7/06/93		0.51 ± 0.02	1.25 ± 0.13	1.25 ± 0.13	1.1 ± 0.4	1.25 ± 0.13	3.3 ± 0.7	4505.7 ± 338
V127	7/13/93		0.99 ± 0.03	1.31 ± 0.11	1.31 ± 0.11	1.5 ± 0.4	1.31 ± 0.11	2.1 ± 0.7	2672.6 ± 200
V128	7/20/93		0.49 ± 0.02	0.79 ± 0.11	0.79 ± 0.11	1.1 ± 0.4	0.79 ± 0.11	8.5 ± 1.1	5627.6 ± 422
V129	7/27/93		0.41 ± 0.02	0.41 ± 0.02	0.60 ± 0.11	1.1 ± 0.4	0.60 ± 0.11	2.5 ± 0.7	2851.7 ± 199
V130	8/3/93		0.41 ± 0.02	0.41 ± 0.02	0.60 ± 0.11	0.25 ± 0.25	0.60 ± 0.11	2 ± 0.6	3030.2 ± 228

Year 3

Sample ID	Al	As	Cd	Cr	Cu	Ni	Pb	S
	ng/m ³ Sigma							
V131	7.95	0.35	0.05	0.32	0.11	0.32	0.11	0.8
V132	41.8	0.79	0.03	0.61	0.11	0.8	0.3	0.8
133								1182.5
134								4949.7
135								88.8
136								369
137								
138								
Vims								
Year 1								
First 12 months statistics								
ave	153	0.629	0.15	0.587	2.90	2.527	4.0	2687
STD	124	0.323	0.10	0.545	1.00	1.098	2.6	927
avg error	14	0.042	0.02	0.313	0.37	0.495	0.7	540
count	19	19	18	19	10	10	14	9
min	11	0.040	0.01	0.020	0.84	0.377	0.9	1446
max	540	1.446	0.42	1.904	4.64	3.878	11.0	4564
median	98	0.548	0.12	0.244	0.00	0.000	2.1	2348
max/mean	3.5	2.3	2.8	3.2	1.6	1.5	2.8	1.7
18 mos.								
Year 2								
First 18 months statistics								
ave	129	0.521	0.12	0.540	2.11	2.703	2.9	2130
STD	103	0.341	0.09	0.447	1.28	1.636	2.4	1352
avg error	11	0.030	0.03	0.199	0.32	0.420	0.6	382
count	34	34	31	33	24	22	29	24
min	11	0.000	0.00	0.020	0.17	0.270	0.1	17
max	540	1.446	0.42	1.904	4.64	6.231	11.0	4681
median	92	0.438	0.09	0.286	0.85	1.085	1.3	1798
max/mean	4.2	2.8	3.5	3.5	2.2	2.3	3.8	2.2
Year 3								
First 18 months statistics								
ave	84	0.447	0.11	0.471	1.58	3.143	2.9	1709
STD	55	0.276	0.06	0.200	1.15	1.654	1.9	1060
avg error	5	0.014	0.03	0.051	0.25	0.210	0.5	214
count	34	32	31	33	32	30	33	34
min	11	0.000	0.00	0.178	0.17	0.270	0.1	7
max	270	1.016	0.23	0.925	4.70	6.231	7.1	4681
median	66	0.434	0.11	0.434	1.13	2.414	2.6	1578
max/mean	3.2	2.3	2.0	2.0	3.0	2.0	2.5	2.7
Year 3								
First 18 months statistics								
ave	66	0.600	0.17	0.479	1.70	2.592	4.0	1871
STD	51	0.320	0.14	0.310	0.95	1.297	1.9	1297
avg error	13	0.027	0.01	0.108	0.35	0.183	0.7	143
count	33	60	48	55	54	46	58	60
min	4	0.044	0.01	0.036	0.25	0.083	0.5	124
max	247	1.835	0.81	1.333	5.50	13.343	11.6	5627
median	49	0.526	0.13	0.422	1.50	1.790	3.7	1510
max/mean	3.7	3.1	4.7	2.8	3.2	5.1	2.9	3.0

¹ Sampler malfunction, either no sample was collected or the air volume was inaccurate, values (when reported) are ng/filter, but not included in averages or flux estimates.
² Value not included in average.
 ND = not detected.

Se ng/m³ Sigma ng/m³ Sigma Zn ng/m³ Sigma V ng/m³ Sigma Br ng/m³ Sigma Fe ng/m³ Sigma Mn ng/m³ Sigma

(Haven Beach Site)

Se ng/m³ Sigma Zn ng/m³ Sigma V ng/m³ Sigma Br ng/m³ Sigma Fe ng/m³ Sigma Mn ng/m³ Sigma

2.65 ± 0.23	(39.3 ± 3.7)	3.89 ± 0.33	5.38 ± 0.45	129 ± 4.4	4.08 ± 0.16
2.52 ± 0.21	(104 ± 9.5)	5.21 ± 0.44	5.58 ± 0.46	144 ± 4.0	3.90 ± 0.15
(0.265 ± 0.06)	(166 ± 15)	(1.19 ± 0.10)	(0.43 ± 0.066)	(175 ± 6.7)	(1.18 ± 0.027)
1.61 ± 0.14	(116 ± 11)	4.71 ± 0.40	4.12 ± 0.35	104 ± 3.6	2.39 ± 0.10
0.937 ± 0.09	(51.4 ± 4.8)	2.49 ± 0.21	2.67 ± 0.23	28.2 ± 3.0	0.896 ± 0.038
2.98 ± 0.27	(33.9 ± 3.3)	3.86 ± 0.31	3.98 ± 0.27	67.4 ± 5.0	1.54 ± 0.034
3.90 ± 0.33	(35.5 ± 3.6)	7.39 ± 0.62	7.05 ± 0.57	89.5 ± 4.6	3.14 ± 0.12
0.240 ± 0.02	2.33 ± 0.32	0.419 ± 0.04	0.259 ± 0.026	6.30 ± 0.91	< 0.075 ± 0.075
(2.83 ± 0.26)	(279 ± 26)	(18.2 ± 1.5)	(7.5 ± 0.63)	(155 ± 8.6)	(6.8 ± 0.27)
2.46 ± 0.23	21.5 ± 2.4	6.89 ± 0.56	3.31 ± 0.23	110 ± 7.5	3.26 ± 0.070
1.65 ± 0.14	17.9 ± 1.8	5.71 ± 0.48	3.41 ± 0.28	80.5 ± 3.2	2.32 ± 0.092
(3.09 ± 0.27)	(28.1 ± 2.8)	(8.48 ± 0.68)	(2.94 ± 0.21)	(203 ± 8.6)	(4.28 ± 0.09)
0.959 ± 0.088	12.7 ± 1.5	4.28 ± 0.38	3.89 ± 0.33	225 ± 7.3	3.91 ± 0.15
1.13 ± 0.10	12.3 ± 1.2	3.99 ± 0.32	2.65 ± 0.18	107 ± 3.4	2.33 ± 0.05
0.768 ± 0.072	7.48 ± 0.86	3.35 ± 0.29	4.10 ± 0.35	108 ± 4.4	3.04 ± 0.12
1.30 ± 0.13	6.98 ± 1.3	6.59 ± 0.58	3.17 ± 0.27	81.7 ± 5.7	2.45 ± 0.10
1.36 ± 0.12	11.2 ± 1.1	9.08 ± 0.73	2.31 ± 0.17	73.3 ± 3.0	2.75 ± 0.06
1.09 ± 0.10	9.57 ± 1.0	2.69 ± 0.23	2.88 ± 0.24	130 ± 3.9	3.11 ± 0.12
1.67 ± 0.14	10.8 ± 1.1	3.59 ± 0.30	3.79 ± 0.32	141 ± 4.4	4.02 ± 0.16
0.748 ± 0.068	11.4 ± 1.1	3.06 ± 0.28	3.35 ± 0.31	136 ± 4.7	2.66 ± 0.10
0.710 ± 0.067	4.19 ± 0.62	5.17 ± 0.44	2.26 ± 0.30	300 ± 8.5	6.22 ± 0.24
3.00 ± 0.25	15.1 ± 1.5	5.59 ± 0.47	5.35 ± 0.45	187 ± 4.9	4.00 ± 0.16
(2.03 ± 0.18)	(29.3 ± 2.9)	(6.48 ± 0.55)	(8.30 ± 0.68)	(361 ± 10)	(7.19 ± 0.28)
1.67 ± 0.10	28.02 ± 1.97	4.05 ± 0.24	1.626 ± 0.129	234.7 ± 5.7	3.66 ± 0.15
0.98 ± 0.06	5.54 ± 0.46	5.17 ± 0.28	1.873 ± 0.083	77.2 ± 2.6	1.05 ± 0.09
1.09 ± 0.07	1.50 ± 0.33	3.71 ± 0.22	2.318 ± 0.152	11.1 ± 2.6	1.83 ± 0.08
1.08 ± 0.03	8.45 ± 0.63	4.03 ± 0.22	2.351 ± 0.154	82.8 ± 2.5	1.78 ± 0.04
0.62 ± 0.04	5.05 ± 0.40	2.36 ± 0.14	1.070 ± 0.068	56.5 ± 1.9	1.09 ± 0.04
0.95 ± 0.07	8.04 ± 0.80	5.82 ± 0.32	1.356 ± 0.084	113.5 ± 5.1	1.36 ± 0.04
±	6.253 ± 0.551	4.26 ± 0.23	3.866 ± 0.177	31.7 ± -2.6	1.32 ± 0.04
0.85 ± 0.06	15.06 ± 1.10	2.50 ± 0.13	2.202 ± 0.089	233.1 ± 5.8	2.49 ± 0.05
1.10 ± 0.08	14.13 ± 1.10	4.61 ± 0.27	4.473 ± 0.251	91.7 ± 4.0	3.82 ± 0.15
1.99 ± 0.12	27.97 ± 1.95	4.85 ± 0.26	8.393 ± 0.335	125.5 ± 4.1	6.93 ± 0.15
1.81 ± 0.11	(117) ± (7.92)	5.18 ± 0.31	6.500 ± 0.436	182.8 ± 6.2	< (8.81) ±
2.18 ± 0.13	(48.2) ± (3.32)	8.40 ± 0.45	4.595 ± 0.162	142.2 ± 5.0	2.99 ± 0.07
2.45 ± 0.16	(62.4) ± (4.27)	9.63 ± 0.57	5.484 ± 0.341	111.2 ± 5.2	3.16 ± 0.13
1.07 ± 0.07	(40.28) ± (2.75)	4.09 ± 0.22	2.403 ± 0.350	96.2 ± 3.5	(14.8) ± (0.308)
2.97 ± 0.20	(358) ± (24)	2.89 ± 0.17	6.696 ± 0.404	(644) ± (14.9)	(9.43) ± (0.37)
0.17 ± 0.02	6.25 ± 0.54	0.68 ± 0.04	0.962 ± 0.041	49.5 ± 2.7	0.67 ± 0.02
2.97 ± 0.17	25.44 ± 1.77	5.33 ± 0.32	4.596 ± 0.260	(448) ± (9.73)	3.80 ± 0.16
1.76 ± 0.11	11.19 ± 0.83	4.15 ± 0.22	2.569 ± 0.087	69.9 ± 2.8	1.80 ± 0.04
1.94 ± 0.10	12.66 ± 0.94	3.99 ± 0.24	2.837 ± 0.156	54.3 ± 2.9	1.23 ± 0.05
1.55 ± 0.09	9.36 ± 0.70	4.11 ± 0.22	3.728 ± 0.123	38.3 ± 2.2	1.48 ± 0.04
1.98 ± 0.11	13.35 ± 0.91	3.46 ± 0.06	3.353 ± 0.020	44.4 ± 3.4	1.2 ± 0.2
1.44 ± 0.08	10.58 ± 0.75	6.62 ± 0.13	1.354 ± 0.008	44.0 ± 4.3	1.1 ± 0.2
1.00 ± 0.06	8.84 ± 0.62	4.54 ± 0.08	0.375 ± 0.006	25.0 ± 4.1	0.75 ± 0.2
1.72 ± 0.10	10.64 ± 0.75	3.15 ± 0.07	1.434 ± 0.018	75.7 ± 4.5	2 ± 0.3
1.80 ± 0.10	15.32 ± 1.03	5.25 ± 0.10	2.849 ± 0.011	84.6 ± 2.9	2.4 ± 0.2
2.17 ± 0.12	13.68 ± 0.93	7.09 ± 0.14	2.464 ± 0.018	95.1 ± 3.6	2.6 ± 0.3
1.81 ± 0.10	19.50 ± 1.32	4.68 ± 0.09	3.801 ± 0.034	113.8 ± 4.0	2.4 ± 0.3

Se	Zn	V	Br	Fa	Mn
ng/m ³ Sigma					
0.91 ± 0.05	8.00 ± 0.57	5.39 ± 0.10	1.554 ± 0.027	79.0 ± 3.6	1.3 ± 0.2
1.04 ± 0.06	9.36 ± 0.65	2.81 ± 0.06	1.396 ± 0.030	48.9 ± 3.4	1 ± 0.2
1.39 ± 0.08	22.43 ± 1.52	4.02 ± 0.08	2.895 ± 0.026	135.6 ± 4.5	3.7 ± 0.4
1.15 ± 0.07	8.53 ± 0.60	7.14 ± 0.12	2.086 ± 0.017	55.9 ± 3.5	1.1 ± 0.2
1.10 ± 0.06	7.67 ± 0.55	3.86 ± 0.08	2.779 ± 0.032	ND	2.4 ± 0.3
0.73 ± 0.04	5.39 ± 0.41	2.89 ± 0.05	1.484 ± 0.024	38.2 ± 3.5	0.9 ± 0.2
1.08 ± 0.06	9.50 ± 0.66	5.50 ± 0.11	0.601 ± 0.018	52.5 ± 3.4	0.8 ± 0.2
1.11 ± 0.06	11.78 ± 0.81	6.26 ± 0.12	2.511 ± 0.017	115.5 ± 4.0	2.20 ± 0.30
0.82 ± 0.05	6.43 ± 0.48	2.48 ± 0.05	2.178 ± 0.019	66.9 ± 3.8	1.70 ± 0.20
1.39 ± 0.08	8.16 ± 0.58	3.67 ± 0.07	2.165 ± 0.027	56.3 ± 3.5	1.30 ± 0.20
1.41 ± 0.08	8.35 ± 0.59	2.82 ± 0.05	1.872 ± 0.020	78.3 ± 3.8	1.10 ± 0.20
2.76 ± 0.16	12.03 ± 0.85	7.10 ± 0.27	2.384 ± 0.025	(1050.4 ± 21.38)	(13.1 ± 1.1)
1.90 ± 0.15	13.8 ± 1.6	7.2 ± 1.2	3.93 ± 0.10	(1615.28 ± 37.41)	(24.1 ± 2)
1.28 ± 0.09	8.8 ± 0.6	4.8 ± 0.7	3.08 ± 0.08	94 ± 5	2.1 ± 0.3
1.45 ± 0.09	11.2 ± 0.8	5.3 ± 0.7	3.00 ± 0.08	84 ± 4	2.4 ± 0.3
1.84 ± 0.17	10.0 ± 1.6		2.81 ± 0.31	138 ± 18	
2.79 ± 0.20	12.8 ± 1.3		3.93 ± 0.16	184 ± 10	3.4 ± 0.8
0.40 ± 0.05	2.2 ± 0.3	2.7 ± 0.5	0.48 ± 0.08	20 ± 4	
1.03 ± 0.08	5.0 ± 0.5	2.8 ± 0.7	2.99 ± 0.08	64 ± 5	1.3 ± 0.3
1.58 ± 0.11	14.2 ± 1.1	1.7 ± 0.5	2.13 ± 0.08	101 ± 6	3.7 ± 0.5
0.40 ± 0.04	2.2 ± 0.3	3.3 ± 0.5	2.53 ± 0.08	14 ± 3	
0.73 ± 0.06	6.2 ± 0.5	2.7 ± 0.6	1.65 ± 0.12	49 ± 5	1.2 ± 0.3
0.66 ± 0.06	6.7 ± 0.6	2.2 ± 0.5	1.47 ± 0.07	55 ± 4	1.5 ± 0.3
0.45 ± 0.04	6.4 ± 0.5	2.9 ± 0.6	2.82 ± 0.10	39 ± 4	1.4 ± 0.3
2.03 ± 0.12	12.0 ± 0.9	9.7 ± 1	5.06 ± 0.07	87 ± 3	2.2 ± 0.2
1.10 ± 0.07	6.4 ± 0.5	3.9 ± 0.5	2.79 ± 0.05	44 ± 2	2 ± 0.2
1.07 ± 0.09	10.7 ± 0.9	4 ± 0.6	2.92 ± 0.09	61 ± 5	1.9 ± 0.3
1.16 ± 0.08	12.3 ± 0.9	5.8 ± 0.8	2.65 ± 0.08	42 ± 4	1.7 ± 0.3
1.10 ± 0.08	13.5 ± 1.0	10.6 ± 1.2	4.96 ± 0.11	114 ± 5	2 ± 0.3
0.70 ± 0.05	6.2 ± 0.5	4 ± 0.6	5.00 ± 0.12	24 ± 4	
2.65 ± 0.17	12.2 ± 0.9	2.8 ± 0.6	6.16 ± 0.09	48 ± 5	1.3 ± 0.3
2.36 ± 0.15	26.5 ± 1.9	3.1 ± 0.6	4.76 ± 0.12	93 ± 6	2.8 ± 0.4
0.66 ± 0.05	11.4 ± 0.8	4.8 ± 0.7	5.01 ± 0.13	27 ± 4	1.2 ± 0.3
3.12 ± 0.19	10.4 ± 0.8	13.1 ± 1.4	5.85 ± 0.11	52 ± 5	
1.50 ± 0.11	8.4 ± 0.7	9.5 ± 1.1	2.72 ± 0.08	29 ± 5	1.7 ± 0.3
0.99 ± 0.07	13.5 ± 1.0	2.4 ± 0.5	1.55 ± 0.08	58 ± 5	1.5 ± 0.3
	4.0 ± 0.4	1.9 ± 0.3	0.20 ± 0.04	10 ± 3	
1.85 ± 0.11	13.5 ± 1.0	4.2 ± 0.6	4.45 ± 0.09	58 ± 3	1.9 ± 0.3
0.68 ± 0.05	9.1 ± 0.7	3.2 ± 0.5	3.82 ± 0.10	34 ± 3	1.1 ± 0.2
1.24 ± 0.08	8.8 ± 0.7	6 ± 0.7	2.88 ± 0.06	79 ± 4	2.3 ± 0.3
0.94 ± 0.07	9.5 ± 0.7	5.4 ± 0.7	7.03 ± 0.15	25 ± 3	1 ± 0.2
1.22 ± 0.07	7.1 ± 0.5	2.5 ± 0.5	3.11 ± 0.07	48 ± 4	1.3 ± 0.3
1.46 ± 0.10	7.3 ± 0.6	0.4 ± 0.4	2.19 ± 0.07	37 ± 4	1.9 ± 0.3
1.66 ± 0.10	12.8 ± 0.9	7.1 ± 0.9	4.31 ± 0.07	16 ± 3	0.5 ± 0.2
0.81 ± 0.06	4.2 ± 0.4	2.4 ± 0.5	1.51 ± 0.06	31 ± 3	3.6 ± 0.4
0.55 ± 0.04	4.2 ± 0.4	1.3 ± 1.3	2.17 ± 0.06	32 ± 3	1.9 ± 0.3
1.49 ± 0.09	9.0 ± 0.7	3.8 ± 0.6	2.71 ± 0.06	57 ± 4	1.5 ± 0.3
0.70 ± 0.05	13.2 ± 1.0	6.8 ± 0.8	3.28 ± 0.06	36 ± 3	0.9 ± 0.2
1.06 ± 0.07	6.9 ± 0.6	7.5 ± 0.9	2.14 ± 0.06	100 ± 4	2.1 ± 0.4
0.60 ± 0.05	5.6 ± 0.5	3.2 ± 0.7	3.45 ± 0.13	61 ± 4	1.6 ± 0.3
0.37 ± 0.04	10.2 ± 0.8	2.8 ± 0.5	1.58 ± 0.08	30 ± 3	0.6 ± 0.2
	1.2 ± 0.3	7.2 ± 0.9	2.58 ± 0.06	161 ± 5	4.2 ± 0.4
2.12 ± 0.13	10.3 ± 0.8	5.9 ± 0.8	1.23 ± 0.07	28 ± 4	1.4 ± 0.3
0.80 ± 0.06	3.8 ± 0.4	2.9 ± 0.6	3.66 ± 0.06	100 ± 4	1.8 ± 0.3
1.27 ± 0.08	7.5 ± 0.6	1.9 ± 0.5	3.08 ± 0.08	47 ± 4	0.7 ± 0.3
1.27 ± 0.08	8.8 ± 0.6	0.84 ± 0.03	4.25 ± 0.08	84 ± 4	3.2 ± 0.4
1.34 ± 0.08	9.0 ± 0.7	6.1 ± 0.8	1.78 ± 0.08	32 ± 4	1.2 ± 0.3
	2.6 ± 0.3	3.7 ± 0.6	2.18 ± 0.06	89 ± 4	2.2 ± 0.3
1.27 ± 0.09	9.3 ± 0.7	1.3 ± 0.5	3.26 ± 0.08	39 ± 4	1.1 ± 0.3
1.17 ± 0.07	4.8 ± 0.4	4.1 ± 0.7	3.26 ± 0.08	191 ± 7	4 ± 0.5
	8.0 ± 0.8	4 ± 0.8	4.09 ± 0.07	106 ± 5	1.5 ± 0.3
1.02 ± 0.07	3.1 ± 0.4	3.3 ± 0.7	2.82 ± 0.08	144 ± 5	4.2 ± 0.5
2.61 ± 0.15	13.2 ± 1.0	4.9 ± 0.7	2.93 ± 0.06	76 ± 4	2 ± 0.3
0.87 ± 0.06	5.8 ± 0.5	3.8 ± 0.6		97 ± 4	2.1 ± 0.3
1.68 ± 0.10	8.2 ± 0.7	8.3 ± 1			

Se	Zn	V	Br	Fa	Mn
ng/m ³ Sigma					
0.51 0.04	2.2 0.3	3.4 0.6	2.37 0.08	24 4	0.5 0.2
2.45 0.15	14.8 1.1	5.1 0.7	4.73 0.07	75 4	2.1 0.3

First 12 months statistics

1.7	11.0	4.6	3.659	118	2.852
1.0	5.1	1.9	1.451	65	1.302
0.1	1.2	0.4	0.305	5	0.111
19	13	19	19	19	19
0.2	2.3	0.4	0.259	6	0.075
3.9	21.5	9.1	7.045	300	6.224
1.2	5.6	3.9	3.330	105	2.702
2.3	1.9	2.0	1.9	2.5	2.1

1.6	11.5	4.7	3.668	118	2.824
0.9	7.0	1.9	1.806	65	1.437
0.1	1.1	0.3	0.265	4	0.101
33	23	34	34	33	31
0.2	1.5	0.4	0.269	6	0.075
3.9	28.0	9.6	8.393	300	6.932
1.1	5.3	4.1	3.239	100	2.422
2.5	2.4	2.1	2.3	2.6	2.5

1.4	12.0	4.6	2.892	87	2.062
0.7	6.7	1.7	1.814	54	1.313
0.1	0.9	0.2	0.123	4	0.153
33	29	34	34	31	31
0.2	1.5	0.7	0.375	11	0.669
3.0	28.0	9.6	8.393	235	6.932
1.3	8.8	4.1	2.403	73	1.355
2.1	2.3	2.1	2.9	2.7	3.4

1.3	8.7	4.6	3.050	66	1.867
0.7	4.2	2.7	1.320	40	0.903
0.1	0.7	0.6	0.082	4	0.304
56	60	58	60	57	52
0.4	1.2	0.4	0.203	10	0.500
3.1	26.5	13.1	7.027	191	4.200
1.2	8.4	3.9	2.616	56	1.700
2.4	3.0	2.8	2.3	2.9	2.2

Appendix A3

Table A3.1. Mean field measured vapor concentrations, pg/m³, and mean analytical detection limits for PAHs, Chesapeake Bay, 1992.

Polycyclic Aromatic Hydrocarbon	Mean Vapor Concentration WYE	Mean Analytical Detection Limits WYE	Mean Vapor Concentration ELMS	Mean Analytical Detection Limits ELMS	Mean Vapor Concentration HAVEN BEACH	Mean Analytical Detection Limits HAVEN BEACH
Fluorene	863.7	0.2	810.4	0.13	832.1	0.37
Phenanthrene	3046.7	0.2	1937.3	0.13	2460.2	0.37
Anthracene	31.8	0.2	47.2	0.13	12.2	0.37
Fluoranthene	582.6	0.3	231.4	0.26	530.7	0.73
Pyrene	643.3	0.3	397.8	0.26	1055.4	0.73
Benzo[a]anthracene	3.6	0.3	2.6	0.26	3.3	0.73
Chrysene	20.2	0.3	53.8	0.26	62.0	0.73
Benzo[b]fluoranthene	2.5	0.3	34.3	0.26	3.8	0.73
Benzo[k]fluoranthene	1.9	0.3	27.9	0.26	3.3	0.73
Benzo[e]pyrene	1.9	0.3	1.2	0.26	5.2	0.73
Benzo[a]pyrene	0.1	0.3	3.7	0.26	2.1	0.73
Indeno[123-cd]pyrene	0.5	1.5	0.8	1.28	2.2	3.56
Dibenz[ah]anthracene	0.5	1.5	0.2	1.28	0.3	3.56
Benzo[ghi]perylene	1.8	0.6	0.9	0.53	9.0	1.46

Table A3.2 Summary of Laboratory Surrogate Compounds for Chesapeake Bay Atmospheric Deposition Samples, Elms Site, 1992

Deuterated PAH	Sample Matrix	Mean Recovery	Relative* STD dev (%)	N**	70% < N N < 130%
d-10 Fluoranthene	Rain Dissolved	60.6	5.45	5	0
d-12 Chrysene		84.7	28.83	18	15
d-10 Anthracene	Rain Filter	53.6	32.29	17	1
d-10 Fluoranthene		51.6	39.9	18	4
d-12 Chrysene		80	25.47	23	9
Non-Commercial (PCB)	Sample Matrix	Mean Recovery	Relative* STD dev (%)	N**	70% < N N < 130%
3,5-dichlorobiphenyl	Rain Dissolved	73.3	46.53	22	10
3,5-dichlorobiphenyl	Rain Filter	63.8	54.33	22	12

Relative STD DEV = STD DEV 100/[Mean].

**Number of samples.

Table A3.3 Summary of Laboratory Surrogate Compounds for Chesapeake Bay Atmospheric Deposition Samples, Haven Beach, 1992

Deuterated PAH	Matrix	Mean Recovery	Relative* STD dev (%)	N**	70% < N N < 130%
d-10 Anthracene	Rain Dissolved	51.7	32.45	22	4
d-10 Fluoranthene		76	58.71	21	11
d-12 Chrysene		90.97	42.59	28	16
d-10 Anthracene	Rain Filter	32.74	76.51	25	2
d-10 Fluoranthene		53.84	93.01	25	3
d-12 Chrysene		42.57	87.62	25	2
Non-Commercial (PCB)	Matrix	Mean Recovery	Relative* STD dev (%)	N**	70% < N N < 130%
3,5-dichlorobiphenyl	Rain Dissolved	59.7	43.22	28	12
3,5-dichlorobiphenyl	Rain Filter	52.22	45.22	28	6

Table A3.4. Mean Mass (ng) Summary of Polycyclic Hydrocarbons and Polychlorinated Biphenyls in Laboratory and Field Blank Matrices, Chesapeake Bay Atmospheric Deposition Project, Elms, 1992.

Polycyclic Aromatic Hydrocarbon	LABORATORY				FIELD			
	Mean Resin (N=6)	%Rel STD^	Mean Filter (N=3)	%Rel STD^	Mean Resin (N=1)	%Rel STD^	Mean Filter (N=1)	%Rel STD^
Fluorene	5.02	138	ND	--	25.52	--	ND	--
Phenanthrene	25.74	111	ND	--	82.67	--	21.11	--
Anthracene	11.81	186	ND	--	1.95	--	3.82	--
Fluoranthene	63.17	139	ND	--	31.59	--	31.34	--
Pyrene	9.48	220	ND	--	14.83	--	19.09	--
Benzo[a]anthracene	0.05	224	ND	--	0.06	--	10.57	--
Chrysene	0.10	224	ND	--	7.30	--	16.20	--
Benzo[b]fluoranthene	85.42	203	ND	--	5.21	--	50.46	--
Benzo[k]fluoranthene	5.74	220	ND	--	1.77	--	22.84	--
Benzo[e]pyrene	ND	0	ND	--	3.84	--	18.66	--
Benzo[a]pyrene	47.25	204	ND	--	1.60	--	12.89	--
Indeno[123-cd]pyrene	0.18	224	ND	--	4.16	--	16.24	--
Dibenz[ah]anthracene	ND	0	ND	--	0.00	--	0.00	--
Benzo[ghi]perylene	0.36	224	ND	--	4.55	--	14.57	--
Total PCB*	4.66	30.56	8.44	37.26	20.87	--	29.26	--

Relative STD DEV = STD dev* 100/mean.

*N = 5 for lab resins.

Table A3.5. Mean Mass (ng) Summary of Polycyclic Hydrocarbons and Polychlorinated Biphenyls in Laboratory and Field Blank Matrices, Chesapeake Bay Atmospheric Deposition Project, Haven Beach, 1992.

Polycyclic Aromatic Hydrocarbon	LABORATORY				FIELD			
	Mean Resin (N=1)	Rel STD^	Mean Filter (N=1)	Rel STD^	Mean Resin (N=1)	Rel STD^	Mean Filter (N=1)	Rel STD^
Fluorene	ND	--	ND	--	7.09	--	ND	--
Phenanthrene	ND	--	ND	--	4.17	--	ND	--
Anthracene	ND	--	ND	--	0.39	--	ND	--
Fluoranthene	ND	--	ND	--	6.85	--	ND	--
Pyrene	ND	--	ND	--	1.92	--	ND	--
Benzo[a]anthracene	ND	--	ND	--	1.52	--	ND	--
Chrysene	ND	--	ND	--	1.26	--	ND	--
Benzo[b]fluoranthene	ND	--	ND	--	ND	--	ND	--
Benzo[k]fluoranthene	ND	--	ND	--	ND	--	ND	--
Benzo[e]pyrene	ND	--	ND	--	ND	--	ND	--
Benzo[a]pyrene	ND	--	ND	--	ND	--	ND	--
Indeno[123-cd]pyrene	ND	--	ND	--	ND	--	ND	--
Dibenz[ah]anthracene	ND	--	ND	--	ND	--	ND	--
Benzo[ghi]perylene	ND	--	ND	--	ND	--	ND	--
Total PCB	6.47	--	6.81	--	13.07	--	10.2	--

Relative STD DEV = STD dev* 100/mean.

